

Annexes

The following twenty-six annexes provide additional information to the material presented in the main body of this report. Annexes A through O discuss methodologies for individual source categories in greater detail than was presented in the main body of the report and include explicit activity data and emission factor tables. Annex P analyzes the key sources in this report. Annex Q presents a summary of Global Warming Potential values. Annexes R and S summarize U.S. emissions of ozone depleting substances (e.g., CFCs and HCFCs) and sulfur dioxide (SO₂), respectively. Annex T provides a complete list of emission sources assessed in this report. Annex U presents the IPCC reference approach for estimating CO₂ emissions from fossil fuel combustion. Annex V addresses the criteria for the inclusion of an emission source category and some of the sources that meet the criteria but are nonetheless excluded from U.S. estimates. Annex W provides some useful constants, unit definitions, and conversions. Annexes X and Y provide a listing of abbreviations and chemical symbols used. Finally, Annex Z contains a glossary of terms related to greenhouse gas emissions and inventories.

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ANNEX A: Methodology for Estimating Emissions of CO₂ from Fossil Fuel Combustion

Carbon dioxide (CO₂) emissions from fossil fuel combustion were estimated using a “bottom-up” methodology characterized by six steps. These steps are described below.

Step 1: Determine Energy Consumption by Fuel Type and Sector

The bottom-up methodology used by the United States for estimating CO₂ emissions from fossil fuel combustion is conceptually similar to the approach recommended by the Intergovernmental Panel on Climate Change (IPCC) for countries that intend to develop detailed, sectoral-based emission estimates (IPCC/UNEP/OECD/IEA 1997). Basic consumption data are presented in Columns 2-8 of Table A-1 through Table A-11, with totals by fuel type in Column 8 and totals by end-use sector in the last rows. Fuel consumption data for the bottom-up approach were obtained directly from the Energy Information Administration (EIA) of the U.S. Department of Energy. These data were first gathered in physical units, and then converted to their energy equivalents (see “Converting Physical Units to Energy Units” in Annex W). The EIA data were collected through a variety of consumption surveys at the point of delivery or use and qualified with survey data on fuel production, imports, exports, and stock changes. Individual data elements were supplied by a variety of sources within EIA. Most information was taken from published reports, although some data were drawn from unpublished energy studies and databases maintained by EIA.

Energy consumption data were aggregated by sector (i.e., residential, commercial, industrial, transportation, electricity generation, and U.S. territories), primary fuel type (e.g., coal, natural gas, and petroleum), and secondary fuel type (e.g., motor gasoline, distillate fuel, etc.). The 2000 total energy consumption across all sectors, including territories, and energy types was 83,610 trillion British thermal units (Tbtu), as indicated in the last entry of Column 8 in Table A-1. This total includes fuel used for non-energy purposes and fuel consumed as international bunkers, both of which are deducted in later steps.

Fuel consumption data for electricity generation data by nonutility power producers are initially categorized by the EIA as part of the industrial sector. These data are then combined with fuel consumption by electric utilities to form the electricity generation sector. The method for this reallocation is described in detail in EIA’s report on U.S. GHG emissions, *Emissions of Greenhouse Gases in the United States, 2000* (EIA 2001c).

There were a number of modifications made in this report that may cause consumption information herein to differ from figures given in the cited literature. These are 1) the reallocation of some coking coal, petroleum coke, and natural gas consumption for ammonia production to the Industrial Processes chapter, 2) corrections for synthetic natural gas production, 3) corrections for ethanol added to motor gasoline, and 4) corrections for biogas in natural gas.

First, portions of the fuel consumption data for three fuel categories—coking coal, petroleum coke, and natural gas—were reallocated to the Industrial Processes chapter, as these portions were actually consumed as raw material during non-energy related industrial processes. Coking coal, also called “coal coke,” is used as a raw material (specifically as a reducing agent) in the blast furnace process to produce iron and steel, and therefore is not used as a fuel for this process. Similarly, petroleum coke is used in multiple processes as a raw material, and is thus not used as a fuel in those applications. The processes in which petroleum coke is used include 1) ferroalloy production, 2) aluminum production (for the production of carbon anodes and cathodes), and 3) titanium dioxide production (in the chloride process). Finally, natural gas consumption is used for the production of ammonia. Consumption of these fuels for non-energy purposes is presented in the Industrial Processes chapter, and is removed from the energy and non-energy consumption estimates within the Energy chapter.

Second, a portion of industrial coal accounted for in EIA combustion figures is actually used to make “synthetic natural gas” via coal gasification. The energy in this gas enters the natural gas stream, and is accounted for in natural gas consumption statistics. Because this energy is already accounted for as natural gas, it is deducted from industrial

coal consumption to avoid double counting. This makes the figure for other industrial coal consumption in this report slightly lower than most EIA sources.

Third, ethanol has been added to the motor gasoline stream for several years, but prior to 1993 this addition was not captured in EIA motor gasoline statistics. Starting in 1993, ethanol was included in gasoline statistics. However, because ethanol is a biofuel, which is assumed to result in no net CO₂ emissions, the amount of ethanol added is subtracted from total gasoline consumption. Thus, motor gasoline consumption statistics given in this report may be slightly lower than in EIA sources.

Fourth, EIA natural gas consumption statistics include “biomass gas,” which is upgraded landfill methane that is sold to pipelines. However, because this gas is biogenic, the biomass gas total is deducted from natural gas consumption. The subtraction is done only from natural gas in the industrial sector, as opposed to all end-sectors, because the biogas amount is small. Due to this adjustment—and the ammonia adjustment mentioned previously—industrial natural gas consumption in this report is slightly lower than in EIA sources.

There are also three basic differences between the consumption figures presented in Table A-1 through Table A-11 and those recommended in the IPCC emission inventory methodology.

First, consumption data in the U.S. inventory are presented using higher heating values (HHV)¹ rather than the lower heating values (LHV)² reflected in the IPCC emission inventory methodology. This convention is followed because data obtained from EIA are based on HHV. Of note, however, is that EIA renewable energy statistics are often published using LHV. The difference between the two conventions relates to the treatment of the heat energy that is consumed in the process of evaporating the water contained in the fuel. The simplified convention used by the International Energy Agency for converting from HHV to LHV is to multiply the energy content by 0.95 for petroleum and coal and by 0.9 for natural gas.

Second, while EIA's energy use data for the United States includes only the 50 U.S. states and the District of Columbia, the data reported to the Framework Convention on Climate Change are to include energy consumption within territories. Therefore, consumption estimates for U.S. territories were added to domestic consumption of fossil fuels. Energy consumption data from U.S. territories are presented in Column 7 of Table A-1 through Table A-11. It is reported separately from domestic sectoral consumption, because it is collected separately by EIA with no sectoral disaggregation.

Third, the domestic sectoral consumption data in Table A-1 through Table A-11 include bunker fuels used for international transport activities and non-energy uses of fossil fuels. The IPCC requires countries to estimate emissions from international bunker fuels separately and exclude these emissions from national totals, so international bunker fuel emissions have been estimated in Table A-12 and deducted from national estimates (see Step 4). Similarly, fossil fuels used to produce non-energy products that store carbon rather than release it to the atmosphere are provided in Table A-13 and deducted from national emission estimates (see Step 3). The final fate of these fossil fuel based products is dealt with under the waste combustion source category in cases where the products are combusted through waste management practices.

Step 2: Determine the Carbon Content of All Fuels

The carbon content of combusted fossil fuels was estimated by multiplying energy consumption (Columns 2 through 8 of Table A-1 through Table A-11) by fuel-specific carbon content coefficients (see Table A-14 and Table A-15) that reflect the amount of carbon per unit of energy in each fuel. The resulting carbon contents are sometimes referred to as potential emissions, or the maximum amount of carbon that could potentially be released to the

¹ Also referred to as Gross Calorific Values (GCV).

² Also referred to as Net Calorific Values (NCV).

atmosphere if all carbon in the fuels were oxidized. The carbon content coefficients used in the U.S. inventory were derived by EIA from detailed fuel information and are similar to the carbon content coefficients contained in the IPCC's default methodology (IPCC/UNEP/OECD/IEA 1997), with modifications reflecting fuel qualities specific to the United States.

Step 3: Adjust for the amount of Carbon Stored in Products

Depending on the end-use, non-energy uses of fossil fuels can result in long term storage of some or all of the carbon contained in the fuel. For example, asphalt made from petroleum can sequester up to 100 percent of the carbon contained in the petroleum feedstock for extended periods of time. Other non-energy fossil fuel products, such as lubricants or plastics also store carbon, but can lose or emit some of this carbon when they are used and/or burned as waste.³

The amount of carbon in non-energy fossil fuel products was based upon data that addressed the fraction of carbon that remains in products after they are manufactured, with all non-energy use attributed to the industrial, transportation, and territories end-use sectors. This non-energy consumption is presented in Table A-13. This data were then multiplied by fuel-specific carbon content coefficients (Table A-14 and Table A-15) to obtain the carbon content of the fuel, or the maximum amount of carbon that could remain in non-energy products (Columns 5 and 6 of Table A-13). This carbon content was then multiplied by the fraction of carbon assumed to actually have remained in products (Column 7 of Table A-13), resulting in the final estimates by sector and fuel type, which are presented in Columns 8 through 10 of Table A-13. A detailed discussion of carbon stored in products is provided in the Energy chapter and in Annex B.

Step 4: Subtract Carbon in International Bunker Fuels

Emissions from international transport activities, or international bunker fuel consumption, are not included in national totals, as required by the IPCC (IPCC/UNEP/OECD/IEA 1997). There is currently disagreement internationally as to how these emissions should be allocated, and until this issue is resolved, countries are asked to report them separately. EIA energy statistics, however, include these bunker fuels—jet fuel for aircraft, and distillate fuel oil and residual fuel oil for marine shipping—as part of fuel consumption by the transportation end-use sector. To compensate for this inclusion, international bunker fuel emissions⁴ were calculated separately (see Table A-12) and the carbon content of these fuels was subtracted from the transportation end-use sector. International bunker fuel emissions from military activities were developed using data provided by the Department of Defense as described in the International Bunker Fuels section of the Energy chapter and in Annex I. The calculations of international bunker fuel emissions followed the same procedures used for other fuel emissions (i.e., estimation of consumption, determination of carbon content, and adjustment for the fraction of carbon not oxidized).

Step 5: Account for Carbon that Does Not Oxidize During Combustion

Because combustion processes are not 100 percent efficient, some of the carbon contained in fuels is not emitted in a gaseous form to the atmosphere. Rather, it remains behind as soot, particulate matter and ash. The estimated fraction of carbon not oxidized in U.S. energy conversion processes due to inefficiencies during combustion ranges from 0.5 percent for natural gas to 1 percent for petroleum and coal. Except for coal these assumptions are consistent with the default values recommended by the IPCC (IPCC/UNEP/OECD/IEA 1997). In the United States, unoxidized carbon from coal combustion was estimated to be no more than one percent (Bechtel 1993). Table A-14

³ See Waste Combustion section of the Energy chapter for a discussion of emissions from the combustion of plastics in the municipal solid waste stream.

⁴ Refer to the International Bunker Fuels section of the Energy chapter for a description of the methodology for distinguishing between bunker and non-bunker fuel consumption.

presents fractions oxidized by fuel type, which are multiplied by the net carbon content of the combusted energy to give final emissions estimates.

Of the fraction of carbon that is oxidized (e.g., 99 to 99.5 percent), the vast majority is emitted in its fully oxidized form as carbon dioxide (CO₂). A much smaller portion of this “oxidized” carbon is also emitted as carbon monoxide (CO), methane (CH₄), and non-methane volatile organic compounds (NMVOCs). These partially oxidized or unoxidized carbon compounds when in the atmosphere, though, are generally oxidized to CO₂ through atmospheric processes (e.g., reaction with hydroxyl (OH)).⁵

Step 6: Summarize Emission Estimates

Actual CO₂ emissions in the United States were summarized by major fuel (i.e., coal, petroleum, natural gas, geothermal) and consuming sector (i.e., residential, commercial, industrial, transportation, electricity generation, and U.S. territories). Adjustments for international bunker fuels and carbon in non-energy products were made. Emission estimates are expressed in teragrams of carbon dioxide equivalents (Tg CO₂ Eq.).

To determine total emissions by final end-use sector, emissions from electricity generation were distributed to each end-use sector according to its share of aggregate electricity consumption (see Table A-16). This pro-rated approach to allocating emissions from electricity generation may overestimate or underestimate emissions for particular sectors due to differences in the average carbon content of fuel mixes burned to generate electricity.

⁵ See indirect CO₂ from CH₄ oxidation section in Energy chapter for a discussion of proper accounting of carbon from hydrocarbon and CO emissions.

Table A-12: 2000 CO₂ Emissions From International Bunker Fuel Consumption

Fuel Type	Bunker Fuel Consumption (TBtu)	Carbon Content Coefficient (Tg Carbon/QBtu)¹	Potential Emissions (Tg Carbon)	Fraction Oxidized	Emissions (Tg CO₂ Eq.)
Distillate Fuel Oil	115	19.95	2.3	0.99	8.3
Jet Fuel	816	19.33	15.8	0.99	57.3
Residual Fuel Oil	444	21.49	9.5	0.99	34.6
Total	1,375		27.6		100.2

Note: See Annex I for additional information on military bunkers.

Table A-13: 2000 Carbon In Non-Energy Products

Fuel Type	Non-energy Use (TBtu)	Carbon Content Coefficient (Tg Carbon/QBtu)	Potential Emissions (Tg Carbon)	Fraction Sequestered^a	Carbon Stored (Tg CO₂ Eq.)
Industry	5,512.4		103.6		265.6
Industrial Coking Coal	26.4	25.56	0.7	0.75	1.9
Natural Gas	342.4	14.47	5.0	0.63	11.5
Asphalt & Road Oil	1,275.7	20.62	26.3	1.00	96.4
LPG	1,707.3	16.87	28.8	0.63	66.8
Lubricants	189.9	20.24	3.8	0.09	1.3
Pentanes Plus	286.8	18.24	5.2	0.63	12.1
Petrochemical Feedstocks					
Naphtha (<401 deg. F)	564.2	18.14	10.2	0.63	23.7
Other Oil (>401 deg. F)	664.1	19.95	13.2	0.63	30.7
Still Gas	7.4	17.51	0.1	0.80	0.4
Petroleum Coke	141.4	27.85	3.9	0.50	7.2
Special Naphtha	97.4	19.86	1.9	0.00	0.0
Other (Wax/Misc.)					
Distillate Fuel Oil	7.0	19.95	0.1	0.50	0.3
Residual Fuel	50.3	21.49	1.1	0.50	2.0
Waxes	33.1	19.81	0.7	1.00	2.4
Miscellaneous	119.2	20.23	2.4	1.00	8.8
Transportation	179.4		3.6		1.2
Lubricants	179.4	20.24	3.6	0.09	1.2
U.S. Territories	223.8		4.5		16.5
Lubricants	1.4	20.24	0.0	0.09	0.0
Other Petroleum (Misc.)	222.5	variable	4.5	1.00	16.5
Total	5,915.6		111.7		283.4

^aSee Annex B for additional information.

¹ One QBtu is one quadrillion Btu, or 10¹⁵ Btu. This unit is commonly referred to as a "Quad."

Table A-14: Key Assumptions for Estimating Carbon Dioxide Emissions

Fuel Type	Carbon Content Coefficient (Tg Carbon/QBtu)	Fraction Oxidized
Coal		
Residential Coal	[a]	0.99
Commercial Coal	[a]	0.99
Industrial Coking Coal	[a]	0.99
Industrial Other Coal	[a]	0.99
Coke Imports	27.85	0.99
Transportation Coal	NC	NC
Utility Coal	[a]	0.99
U.S. Territory Coal (bit)	25.14	0.99
Natural Gas	14.47	0.995
Petroleum		
Asphalt & Road Oil	20.62	0.99
Aviation Gasoline	18.87	0.99
Distillate Fuel Oil	19.95	0.99
Jet Fuel	[a]	0.99
Kerosene	19.72	0.99
LPG	[a]	0.99
LPG (energy use/Territories)	[a]	0.99
LPG (non-energy use)	[a]	-
Lubricants	20.24	0.99
Motor Gasoline	[a]	0.99
Residual Fuel	21.49	0.99
Other Petroleum		
AvGas Blend Components	18.87	0.99
Crude Oil	[a]	0.99
MoGas Blend Components	[a]	0.99
Misc. Products	[a]	0.99
Misc. Products (Territories)	variable	0.99
Naphtha (<401 deg. F)	18.14	0.99
Other Oil (>401 deg. F)	19.95	0.99
Pentanes Plus	18.24	0.99
Petrochemical Feedstocks	19.37	0.99
Petroleum Coke	27.85	0.99
Still Gas	17.51	0.99
Special Naphtha	19.86	0.99
Unfinished Oils	[a]	0.99
Waxes	19.81	0.99
Other Wax & Misc.	19.81	0.99
Geothermal	2.05	1.00

Sources: Carbon coefficients from EIA. Combustion efficiency for coal from Bechtel (1993) and for petroleum and natural gas from IPCC (IPCC/UNEP/OECD/IEA 1997).

- Not applicable

NC (Not Calculated)

[a] These coefficients vary annually due to fluctuations in fuel quality (see Table A-15).

Table A-15: Annually Variable Carbon Content Coefficients by Year (Tg Carbon/QBtu)

Fuel Type	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000
Residential Coal	25.92	26.00	26.13	25.97	25.95	26.00	25.92	26.00	26.00	26.00	26.00
Commercial Coal	25.92	26.00	26.13	25.97	25.95	26.00	25.92	26.00	26.00	26.00	26.00
Industrial Coking Coal	25.51	25.51	25.51	25.51	25.52	25.53	25.55	25.56	25.56	25.56	25.56
Industrial Other Coal	25.58	25.59	25.62	25.61	25.63	25.63	25.61	25.63	25.63	25.63	25.63
Utility Coal	25.68	25.69	25.69	25.71	25.72	25.74	25.74	25.76	25.76	25.76	25.76
LPG	16.99	16.98	16.99	16.97	17.01	17.00	16.99	16.99	16.99	16.99	16.99
LPG (energy use/Territories)	17.21	17.21	17.21	17.22	17.22	17.20	17.20	17.18	17.18	17.18	17.18
LPG (non-energy use)	16.83	16.84	16.84	16.80	16.88	16.87	16.86	16.88	16.87	16.88	16.87
Motor Gasoline	19.41	19.41	19.42	19.43	19.45	19.38	19.36	19.35	19.33	19.33	19.34
Jet Fuel	19.40	19.40	19.39	19.37	19.35	19.34	19.33	19.33	19.33	19.33	19.33
MoGas Blend Components	19.41	19.41	19.42	19.43	19.45	19.38	19.36	19.35	19.33	19.33	19.34
Misc. Products	20.16	20.18	20.22	20.22	20.21	20.23	20.25	20.24	20.24	20.19	20.23
Unfinished Oils	20.16	20.18	20.22	20.22	20.21	20.23	20.25	20.24	20.24	20.19	20.23
Crude Oil	20.16	20.18	20.22	20.22	20.21	20.23	20.25	20.24	20.24	20.19	20.23

Source: EIA (2001a) and EIA (2001c)

Table A-16: Electricity Consumption by End-Use Sector (Billion Kilowatt-Hours)

End-Use Sector	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000
Residential	924	955	936	995	1,008	1,043	1,083	1,076	1,130	1,145	1,192
Commercial	839	856	851	886	914	954	981	1,028	1,079	1,105	1,135
Industrial	946	947	973	977	1,008	1,013	1,034	1,038	1,051	1,058	1,068
Transportation	4	4	4	4	4	4	4	4	4	4	4
U.S. Territories*	-	-	-	-	-	-	-	-	-	-	-
Total	2,713	2,762	2,763	2,861	2,935	3,013	3,101	3,146	3,264	3,312	3,398

*EIA data on fuel consumption for electricity generation does not include the U.S. territories.

- Not applicable

Source: EIA (2001a)

ANNEX B: Methodology for Estimating Carbon Stored in Products from Non-Energy Uses of Fossil Fuels

Carbon storage associated with the non-energy use of fossil fuels was calculated by multiplying each fuel's potential emissions (i.e., each fuel's total carbon content) by a fuel-specific storage factor. This Annex explains the methods and data sources employed in developing the storage factors for asphalt and road oil, lubricants, petrochemical feedstocks, liquefied petroleum gases (LPG), pentanes plus, and natural gas used for chemical manufacturing plant feedstocks (i.e., not used as fuel.) The storage factors for the remaining non-energy fuel uses are based on values reported by Marland and Rotty (1984).

Table B-1: Fuel Types and Percent of Carbon Stored for Non-Energy Uses

Fuel Type	Storage Factor
Industrial Coking Coal ^a	75%
Natural Gas to Chemical Plants	-
Nitrogenous Fertilizers	0%
Other Uses	63%
Asphalt & Road Oil	100%
Liquefied Petroleum Gas (LPG)	63%
Lubricants	9%
Pentanes Plus	63%
Petrochemical Feedstocks	-
Naphtha (b.p.<401° F)	63%
Other Oil (b.p.>401° F)	63%
Petroleum Coke ^b	50%
Special Naphtha	0%
Other	-
Distillate Fuel Oil	50%
Residual Fuel	50%
Waxes	100%
Miscellaneous Products	100%

- Not applicable

^a Includes processes for which specific coking coal consumption and emission factor data are not available. Consumption of coking coal for production of iron and steel is covered in the Industrial Processes chapter.

^b Includes processes for which specific petroleum coke consumption and emission factor data are not available (e.g., carbon fibers and textiles, refractory, electric motor parts, brake parts, batteries). Consumption of petroleum coke for production of primary aluminum anodes, electric arc furnace anodes, titanium dioxide, and ferroalloys is covered in the Industrial Processes chapter.

The following sections describe the selected non-energy uses in greater detail, outlining the methods employed and data used in estimating each storage factor. Several of the fuel types tracked by EIA—petrochemical feedstocks, pentanes plus, LPG, and natural gas—are used in organic chemical synthesis and in other manufacturing processes. Because the methods and data used to analyze them overlap, they are handled as a group and are discussed first. Discussions of the storage factors for asphalt and road oil and lubricants follow.

Petrochemical Feedstocks, Pentanes Plus, Liquefied Petroleum Gases, and Natural Gas

Petrochemical feedstocks, pentanes plus, liquefied petroleum gases (LPG) and natural gas¹ are used in the manufacture of a wide variety of man-made chemicals and products. Plastics, rubber, synthetic fibers, solvents, paints, fertilizers, pharmaceuticals, and food additives are just a few of the derivatives of these four fuel types. Chemically speaking, these fuels are diverse, ranging from simple natural gas (i.e., predominantly methane, CH₄) to heavier, more complex naphthas and other oils.² The storage factor for petrochemical feedstocks, pentanes plus, LPG, and natural gas used for purposes other than fuel is estimated based on data for the year 1998, which is the latest year for which data are available from several key data sources concerning consumption for non-fuel uses.

The four fuel categories constituted approximately 234.2 Tg CO₂ Eq., or 49 percent, of the 478.22 Tg CO₂ Eq. of non-energy fuel consumption in 1998 (including net exports of petrochemical feedstocks). Of this amount for the four fuels, 18.8 Tg CO₂ Eq. was exported, resulting in a net U.S. consumption of 215.4 Tg CO₂ Eq. in 1998. Of this net consumption, 136.3 Tg CO₂ Eq. ended up stored in products—including products subsequently combusted for waste disposal—while the remaining 79.1 Tg CO₂ Eq. was emitted to the atmosphere directly as CO₂ (e.g., through combustion of industrial byproducts) or indirectly as CO₂ precursors (e.g., through evaporative product use). The indirect emissions include a variety of organic gases such as volatile organic compounds (VOCs) and carbon monoxide (CO), which eventually oxidize into CO₂ in the atmosphere. For 1998 the storage factor for the four fuel categories was 63.3 percent; this factor was assumed to remain constant. The derivation of the storage factor is described in the following sections.

Methodology and Data Sources

An empirically determined storage factor was developed for the carbon consumed for non-energy end uses of petrochemical feedstocks, pentanes plus, LPG, and natural gas (henceforth referred to as feedstocks). The storage factor is equal to the ratio of carbon stored in the final products to total carbon content for the non-energy fossil fuel feedstocks used in industrial processes, after adjusting for net exports of feedstocks. Only one aggregate storage factor was calculated for the four fuel feedstock types. The feedstocks were grouped because of the overlap of their derivative products. Due to the many reaction pathways involved in producing petrochemical products (or wastes), it becomes extraordinarily complex to link individual products (or wastes) to their parent fuel feedstocks.

Import and export data for feedstocks were obtained from the Energy Information Administration (EIA) and from the National Petroleum Refiners Association (NPRA) for the major categories of petrochemical feedstocks (EIA 2001a, NPRA 2001). The EIA tracks imports and exports of petrochemical feedstocks in its publication *Petroleum Supply Annual* (EIA 2001c), including butanes, butylenes, ethane, ethylene, propane, propylene, LPG, and naphthas (i.e., most of the large volume primary chemicals produced by petroleum refineries). These imports and exports are already factored into the U.S. fuel consumption statistics. However, EIA does not track imports and exports of chemical intermediates and products produced by the chemical industry (e.g., xylenes, vinyl chloride, and polypropylene resins), which were derived from the primary chemicals produced by the refineries. For these products, the NPRA data were used. NPRA notes that their data set does not cover all imports and exports, due to the extremely broad range of products and the large number of importers and exporters; they estimate that their data set accounts for roughly 75 percent of the total mass imports and exports of feedstocks. Therefore, after calculating the net carbon flows from the NPRA data, the estimate was adjusted (i.e., by 100/75, or 1.33) to account for 100 percent of the total. Overall, the United States is a net exporter of chemical intermediates and products. Net exports

¹ Natural gas has two categories of non-energy consumption: for fertilizer and for other chemical syntheses. Only natural gas that is supplied to chemical plants for other uses is included here. Natural gas used for fertilizer is not included because it covered in the Industrial Processes chapter.

² Naphthas are compounds distilled from petroleum containing 4 to 12 carbon atoms per molecule and having a boiling point less than 401° F. Other oils are distillates containing 12 to 25 carbon atoms per molecule and having a boiling point greater than 401° F.

of these materials amounted to 18.8 Tg CO₂ in 1998, corresponding to 8 percent of the total domestic fuel feedstocks production of 234.2 Tg CO₂. Net domestic consumption of fuel feedstocks (i.e., adjusted for net exports of intermediates and products) amounted to 215.4 Tg CO₂ in 1998.

The overall storage factor for the feedstocks was determined by developing a mass balance on the carbon in feedstocks, and characterizing products, uses, and environmental releases as resulting in either storage or emissions. The total carbon in the system is estimated by multiplying net domestic consumption—reported by EIA and then adjusted by assuming 8 percent net exports—by the carbon content of each of the feedstocks (i.e., petrochemical feedstocks-naptha, petrochemical feedstocks-other oils, LPG, pentanes plus, natural gas). Carbon content values for the fuel feedstocks are discussed in Annex A.

Next, carbon pools and releases in a variety of products and processes were characterized. Plastics, synthetic rubber, synthetic fibers, carbon black, industrial non-methane volatile organic compound (NMVOC) emissions, industrial toxic chemical (i.e., TRI) releases, pesticides, and organic solvents were identified as the major product categories.³

The carbon in each product or waste produced was categorized as either stored or emitted. The aggregate storage factor is the carbon-weighted average of storage across fuel types. As discussed later in the section on uncertainty, data were not available for all of the non-energy end uses of fuel feedstocks, so the uses analyzed represent a sample of the total carbon consumed. The sample accounts for 195.0 Tg CO₂ Eq., or 91 percent, of the 215.4 Tg CO₂ Eq. of carbon within these fuel feedstock types that is consumed for non-energy purposes. The remaining 9 percent (18.8 Tg CO₂ Eq.) that is “unaccounted for” is assumed to be emitted, not stored. The total amount of carbon that is stored in products, including the amount subsequently lost to waste disposal, corresponds to 136.3 Tg CO₂ Eq. Emissions of CO₂ from waste disposal are accounted for separately in the Inventory and are discussed in the Waste Combustion section of the Energy chapter.

The following sections provide details on the calculation steps, assumptions, and data sources employed in estimating and classifying the carbon in each product and waste. Summing the carbon stored and dividing it by the total fuel feedstock carbon used yields the overall storage factor, as shown in Table B-2 and the equation below. The major products and their carbon contents are also shown in the Table.

$$\text{Overall Storage Factor} = \text{Carbon Stored} / \text{Total Carbon} = 136.3 \text{ Tg CO}_2 \text{ Eq.} / 215.4 \text{ Tg CO}_2 \text{ Eq.} = 63.3\%$$

Table B-2: Carbon Stored and Emitted by Products from Petrochemical Feedstocks, Pentanes Plus, LPG, and Natural Gas in 1998 (Tg CO₂ Eq.)

Product/Waste Type	Carbon Stored	Carbon Emitted
Plastics	110.4	-
Synthetic Rubber	7.7	-
Synthetic Fiber	11.8	-
Carbon Black	5.9	-
Pesticides	0.4	0.2
Industrial Releases	-	-
Industrial VOCs		4.0
TRI Releases	0.1	1.0
Non-Combustion CO	-	1.3
Solvent VOCs	-	9.2

³ For the most part, the releases covered by the U.S. Toxic Release Inventory (TRI) represent air emissions or water discharges associated with production facilities. Similarly, VOC emissions are generally associated with production facilities. These emissions could have been accounted for as part of the Waste chapter, but because they are not necessarily associated with waste management, they were included here. Toxic releases are not a “product” category, but they are referred to as such for ease of discussion.

Energy Recovery	-	41.1
Hazardous Waste Incin.	-	1.9
“Unaccounted for”*		20.4
Total	136.3	79.1

- Not applicable

* Unaccounted for carbon was assumed to be emitted.

Note: Totals may not sum due to independent rounding.

Plastics

Data on annual production of plastics were taken from the American Plastics Council, as published in *Chemical & Engineering News* and through direct communication with the APC (APC 2000, Eldredge-Roebuck 2000). These data were organized by year and resin type (see Table B-3). A carbon content was assigned for each resin. These contents were based on molecular formulas and are listed in Table B-4 and Table B-5. In cases where the resin type is generic, referring to a group of chemicals and not a single polymer (e.g., phenolic resins, other styrenic resins), a representative compound was chosen. For engineering resins and other resins, a weighted carbon content of 65 percent was assumed (i.e., it was assumed that these resins had the same content as those for which a representative compound could be assigned).

There were no emissive uses of plastics identified, so 100 percent of the carbon was considered stored in products. However, an estimate of emissions related to the combustion of these plastics in the municipal solid waste stream can be found in the Waste Combustion section of the Energy chapter.

Table B-3: 1998 Plastic Resin Production (Tg dry weight) and Carbon Stored (Tg CO₂ Eq.)

Resin Type	1998 Production ^a	Carbon Stored
Epoxy	0.29	0.8
Polyester (Unsaturated)	0.78	1.8
Urea	1.17	1.5
Melamine	0.13	0.1
Phenolic	1.79	5.0
Low-Density Polyethylene (LDPE)	3.44	10.8
Linear Low-Density Polyethylene (LLDPE)	3.28	10.3
High Density Polyethylene (HDPE)	5.86	18.4
Polypropylene (PP)	6.27	19.7
Acrylonitrile-butadiene-styrene (ABS)	0.65	2.0
Styrene-acrylonitrile (SAN)	0.06	0.2
Other Styrenics	0.75	2.5
Polystyrene (PS)	2.83	9.6
Nylon	0.58	1.4
Polyvinyl chloride (PVC) ^b	6.58	9.3
Thermoplastic Polyester	2.01	4.6
Engineering Resins	1.25	3.0
All Other	3.88	9.4
Total	41.59	110.4

^a Includes production from Canada for ABS, SAN, PVC, PP, Phenolic, Urea, Melamine, and Thermoplastic Polyester

^b Includes copolymers

Note: Totals may not sum due to independent rounding

Table B-4: Assigned Carbon Contents of Plastic Resins (by weight)

Resin Type	Carbon Content	Source of Carbon Content Assumption
Epoxy	76%	Typical epoxy resin made from epichlorhydrin and bisphenol A
Polyester (Unsaturated)	63%	Poly (ethylene terephthalate) (PET)

Urea	34%	50% carbamal, 50% N-(hydroxymethyl) urea *
Melamine	29%	Trimethylol melamine *
Phenolic	77%	Phenol
Low-Density Polyethylene (LDPE)	86%	Polyethylene
Linear Low-Density Polyethylene (LLDPE)	86%	Polyethylene
High Density Polyethylene (HDPE)	86%	Polyethylene
Polypropylene (PP)	86%	Polypropylene
Acrylonitrile-Butadiene-Styrene (ABS)	85%	50% styrene, 25% acrylonitrile, 25% butadiene
Styrene-Acrylonitrile (SAN)	80%	50% styrene, 50% acrylonitrile
Other Styrenics	92%	Polystyrene
Polystyrene (PS)	92%	Polystyrene
Nylon	65%	Average of nylon resins (see Table B-5)
Polyvinyl Chloride (PVC)	38%	Polyvinyl chloride
Thermoplastic Polyester	63%	Polyethylene terephthalate
Engineering Resins	66%	Weighted average of other resin production
All Other	66%	Weighted average of other resin production

*Does not include alcoholic hydrogens.

Table B-5: Major Nylon Resins and their Carbon Contents (by weight)

Resin	Carbon Content
Nylon 6	64%
Nylon 6,6	64%
Nylon 4	52%
Nylon 6,10	68%
Nylon 6,11	69%
Nylon 6,12	70%
Nylon 11	72%

Synthetic Rubber

Data on annual consumption of synthetic rubber were obtained from the International Institute of Synthetic Rubber Producers (IISRP) press release “Synthetic Rubber Use Growth to Continue Through 2004, Says IISRP and RMA” (IISRP 2000). Due to the fact that production data for synthetic rubber were unavailable, consumption was assumed to equal production. These data were organized by year and elastomer type. A carbon content was assigned for each elastomer type. These contents, based on stoichiometry, are listed in Table B-6. For the “Others” category, a weighted carbon content was calculated from total 1998 consumption data.

There were no emissive uses of rubber identified, so 100 percent of the carbon was assumed stored. However, emissions related to the combustion of scrap tires and rubber consumer goods can be found in the Waste Combustion section of the Energy chapter.

Table B-6: 1998 Rubber Consumption, Carbon Content, and Carbon Stored

Elastomer Type	1998 Consumption (Thousand Metric Tons) *	Carbon Content	Carbon Stored (Tg CO ₂ Eq.)
SBR Solid	908	91%	3.0
Polybutadiene	561	89%	1.8
Ethylene Propylene	320	86%	1.0
Polychloroprene	69	59%	0.1
NBR Solid	87	77%	0.2
Polyisoprene	78	88%	0.3
Others	369	88%	1.2
Total	2,392	-	7.7

* Includes consumption in Canada.

- Not applicable

Note: Totals may not sum due to independent rounding

Synthetic Fibers

Annual synthetic fiber production data were obtained from the Fiber Economics Bureau, as published in *Chemical & Engineering News* and exhibited on the FiberSource website (FEB 2000). These data are organized by year and fiber type. For each fiber, a carbon content was assigned based on stoichiometry (see Table B-7). For polyester, the carbon content for poly(ethylene terephthalate) (PET) was used as a representative compound. For nylon, the average carbon content of nylon 6 and nylon 6,6 was used, since these are the most widely produced nylon fibers. Cellulosic fibers, such as acetate and rayon, have been omitted from the synthetic fibers' carbon accounting because much of their carbon is of biogenic origin. These fibers account for only 4 percent of overall fiber production by weight.

There were no emissive uses of fibers identified, so 100 percent of the carbon was considered stored. However, emissions related to the combustion of textiles in the municipal solid waste stream is accounted for under the Waste Combustion section of the Energy chapter.

Table B-7: 1998 Fiber Production, Carbon Content, and Carbon Stored

Fiber Type	Production (Tg)	Carbon Content	Carbon Stored (Tg CO₂ Eq.)
Polyester	1.8	63%	4.1
Nylon	1.3	64%	3.0
Olefin	1.3	86%	4.1
Acrylic	0.2	68%	0.5
Total	4.6	-	11.7

- Not applicable

Note: Totals may not sum due to independent rounding

Carbon Black

Carbon black is a finely divided solid form of carbon produced from the partial oxidation of heavy oil fractions.⁴ It is used primarily in manufacture of tire treads and other abrasion resistant rubber products, but can also be used in pigments for paints and inks. In 1998, carbon black ranked 35th in chemical production in the United States with 1,610,280 metric tons produced (CMA 1999). Since carbon black is essentially pure carbon, its carbon content was assumed to be 100 percent. Also, since it is used in solid products and resists degradation, it was considered 100 percent stored. For 1998, carbon stored as a result of carbon black production was estimated to be 5.9 Tg CO₂ Eq.

Energy Recovery

The amount of fuel feedstocks that are combusted for energy recovery was estimated from data included in EIA's Manufacturers Energy Consumption Survey (MECS) for 1998 (EIA 2001b). Fuel feedstocks may be combusted for energy recovery because the chemical reactions in which fuel feedstocks are used are not 100 percent efficient. These chemical reactions may generate unreacted raw material feedstocks or generate byproducts that have a high energy content. The chemical industry and many downstream industries are energy-intensive, and therefore unreacted feedstocks or byproducts of production may be combusted for energy recovery in industrial boilers. Also, hazardous waste regulations provide a strong incentive—and in some cases require—burning of organic wastes generated from chemical production processes. Combustion of hazardous waste without energy recovery is referred to as "incineration," and is discussed separately in this Annex.

⁴ Carbon black can also be produced from the cracking of natural gas, but this method is uncommon.

MECS data include data on the consumption for energy recovery of “other” fuels in the petroleum and coal products, chemicals, primary metals, nonmetallic minerals, and other manufacturing sectors. These “other” fuels include refinery still gas; waste gas; waste oils, tars, and related materials; petroleum coke, coke oven and blast furnace gases; and other uncharacterized fuels. Fuel use of petroleum coke is included separately in the fuel use data provided annually by EIA, and energy recovery of coke oven gas and blast furnace gas (i.e., byproducts of the iron and steel production process) is addressed in the Iron and Steel production section in the Industrial Processes chapter. Consumption of refinery still gas and “other” fuels in the refinery sector is also included separately in the fuel use data from EIA. Therefore these categories of “other” fuels are addressed elsewhere in the inventory and not considered as part of the petrochemical feedstocks energy recovery discussion. The remaining categories of fuels included in the 1998 MECS data (Table B-8) including waste gas; waste oils, tars, and related materials; and other uncharacterized fuels are assumed to be petrochemical feedstocks burned for energy recovery. The conversion factors listed in Annex A were used to convert the Btu values for each fuel feedstock to Tg CO₂. Petrochemical feedstocks combusted for energy recovery corresponded to 41.1 Tg CO₂ Eq. in 1998.

Table B-8: Summary of 1998 MECS Data for Other Fuels used in Manufacturing/Energy Recovery (Trillion Btu)

Subsector and Industry	NAICS Code	Waste Gas ^b	Waste Oils/Tars ^c	Refinery Still Gas ^d	Net Steam ^e	Other Fuels ^{f,g}
Printing and Related Support	323	0	1	0	0	0
Petroleum and Coal Products	324	0	1	1399	0	324 ^a
Chemicals	325	416	16	0	194	118 ^a
Plastics and Rubber Products	326	0	0	0	0	0
Nonmetallic Mineral Products	327	2	9	0	0	14
Primary Metals	331	2	2	0	0	22
Fabricated Metal Products	332	1	0	0	0	0
Machinery	333	0	1	0	0	0
Computer and Electronic Products	334	0	0	0	0	0
Electrical Equipment, Appliances, Components	335	1	1	0	0	0
Transportation Equipment	336	1	2	0	0	0
Total		423	33	1399	194	478
Carbon Content (Tg/QBTU)		18.1	20.6	17.5	0	19.4
Fraction Oxidized		99%	99%	99%	0%	99%
Total Carbon (Tg)		7.6	0.7	24.2	0.0	9.2
Total Carbon (Non-Refining)		7.6	0.7	0.0	0.0	3.0

^a EIA personal communication, 2001.

^b C content assumed to be naphtha <401F.

^c C content assumed to be asphalt and road oil.

^d Refinery “other” fuel consumption is reported elsewhere in the inventory and is excluded from the total carbon. The total non-refining total carbon excludes all “other” fuel consumption under NAICS Code 325 (Petroleum and Coal Products).

^e Net steam is reported elsewhere in the inventory and is excluded from the total carbon content estimate.

^f Assumed to be petrochemical feedstocks.

^g Includes net steam, except for NAICS Code 325 (Chemicals). Net steam for other sectors assumed to be negligible.

Note: Totals may not sum due to independent rounding.

Industrial and Solvent Evaporation Volatile Organic Compound Emissions

Data on annual non-methane volatile organic compound (NMVOC) emissions were obtained from the *National Air Quality and Emissions Trends Report* (EPA 2000b). NMVOC emissions were organized by end-use category. NMVOC emissions from solvent utilization were considered to be a result of non-energy use of petrochemical feedstocks. The end-use categories that represent “Industrial NMVOC Emissions” include chemical and allied products, petroleum and related industries, and other industrial processes. These categories are non-energy uses of the four fuel types; other categories where VOCs could be emitted due to combustion of fossil fuels were excluded to avoid double counting.

Because solvent evaporation and industrial NMVOC emission data are provided in tons of total NMVOCs, assumptions were made concerning the average carbon content of the NMVOCs for each category of emissions. The assumptions for calculating the carbon fraction of industrial and solvent utilization emissions were made separately and differ significantly. For industrial NMVOC emissions, the carbon content of 85 percent was assumed. This value was chosen to reflect the carbon content of an average volatile organic compound based on the list of the most abundant, NMVOCs provided in the *Trends Report*. The list contains only pure hydrocarbons, including saturated alkanes (carbon contents ranging from 80 to 85 percent based upon carbon number), alkenes (carbon contents equal 85.7 percent), and some aromatics (carbon contents approximately 90 percent, depending upon substitution).

EPA (2000b) solvent evaporation emissions data were used directly to estimate the carbon content of solvent emissions. The EPA data identify solvent emissions by compound or compound category for six different solvent end-use categories: degreasing, graphic arts, dry cleaning, surface coating, other industrial processes, and non-industrial processes. The percent carbon of each compound identified in the EPA solvent evaporation emissions data was calculated based on the stoichiometry of the individual compound (e.g., the carbon content of methylene chloride is 14 percent; the carbon content of toluene is 91 percent). For solvent emissions that are identified in the EPA data only by chemical category (e.g., butanediol derivatives) a single individual compound was selected to represent each category, and the carbon content of the category was estimated based on the carbon content of the representative compound. The overall carbon content of the solvent evaporation emissions is estimated to be 56 percent.

The results of the industrial and solvent NMVOC emissions analysis are provided in Table B-9.

Table B-9: 1998 Industrial and Solvent NMVOC Emissions

Activity	Thousand short tons	Carbon Content	Carbon Emitted (Tg CO₂ Eq.)
Industrial NMVOCs ^a	1,417	85%	4.0
Solvent Evaporation ^b	4,950	56%	9.3

a Includes emissions from chemical and allied products, petroleum and related industries, and other industrial processes categories.

b Includes solvent usage and solvent evaporation emissions from degreasing, graphic arts, dry cleaning, surface coating, other industrial processes, and non-industrial processes.

Hazardous Waste Incineration

Hazardous wastes are defined by the EPA under the Resource Conservation and Recovery Act (RCRA).⁵ Industrial wastes, such as rejected products, spent reagents, reaction by-products, and sludges from wastewater or air pollution control, are federally regulated as hazardous wastes if they are found to be ignitable, corrosive, reactive, or toxic according to standardized tests or studies conducted by the EPA.

Hazardous wastes must be treated prior to disposal according to the federal regulations established under the authority of RCRA. Combustion is one of the most common techniques for hazardous waste treatment, particularly for those wastes that are primarily organic in composition or contain primarily organic contaminants. Generally speaking, combustion devices fall into two categories: incinerators that burn waste solely for the purpose of waste management, and boilers and industrial furnaces (BIFs) that burn waste in part to recover energy from the waste. More than half of the hazardous waste combusted in the U.S. is burned in BIFs; these processes are included in the energy recovery calculations described above.

The EPA's Office of Solid Waste requires biennial reporting of hazardous waste management activities, and these reports provide estimates of the amount of hazardous waste burned for incineration or energy recovery.

⁵ [42 U.S.C. §6924, SDWA §3004]

EPA stores this information in its Biennial Reporting System (BRS) database (EPA 2000a). Combusted hazardous wastes are identified based on EPA-defined management system types M041 through M049 (incineration). Combusted quantities are grouped into four representative waste form categories based on the form codes reported in the BRS: aqueous liquids, organic liquids and sludges, organic solids, and inorganic solids. To relate hazardous waste quantities to carbon emissions, “fuel equivalent” factors were derived for hazardous waste by assuming that the hazardous wastes are simple mixtures of a common fuel, water, and noncombustible ash. For liquids and sludges, crude oil is used as the fuel equivalent and coal is used to represent solids.

Fuel equivalent factors were multiplied by the tons of waste incinerated to obtain the tons of fuel equivalent. Multiplying the tons of fuel equivalent by the appropriate carbon content factors from Marland and Rotty (1984) yields tons of carbon emitted. Implied carbon content is calculated by dividing the tons of carbon emitted by the associated tons of waste incinerated.

Waste quantity data for hazardous wastes were obtained from the EPA’s BRS database for reporting years 1989, 1991, 1993, 1995, and 1997 (EPA 2000a). Combusted waste quantities were obtained from Form GM (Generation and Management) for wastes burned on site and Form WR (Wastes Received) for waste received from off-site for combustion. For each of the waste types, assumptions were developed on average waste composition (see Table B-10). Carbon emission factors for equivalent fuels were obtained from Marland and Rotty (1984). Regulations require incinerators to achieve at least 99.99 percent destruction of organics; this formed the basis for assuming the fraction of carbon oxidized. A least-squares linear regression from the time series 1989 through 1997 was used to estimate emissions for 1998.

Table B-10: Assumed Composition of Combusted Hazardous Waste by Weight (Percent)

Waste Type	Water	Noncombustibles	Fuel Equivalent
Aqueous Waste	90	5	5
Organic Liquids and Sludges	40	20	40
Organic Solids	20	40	40
Inorganic Solids	20	70	10

Non-Combustion Carbon Monoxide Emissions

Carbon monoxide (CO) emissions data were obtained from the National Air Quality and Emissions Trends Report (EPA 2000a). There are four categories of CO emissions in EPA (2000a) that are classified as process-related emissions not related to fuel combustion. These include chemical and allied products manufacturing, metals processing, petroleum and related industries, and other industrial processes. Some of these CO emissions are accounted for in the Industrial Processes section of this report, and are therefore not accounted for in this section, including total carbon emissions from the primary aluminum, titanium dioxide, iron and steel, and ferroalloys production processes. The total carbon (CO and CO₂) emissions from oil and gas production and asphalt manufacturing are also accounted for elsewhere in this Inventory. Sustainably harvested biogenic emissions (e.g., pulp and paper process emissions) are also excluded from calculation of CO emissions in this section. Those CO emissions that are not accounted for elsewhere are considered to be byproducts of non-fuel use of feedstocks and are included in the calculation of the petrochemical feedstocks storage factor. Table B-11 lists the industrial processes and CO emissions that remain after excluding those reported in the Industrial Processes chapter, or where the carbon originates from sustainably harvested biogenic sources.

Table B-11: Carbon Monoxide Non-Combustion Emissions (Gg)

Source	1998
Chemical and Allied Products	
Organic Chemical Manufacture	83.6
Inorganic Chemical Manufacture, Other	2.7
Polymer and Resin Manufacture	4.5
Agricultural Chemical Manufacture	11.8
Paint, Varnish, Lacquer Manufacture	-
Pharmaceutical Manufacture	-
Other Chemical Manufacture	23.6
Metals Processing	

Nonferrous Metals Processing, Other	145.5
Metals Processing NEC	40.0
Petroleum and Related Industries	
Petroleum Refineries and Related Industry	300.9
Other Industrial Processes	
Rubber and Misc. Plastic Products	-
Mineral Products	169.1
Machinery Products	0.91
Electronic Equipment	-
Transportation Equipment	-
Misc. Industrial Processes	18.2
Total	801.7
Total (Tg CO₂ Eq.)	1.3

Pesticides

Pesticide consumption data were obtained from the *1996/1997 Pesticides Industry Sales and Usage* (EPA 1999) report. Although some production data were available, consumption data were used because these data provided information on active ingredients. Active ingredient compound names and consumption weights were available for the top 25 agriculturally-used pesticides and top 9 pesticides used in the home and garden and the industry/commercial/government categories. Since the report provides a range of consumption for each active ingredient, the midpoint was used to represent actual consumption. Each of these compounds was assigned a carbon content value based on stoichiometry. If the compound contained aromatic rings substituted with chlorine or other halogens, then the compound was considered persistent and assigned a 100 percent carbon storage factor. All other pesticides were assumed to release their carbon to the atmosphere. Nearly one-third of total pesticide active ingredient consumption was not specified by chemical type in the *Sales and Usage* report (EPA 1999). This unspecified portion of the active ingredient consumption was treated as a single chemical and assigned a carbon content and a storage factor based on the weighted average of the known chemicals' values.

Table B-12: Active Ingredient Consumption in Pesticides (Million lbs.) and Carbon Emitted and Stored (Tg CO₂ Eq.)

Pesticide Use	Active Ingredient	Carbon Emitted	Carbon Stored
Agricultural Uses ^a	551.0	0.1	0.2
Non-Agricultural Uses ^b	84.5	+	+
Home & Garden	34.0	+	+
Industry/Gov't/Commercial	50.5	+	+
Other	334.5	0.1	0.1
Total	970.0	0.2	0.4

+ Less than 0.05 Tg CO₂ Eq.

^a 1997 estimate (EPA 1999).

^b Approximate quantities, 1995/1996 estimates (EPA 1999).

Note: Totals may not sum due to independent rounding.

TRI Releases

Carbon is also found in toxic substances released by industrial facilities. The Toxic Release Inventory (TRI), maintained by the EPA, tracks these releases by chemical and environmental release medium (i.e., land, air, or water) on a biennial basis (EPA 2000b). By examining the carbon contents and receiving media for the top 35 toxic chemicals released, which account for 90 percent of the total mass of chemicals, the quantity of carbon stored and emitted in the form of toxic releases can be estimated.

The TRI specifies releases by chemical, so carbon contents were assigned to each chemical based on stoichiometry. The TRI also classifies releases by disposal location as either off-site or on-site. The on-site releases are further subdivided into air emissions, surface water discharges, underground injection, and releases to land; the latter is

further broken down to disposal in a RCRA Subtitle C (i.e., hazardous waste) landfill or to “Other On-Site Land Disposal.”⁶ The carbon released in each disposal location is provided in Table B-13.

Each on-site classification was assigned a storage factor. A one hundred percent storage factor was applied to disposition of carbon to Underground Injection and to Disposal to RCRA-permitted Landfills, while the other disposition categories were assumed to result in an ultimate fate of emission as CO₂ (i.e., a storage factor of zero was applied to these categories.) The release allocation is not reported for off-site releases; therefore, the approach was to develop a carbon-weighted average storage factor for the on-site carbon and apply it to the off-site releases.

For the remaining 10 percent of the TRI releases, the weights of all chemicals were added and an average carbon content value, based upon the top 35 chemicals’ carbon contents, was applied. The storage and emission allocation for the remaining 10 percent of the TRI releases was carried out in the same fashion as for the 35 major chemicals.

Table B-13: 1998 TRI Releases by Disposal Location (Gg CO₂ Eq.)

Disposal Location	Carbon Stored	Carbon Emitted
Air Emissions	-	924.0
Surface Water Discharges	-	6.7
Underground Injection	89.4	-
RCRA Subtitle C Landfill Disposal	1.4	-
Other On-Site Land Releases	-	15.9
Off-site Releases	6.4	36.0
Total	97.2	982.6

- Not applicable

Note: Totals may not sum due to independent rounding

Uncertainty

There are several cross-cutting sources of uncertainty that pervade the characterization of a storage factor for feedstocks. The aggregate storage factor for petrochemical feedstocks, pentanes plus, liquefied petroleum gases, and natural gas is based on only a partial sampling of the products derived from these fossil fuel feedstocks and imports and exports of petrochemical feedstocks. Including consideration of petrochemical feedstocks that are exported and feedstocks that are burned for energy recovery, approximately 91 percent of the carbon consumed across these four fuel types for non-energy uses is accounted for. The remaining “unaccounted-for” carbon could have a variety of fates. For the purposes of this inventory, all of the unaccounted for carbon was assumed to be emitted, and using this assumption the overall storage factor is 63.3 percent. If the assumption had been made that the products which contained the unaccounted for carbon would store and emit carbon in the same ratio as the investigated products, the overall storage factor would have been 69.9 percent, rather than 63.3 percent.

With respect to the “unaccounted for” carbon, there are uncertainties associated with the EIA and NPRA data for net imports and exports of feedstocks that could affect the calculation of the storage factor. To a lesser extent, there are uncertainties associated with the simplifying assumptions made for each end use category carbon estimate. Generally, the estimate for a product is subject to one or both of the following uncertainties:

- The value used for estimating the carbon content has been assumed or assigned based upon a representative compound.
- The split between carbon storage and emission has been assumed based on an examination of the environmental fate of the products in each end use category.

⁶ Only the top 9 chemicals had their land releases separated into RCRA Landfills and Other Land Disposal. For the remaining chemicals, it was assumed that the ratio of disposal in these two categories was equal to the carbon-weighted average of the land disposal fate of the top 9 chemicals (i.e., 8 percent attributed to RCRA Landfills and 92 percent in the “Other” category).

Another cross-cutting source of uncertainty is that the estimate of the balance between storage and emissions is based on data for only for a single year, 1998. This specific year may not be representative of storage for the entire decade.

Sources of uncertainty associated with specific elements of the analysis are discussed below.

Imports/Export. Import and export data for petrochemical feedstocks were obtained from EIA and the National Petroleum Refiners Association for the major categories of petrochemical feedstocks (EIA 2001a, NPRA 2001). However, the NPRA believes that the import and export data that they provide account for only 75 percent of the total imports and exports of petrochemical feedstocks. The NPRA data provided were adjusted to account for 100 percent of the total, though this adjustment represents a source of uncertainty in the estimate of net exports. Net exports could be lower or higher than that estimated depending upon how comprehensive the combined EIA and NPRA data are in tracking total imports and exports of petrochemical feedstocks.

Oxidation Factors. Oxidation factors have been applied to non-energy uses of petrochemical feedstocks in the same manner as for energy uses. However, this “oxidation factor” may be inherent in the storage factor applied when calculation emissions from non-energy consumption, which would result in a double-counting of the unoxidized carbon. Oxidation factors are small corrections, on the order of 1 percent, and therefore application of oxidation factors to non-energy uses may result in a slight underestimation of carbon emissions from non-energy uses.

Plastics. Uncertainty in the carbon storage estimate for plastics arises primarily from three factors. First, the production data for acrylonitrile-butadiene-styrene, styrene-acrylonitrile, polyvinyl chloride, polypropylene, phenolic, urea, melamine, and thermoplastic polyester resins include Canadian production and may overestimate the amount of plastic produced from U.S. fuel feedstocks. Second, the assumed carbon content values are estimates for representative compounds, and thus do not account for the many formulations of resins available. This uncertainty is greater for resin categories that are generic (e.g., phenolics, other styrenics, nylon) than for resins with more specific formulations (e.g., polypropylene, polyethylene). Lastly, the assumption that all of the carbon contained in plastics is stored ignores certain end uses (e.g., adhesives and coatings) where the resin may be released to the atmosphere; however, these end uses are likely to be small relative to use in plastics.

Rubber. Similar to plastics, uncertainty results from using consumption data for the United States and Canada, rather than just domestic consumption, which may overestimate the amount of rubber produced from U.S. fuel feedstocks. There are also uncertainties as to the assignment of carbon content values; however, they are much smaller than in the case of plastics. There are probably fewer variations in rubber formulations than in plastics, and the range of potential carbon content values is much narrower. Lastly, assuming that all of the carbon contained in rubber is stored ignores the possibility of volatilization or degradation during product lifetimes. However, the proportion of the total carbon that is released to the atmosphere during use is probably negligible.

Fiber. A small degree of uncertainty arises from the assignment of carbon content values; however, the magnitude of this uncertainty is less than that for plastics or rubber. Although there is considerable variation in final textile products, the stock fiber formulations are standardized and proscribed explicitly by the Federal Trade Commission.

Energy Recovery. The amount of feedstocks combusted for energy recovery was estimated from data included in the Manufacturers Energy Consumption Survey (MECS) for 1998 (EIA 2001b). MECS is a comprehensive survey intended to represent U.S. industry as a whole, but because EIA does not receive data from all manufacturers (i.e., it is a sample rather than a census), EIA must extrapolate from the sample. Also, the “other” fuels are identified in the 1998 MECS data in broad categories, including refinery still gas; waste gas; waste oils, tars, and related materials; petroleum coke, coke oven and blast furnace gases; and other uncharacterized fuels, and the industries using these “other” fuels are also identified only in broad categories, including the petroleum and coal products, chemicals, primary metals, nonmetallic minerals, and other manufacturing sectors. The “other” fuel consumption data are reported in BTUs (energy units) and there is uncertainty concerning the selection of a specific conversion factor for each broad “other” fuel category to convert energy units to mass units.

NM VOCs. (Solvent Evaporation and Industry): The principal sources of uncertainty in estimating CO₂ emissions from solvent evaporation and industry are in the estimates of total NMVOC emissions and in the application of

factors for the carbon content of these emissions. Solvent evaporation and industrial NMVOC emissions reported by EPA are based on a number of data sources and emission factors, and may underestimate or overestimate emissions. The carbon content for solvent evaporation emissions is calculated directly from the specific solvent compounds identified by EPA as being emitted, and is thought to have relatively low uncertainty. The carbon content for industrial emissions has more uncertainty, however, as it is calculated from the average carbon content of an average volatile organic compound based on the list of the most abundant measured NMVOCs provided in EPA (2000b).

Hazardous Waste. The greatest uncertainty in the hazardous waste combustion analysis is introduced by the assumptions about the composition of combusted hazardous wastes, including the characterization that hazardous wastes are similar to mixtures of water, noncombustibles, and fuel equivalent materials. Another limitation is the assumption that all of the carbon that enters hazardous waste combustion is emitted—some small fraction is likely to be sequestered in combustion ash—but given that the destruction and removal efficiency for hazardous organics is required to meet or exceed 99.99 percent, this is a minor source of uncertainty. Carbon emission estimates from hazardous waste should be considered central value estimates that are likely to be accurate to within ± 50 percent.

Pesticides. The largest source of uncertainty involves the assumption that a pesticide's active ingredient carbon is either 0 percent stored or 100 percent stored. This split is a generalization of chemical behavior, based upon active-ingredient molecular structure, and not on compound-specific environmental data. The mechanism by which a compound is bound or released from soils is very complicated and can be affected by many variables, including the type of crop, temperature, delivery method, and harvesting practice. Another smaller source of uncertainty arises from the carbon content values applied to the unaccounted for portion of active ingredient. Carbon contents vary widely among pesticides, from 7 to 72 percent, and the remaining pesticides may have a chemical make-up that is very different from the 32 pesticides that have been examined.

TRI. The major uncertainty in using the TRI data are the possibility of double counting of emissions that are already accounted for in the NMVOC data (see above) and in the storage and emission assumptions used. The approach for predicting environmental fate simplifies some complex processes, and the balance between storage and emissions is very sensitive to the assumptions on fate. Extrapolating from known to unknown characteristics also introduces uncertainty. The two extrapolations with the greatest uncertainty are: 1) that the release media and fate of the off-site releases were assumed to be the same as for on-site releases, and 2) that the carbon content of the least frequent 10 percent of TRI releases was assumed to be the same as for the chemicals comprising 90 percent of the releases. However, the contribution of these chemicals to the overall estimate is small. The off-site releases only account for 3 percent of the total releases, by weight, and, by definition, the less frequent compounds only account for 10 percent of the total releases.

Asphalt and Road Oil

Asphalt is one of the principal non-energy uses of fossil fuels. The term “asphalt” generally refers to a mixture of asphalt cement and a rock material aggregate, a volatile petroleum distillate, or water. For the purposes of this analysis, “asphalt” is used interchangeably with asphalt cement, a residue of crude oil. According to EIA (2000d), approximately 100 Tg CO₂ Eq. has been used in the production of asphalt cement annually. Though minor amounts of carbon are emitted during production, asphalt has an overall carbon storage factor of almost 100 percent.

Paving is the primary application of asphalt cement, comprising 86 percent of production. The three types of asphalt paving produced in the United States are hot mix asphalt (HMA), cut-backs, and emulsified asphalt. HMA, which makes up 90 percent of total asphalt paving (EPA 2000d), contains asphalt cement mixed with an aggregate of rock materials. Cut-back asphalt is composed of asphalt cement thinned with a volatile petroleum distillate (e.g., naphtha). Emulsified asphalt contains only asphalt cement and water. Roofing products are the other significant end use of asphalt cement, accounting for approximately 14 percent of U.S. production (Kelly 2000). No data were available on the fate of carbon in asphalt roofing; it was assumed that it has the same fate as carbon in asphalt paving applications.

Methodology

A carbon storage factor was calculated for each type of asphalt paving. The fraction of carbon emitted by each asphalt type was multiplied by consumption data for asphalt paving (EPA 2000d, EIIP 1998) to come up with a weighted average carbon storage factor for asphalt as a whole.

The fraction of carbon emitted by HMA was determined by first calculating the organic emissions (volatile organic compounds [VOCs], carbon monoxide, polycyclic aromatic hydrocarbons [PAHs], hazardous air pollutants [HAPs], and phenol) from HMA paving, using emission factors reported in EPA (2000d) and total HMA production.⁷ The next step was to estimate the carbon content of the organic emissions. This calculation was based on stoichiometry for carbon monoxide (CO) and phenol, and an assumption of 85 percent carbon content for PAHs and HAPs. The carbon content of asphalt paving is a function of the proportion of asphalt cement in asphalt paving and the proportion of carbon in asphalt cement. For the former factor, a 5 percent asphalt cement content was assumed based on personal communication with an expert from the National Asphalt Paving Association (Connolly 2000). For the latter factor, all paving types were characterized as having a mass fraction of 85 percent carbon in asphalt cement, based on the assumption that asphalt is primarily composed of saturated paraffinic hydrocarbons. By combining these estimates, the result is that over 99.99 percent of the carbon in asphalt cement was retained (i.e., stored), and less than 0.01 percent was emitted.

Cut-back asphalt is produced in three forms (i.e., rapid, medium and slow cure). All three forms emit carbon only from the volatile petroleum distillate used to thin the asphalt cement (EPA 1995). Because the petroleum distillates are not included in the EIA statistics for asphalt, the storage factor for cut-back is assumed to be 100 percent.

It was also assumed that there was no loss of carbon from emulsified asphalt (i.e., the storage factor is 100 percent) based on personal communication with an expert from Akzo Nobel Coatings, Inc. (James 2000).

Data Sources

Data on asphalt and road oil consumption and carbon content factors were supplied by the EIA. Hot mix asphalt production and emissions factors were obtained from “Hot Mix Asphalt Plants Emissions Assessment Report” from the EPA publication *AP-42* (EPA 2000d). The asphalt cement content of HMA was provided by Una Connolly of National Asphalt Paving Association (Connolly 2000). The consumption data for cut-back and emulsified asphalts were taken from a Moulthrop, et al. study used as guidance for estimating air pollutant emissions from paving processes (EIIP 1998). “Asphalt Paving Operation” *AP-42* (EPA 1995) provided the emissions source information used in the calculation of the carbon storage factor for cut-back asphalt. The storage factor for emulsified asphalt was provided by Alan James of Akzo Nobel Coatings, Inc. (James 2000).

Uncertainty

The principal source of uncertainty is that the available data are from short-term studies of emissions associated with the production and application of asphalt. As a practical matter, the cement in asphalt deteriorates over time, contributing to the need for periodic re-paving. Whether this deterioration is due to physical erosion of the cement and continued storage of carbon in a refractory form or physicochemical degradation and eventual release of CO₂ is uncertain. Long-term studies may reveal higher lifetime emissions rates associated with degradation.

Many of the values used in the analysis are also uncertain and are based on estimates and professional judgment. For example, the asphalt cement input for HMA was based on expert advice indicating that the range is variable—from about 3 to 5 percent—with actual content based on climate and geographical factors (Connolly 2000). Over this

⁷ The emission factors are expressed as a function of asphalt paving tonnage (i.e., including the rock aggregate as well as the asphalt cement).

range, the effect on the calculated carbon storage factor is minimal (on the order of 0.1 percent). Similarly, changes in the assumed carbon content of asphalt cement would have only a minor effect.

The consumption figures for cut-back and emulsified asphalts are based on information reported for 1994. More recent trends indicate a decrease in cut-back use due to high VOC emission levels and a related increase in emulsified asphalt use as a substitute. However, because the carbon storage factor of each is 100 percent, use of more recent data would not affect the overall result.

Lubricants

Lubricants are used in industrial and transportation applications. They can be subdivided into oils and greases, which differ in terms of physical characteristics (e.g., viscosity), commercial applications, and environmental fate. According to EIA (2000), the carbon content of U.S. production of lubricants in 1999 was approximately 28 Tg CO₂ Eq. Based on apportioning oils and greases to various environmental fates, and characterizing those fates as resulting in either long-term storage or emissions, the overall carbon storage factor was estimated to be 9 percent; thus, storage in 1999 was about 3 Tg CO₂ Eq.

Methodology

For each lubricant category, a storage factor was derived by identifying disposal fates and applying assumptions as to the disposition of the carbon for each practice. An overall lubricant carbon storage factor was calculated by taking a production-weighted average of the oil and grease storage factors.

Oils

Regulation of used oil in the United States has changed dramatically over the past 15 years.⁸ The effect of these regulations and policies has been to restrict land filling and dumping, and to encourage collection of used oil. Given the relatively inexpensive price of crude oil, the economics have not favored re-refining—instead, most of the used oil that has been collected has been combusted.

Table B-14 provides an estimated allocation of the fates of lubricant oils, along with an estimate of the proportion of carbon stored in each fate. The ultimate fate of the majority of oils (about 84 percent) is combustion, either during initial use or after collection as used oil. Combustion results in 99 percent oxidation to CO₂, with correspondingly little long-term storage of carbon in the form of ash. Dumping onto the ground or into storm sewers, primarily by “do-it-yourselfers” who change their own oil, is another fate that results in conversion to CO₂ given that the releases are generally small and most of the oil is biodegraded. In the landfill environment, which tends to be anaerobic, within municipal landfills, it is assumed that 90 percent of the oil persists in an undegraded form. Re-refining adds a recycling loop to the fate of oil; it was assumed that about 97 percent of the carbon in re-refined oil is ultimately oxidized. Because of the dominance of fates that result in eventual release as CO₂, only about 3 percent of the carbon in oil lubricants goes into long-term storage.

Table B-14: Commercial and Environmental Fate of Oil Lubricants (Percent)

⁸ For example, the U.S. EPA “RCRA (Resource Conservation and Recovery Act) On-line” web site (<http://www.epa.gov/rcraonline/>) has over 50 entries on used oil regulation and policy for 1994 through 2000.

Fate of Oil	Portion of Total Oil	Carbon Stored
Combusted During Use	20	1
Not Combusted During Use	80	-
Combusted as Used Oil *	64	1
Dumped on the ground or in storm sewers	6	0
Landfilled	2	90
Re-refined into lube oil base stock and other products	8	3
Weighted Average	-	2.9

* (e.g., in boilers or space heaters)

- Not applicable

Greases

Table B-15 provides analogous estimates for lubricant greases. Unlike oils, grease is generally not combusted during use, and combustion for energy recovery and re-refining are thought to be negligible. Although little is known about the fate of waste grease, it was assumed that 90 percent of the non-combusted portion is landfilled, and the remainder is dumped onto the ground or storm sewers. Because much of the waste grease will be in containers that render it relatively inaccessible to biodegradation, it was assumed that 90 percent and 50 percent of the carbon in landfilled and dumped grease, respectively, would be stored. The overall storage factor is 82 percent for grease.

Table B-15: Commercial and Environmental Fate of Grease Lubricants (Percent)

Fate of Grease	Total Grease	Carbon Stored
Combusted During Use	5	1
Not Combusted During Use	95	-
Landfilled	85.5	90
Dumped on the ground or in storm sewers	9.5	50
Weighted Average	-	81.8

- Not applicable

Having derived separate storage factors for oil and grease, the last step was to estimate the weighted average for lubricants as a whole. No data were found apportioning the mass of lubricants into these two categories, but the U.S. Census Bureau does maintain records of the value of production of lubricating oils and lubricating greases. Assuming that the mass of lubricants can be allocated according to the proportion of value of production (92 percent oil, 8 percent grease), applying these weights to the storage factors for oils and greases (3 percent and 82 percent) yields an overall storage factor of 9 percent.

Data Sources

The estimated volume of lubricants produced annually is based on statistics provided by EIA (2000), which conducts surveys of lubricating oil and grease consumption. Information on the value of lubricating oil and grease production was obtained from reports by the U.S. Census Bureau (1999).

The characterization of fate is based primarily on professional judgment of an EPA regulatory analyst with experience in used oil (Rinehart 2000). For the proportions combusted, one percent was assumed to remain unoxidized in combustion processes (EIIP 1999); for other fates, estimates are based on professional judgment. The assumption that landfilled oil and grease results in 90 percent storage is based on analogy with the persistence of petroleum in native petroleum-bearing strata, which are both anaerobic. The assumption that oil dumped on the ground or in storm sewers is completely degraded is based on the observation that land farming—application to soil—is one of the most frequently used methods for degrading refinery wastes. The lower degradation rate for grease is based on the observation that greases contain longer chain paraffins, which are more persistent. Re-refined oil was assumed to have a storage factor equal to the weighted average for the other fates (i.e., after re-refining, the oil would have the same probability of combustion, landfilling, or dumping as virgin oil).

Uncertainty

The principal sources of uncertainty for the disposition of lubricants are the estimates of the commercial use, post-use, and environmental fate of lubricants, which, as noted above, are largely based on assumptions and judgment. There is no comprehensive system to track used oil and greases, which makes it difficult to develop a verifiable estimate of the commercial fates of oil and grease. The environmental fate estimates for percent of carbon stored are less uncertain, but also introduce uncertainty in the estimate.

The assumption that the mass of oil and grease can be divided according to their value also introduces uncertainty. Given the large difference between the storage factors for oil and grease, changes in their share of total lubricant production has a large effect on the weighted storage factor.

ANNEX C: Methodology for Estimating Emissions of CH₄, N₂O, and Ambient Air Pollutants from Stationary Combustion

Estimates of CH₄ and N₂O Emissions

Methane (CH₄) and nitrous oxide (N₂O) emissions from stationary combustion were estimated using IPCC emission factors and methods. Estimates were obtained by multiplying emission factors—by sector and fuel type—by fossil fuel and wood consumption data. This “top-down” methodology is characterized by two basic steps, described below. Data are presented in Table C-1 through Table C-5.

Step 1: Determine Energy Consumption by Sector and Fuel Type

Greenhouse gas emissions from stationary combustion activities were grouped into four sectors: industrial, commercial, residential, and electricity generation. For CH₄ and N₂O, estimates were based upon consumption of coal, gas, oil, and wood. Energy consumption data were obtained from EIA’s *Annual Energy Review* (EIA 2001), and adjusted from higher to lower heating values by multiplying by 0.9 for natural gas and by 0.95 for coal and petroleum fuel. This is a simplified convention used by the International Energy Agency. Table C-1 provides annual energy consumption data for the years 1990 through 2000.

Step 2: Determine the Amount of CH₄ and N₂O Emitted

Activity data for each sector and fuel type were then multiplied by emission factors to obtain emissions estimates. Emission factors were taken from the *Revised 1996 IPCC Guidelines* (IPCC/UNEP/OECD/IEA 1997). Table C-2 provides emission factors used for each sector and fuel type.

Estimates of NO_x, CO, and NMVOC Emissions

For ambient air pollutants, the major source categories included were those identified in EPA (2001): coal, fuel oil, natural gas, wood, other fuels (i.e., bagasse, liquefied petroleum gases, coke, coke oven gas, and others), and stationary internal combustion, which includes emissions from internal combustion engines not used in transportation. EPA (2001) periodically estimates emissions of NO_x, CO, and NMVOCs by sector and fuel type using a “bottom-up” estimating procedure. In other words, the emissions were calculated either for individual sources (e.g., industrial boilers) or for many sources combined, using basic activity data (e.g., fuel consumption or deliveries, etc.) as indicators of emissions. EPA (2001) projected emissions for years subsequent to their bottom-up estimates. The national activity data used to calculate the individual categories were obtained from various sources. Depending upon the category, these activity data may include fuel consumption or deliveries of fuel, tons of refuse burned, raw material processed, etc. Activity data were used in conjunction with emission factors that relate the quantity of emissions to the activity.

Table C-3 through Table C-5 present ambient air pollutant emission estimates for 1990 through 2000.

The basic calculation procedure for most source categories presented in EPA (2001) is represented by the following equation:

$$E_{p,s} = A_s \times EF_{p,s} \times (1 - C_{p,s}/100)$$

where,

- E = emissions
- p = pollutant
- s = source category
- A = activity level
- EF = emission factor
- C = percent control efficiency

The EPA currently derives the overall emission control efficiency of a category from a variety of sources, including published reports, the 1985 National Acid Precipitation and Assessment Program (NAPAP) emissions inventory, and other EPA databases. The U.S. approach for estimating emissions of NO_x, CO, and NMVOCs from stationary combustion as described above is similar to the methodology recommended by the IPCC (IPCC/UNEP/OECD/IEA 1997).

Table C-1: Fuel Consumption by Stationary Combustion for Calculating CH₄ and N₂O Emissions (TBtu)

Fuel/End-Use Sector	1990	1995	1996	1997	1998	1999	2000
Coal	18,169	19,196	20,117	20,660	20,862	20,940	21,722
Residential	62	53	54	58	44	47	47
Commercial	93	80	82	87	66	70	70
Industry	1,478	1,427	1,370	1,381	1,317	1,271	1,119
Electricity Generation	16,536	17,635	18,610	19,135	19,434	19,553	20,487
Petroleum	11,655	11,277	11,941	12,186	12,199	12,376	12,067
Residential	1,266	1,361	1,457	1,432	1,324	1,456	1,492
Commercial	907	715	741	705	665	672	723
Industry	8,152	8,382	8,858	9,056	8,857	8,895	8,657
Electricity Generation	1,329	819	885	993	1,354	1,352	1,195
Natural Gas	18,230	21,087	21,475	21,522	20,871	21,146	22,233
Residential	4,519	4,984	5,390	5,125	4,669	4,858	5,081
Commercial	2,698	3,117	3,250	3,310	3,098	3,130	3,425
Industry	6,963	7,743	8,052	7,966	7,366	7,279	7,250
Electricity Generation	4,050	5,243	4,783	5,122	5,737	5,880	6,477
Wood	1,872	2,044	2,085	1,993	1,998	2,177	2,187
Residential	581	596	595	433	387	414	433
Commercial	37	45	49	47	47	51	52
Industrial	1,254	1,402	1,441	1,513	1,564	1,711	1,702
Electricity Generation	NO	NO	NO	NO	NO	NO	NO

Note: Totals may not sum due to independent rounding.

NO (Not occurring)

Table C-2: CH₄ and N₂O Emission Factors by Fuel Type and Sector (g/GJ)¹

Fuel/End-Use Sector	CH ₄	N ₂ O
Coal		
Residential	300	1.4
Commercial	10	1.4
Industry	10	1.4
Electricity Generation	1	1.4
Petroleum		
Residential	10	0.6
Commercial	10	0.6
Industry	2	0.6
Electricity Generation	3	0.6
Natural Gas		
Residential	5	0.1
Commercial	5	0.1
Industry	5	0.1
Electricity Generation	1	0.1
Wood		

¹ GJ (Gigajoule) = 10⁹ joules. One joule = 9.486×10⁻⁴ Btu

ANNEX D: Methodology for Estimating Emissions of CH₄, N₂O, and Ambient Air Pollutants from Mobile Combustion

Estimates of CH₄ and N₂O Emissions

Greenhouse gas emissions from mobile combustion are reported by transport mode (e.g., road, rail, air, and water), vehicle type, and fuel type. The EPA does not systematically track emissions of CH₄ and N₂O as is done in EPA (2001) for ambient air pollutants; therefore, estimates of these gases were developed using a methodology similar to that outlined in the *Revised 1996 IPCC Guidelines* (IPCC/UNEP/OECD/IEA 1997).

Activity data were obtained from a number of U.S. government agency publications. Depending on the category, these basic activity data included such information as fuel consumption, fuel deliveries, and vehicle miles traveled (VMT).

Highway Vehicles

Step 1: Determine Vehicle Miles Traveled by Vehicle Type, Fuel Type, and Model Year

Vehicle miles traveled (VMT) by vehicle type were obtained from the Federal Highway Administration's (FHWA) *Highway Statistics* (FHWA 1996 through 2001). As these vehicle categories are not fuel-specific, VMT for each vehicle type was disaggregated by fuel type using fuel economy and consumption data, so that the appropriate emission factors could be applied. First, fuel economy and consumption data from FHWA's *Highway Statistics* were disaggregated by fuel type using a number of sources, including the Department of Energy's (DOE) *Transportation Energy Data Book* (DOE 1993 through 2001), FHWA's *Highway Statistics* (FHWA 1996 through 2001), EPA and DOE's *Fuel Economy 2001 Datafile* (EPA, DOE 2001), and the Vehicle Inventory and Use Survey (Census 1997). These data were used to distribute national VMT estimates across vehicle categories.¹

National VMT data for gasoline and diesel highway vehicles are presented in Table D-1 and Table D-2 respectively. Total VMT for each highway category (i.e., gasoline passenger cars, light-duty gasoline trucks, heavy-duty gasoline vehicles, diesel passenger cars, light-duty diesel trucks, heavy-duty diesel vehicles, and motorcycles) were distributed across 25 model years based on the VMT distribution by vehicle age shown in Table D-5. This distribution was derived by weighting the temporally fixed age distribution of the U.S. vehicle fleet according to vehicle registrations (Table D-3) by the average annual age-specific vehicle mileage accumulation of U.S. vehicles (Table D-4). Both were obtained from EPA's Mobile6 model (EPA 2000).

Activity data for gasoline passenger cars and light-duty trucks in California were developed separately due to the different emission control technologies deployed in that state relative to the rest of the country. Unlike the rest of the United States, beginning in model year 1994, a fraction of the computed California VMT for gasoline passenger cars and light-duty trucks was attributed to low emission vehicles (LEVs). LEVs have not yet been widely deployed in other states. The percent of national VMT represented by California for each year was obtained from the Federal Highway Administration (FHWA 1996 through 2001), and applied to national VMT estimates to estimate California VMT for gasoline passenger cars and light-duty trucks (presented in Table D-1).

Step 2: Allocate VMT Data to Control Technology Type

VMT by vehicle type for each model year were distributed across various control technologies as shown in Table D-7, Table D-8, Table D-9, Table D-10, and Table D-11. Again, California gasoline-fueled passenger cars and light-duty trucks were treated separately due to that state's distinct vehicle emission standards—including the introduction

¹ This methodology is presented in more detail in ICF (2001).

of Low Emission Vehicles (LEVs) in 1994—compared with the rest of the United States. The categories “Tier 0” and “Tier 1” were substituted for the early three-way catalyst and advanced three-way catalyst categories, respectively, as defined in the *Revised 1996 IPCC Guidelines*. Tier 0, Tier 1, and LEV are actually U.S. emission regulations, rather than control technologies; however, each does correspond to particular combinations of control technologies and engine design. Tier 1 and its predecessor Tier 0 both apply to vehicles equipped with three-way catalysts. The introduction of “early three-way catalysts,” and “advance three-way catalysts” as described in the *Revised 1996 IPCC Guidelines*, roughly correspond to the introduction of Tier 0 and Tier 1 regulations (EPA 1998).

Step 3: Determine the Amount of CH₄ and N₂O Emitted by Vehicle, Fuel, and Control Technology Type

VMT for each highway category each year as described in Step 1 (see Table D-12) were first converted to vehicle kilometers traveled (VKT) so that IPCC emission factors could be applied. Emissions of CH₄ and N₂O were then calculated by multiplying emission factors in IPCC/UNEP/OECD/IEA (1997) by the IPCC emission factors, which were derived from the EPA’s MOBILE5a mobile source emissions model (EPA 1997). The MOBILE5a model uses information on ambient temperature, diurnal temperature range, altitude, vehicle speeds, national vehicle registration distributions, gasoline volatility, emission control technologies, fuel composition, and the presence or absence of vehicle inspection/maintenance programs in order to produce these factors.

Emissions of N₂O—in contrast to CH₄, CO, NO_x, and NMVOCs—have not been extensively studied and are currently not well characterized. The limited number of studies that have been performed on highway vehicle emissions of N₂O have shown that emissions are generally greater from vehicles with catalytic converter systems than those without such controls, and greater from aged than from new catalysts. These systems control tailpipe emissions of NO_x (i.e., NO and NO₂) by catalytically reducing NO_x to N₂. Suboptimal catalyst performance, caused by as yet poorly understood factors, results in incomplete reduction and the conversion of some NO_x to N₂O rather than to N₂. Fortunately, newer vehicles with catalyst and engine designs meeting the more recent Tier 1 and LEV standards have shown reduced emission rates of both NO_x and N₂O compared with earlier catalyst designs.

In order to better characterize the process by which N₂O is formed by catalytic controls and to develop a more accurate national emission estimate, the EPA’s Office of Transportation and Air Quality—at its National Vehicle and Fuel Emissions Laboratory (NVFEL)—conducted a series of tests in order to measure emission rates of N₂O from used Tier 1 and LEV gasoline-fueled passenger cars and light-duty trucks equipped with catalytic converters. These tests and a review of the literature were used to develop the emission factors for N₂O (EPA 1998). The following references were used in developing the N₂O emission factors for gasoline-fueled highway passenger cars presented in Table D-12:

- *LEVs*. Tests performed at NVFEL (EPA 1998)²
- *Tier 1*. Tests performed at NVFEL (EPA 1998)
- *Tier 0*. Smith and Carey (1982), Barton and Simpson (1994), and one car tested at NVFEL (EPA 1998)
- *Oxidation Catalyst*. Smith and Carey (1982), Urban and Garbe (1980)
- *Non-Catalyst*. Prigent and de Soete (1989), Dasch (1992), and Urban and Garbe (1979)

Nitrous oxide emission factors for other types of gasoline-fueled vehicles—light-duty trucks, heavy-duty vehicles, and motorcycles—were estimated by adjusting the factors for gasoline passenger cars, as described above, by their relative fuel economies. This adjustment was performed using miles per gallon data derived from (DOE 1993 through 2001), (FHWA 1996 through 2001), (EPA, DOE 2001), and (Census 1997) shown in Table D-12. Data from the literature and tests performed at NVFEL support the conclusion that light-duty trucks have higher emission

² LEVs are assumed to be operated using low-sulfur fuel (i.e., Indolene at 24 ppm sulfur). All other NVFEL tests were performed using a standard commercial fuel (CAAB at 285 ppm sulfur). Emission tests by NVFEL have consistently exhibited higher N₂O emission rates from higher sulfur fuels on Tier 1 and LEV vehicles.

rates than passenger cars. However, the use of fuel-consumption ratios to determine emission factors is considered a temporary measure only, to be replaced as soon as real data are available.

The resulting N₂O emission factors employed for gasoline highway vehicles are lower than the U.S. default values presented in the *Revised 1996 IPCC Guidelines*, but are higher than the European default values, both of which were published before the more recent tests and literature review conducted by the NVFEL. The U.S. defaults in the *Guidelines* were based on three studies that tested a total of five cars using European rather than U.S. test procedures.

Nitrous oxide emission factors for diesel highway vehicles were taken from the European default values found in the *Revised 1996 IPCC Guidelines* (IPCC/UNEP/OECD/IEA 1997). Little data addressing N₂O emissions from U.S. diesel-fueled vehicles exists, and in general, European countries have had more experience with diesel-fueled vehicles.

Compared to regulated tailpipe emissions, relatively little data are available to estimate emission factors for N₂O. Nitrous oxide is not a regulated ambient air pollutant, and measurements of it in automobile exhaust have not been routinely collected. Further testing is needed to reduce the uncertainty in nitrous oxide emission factors for all classes of vehicles, using realistic driving regimes, environmental conditions, and fuels.

Non-Highway Vehicles

Activity data for non-highway vehicles were based on annual fuel consumption statistics by transportation mode and fuel type. Consumption data for distillate and residual fuel oil by ships and boats (i.e., vessel bunkering), construction equipment, farm equipment, and locomotives were obtained from EIA (2000b). In the case of ships and boats, the EIA (2000b) vessel bunkering data were reduced by the amount of fuel used for international bunkers.³ Data on the consumption of jet fuel in aircraft were obtained directly from DOT/BTS, as described under CO₂ from Fossil Fuel Combustion, and were reduced by the amount allocated to international bunker fuels. Data on aviation gasoline consumed in aircraft were taken from FAA (2000). Data on the consumption of motor gasoline by ships and boats, construction equipment, farm equipment, and locomotives data were drawn from FHWA (1996 through 2000). The activity data used for non-highway vehicles are included in Table D-6.

Emissions of CH₄ and N₂O from non-highway vehicles were calculated by multiplying U.S. default emission factors in the *Revised 1996 IPCC Guidelines* (IPCC/UNEP/OECD/IEA 1997) by activity data for each vehicle type (see Table D-13).

Table D-14 and Table D-15 provide complete emissions of CH₄ and N₂O emissions, respectively, for 1990 through 2000.

Estimates of NO_x, CO, and NMVOC Emissions

The emission estimates of NO_x, CO, and NMVOCs for mobile combustion were taken directly from the EPA's *National Air Pollutant Emissions Trends, 1900-2000* (EPA 2001). This EPA report provides emission estimates for these gases by sector and fuel type using a "top down" estimating procedure whereby emissions were calculated using basic activity data, such as amount of fuel delivered or miles traveled, as indicators of emissions.

Table D-16 through Table D-18 provide complete emissions estimates for 1990 through 2000.

³ See International Bunker Fuels section of the Energy Chapter.

Table D-1: Vehicle Miles Traveled for Gasoline Highway Vehicles (10⁹ Miles)

Year	Passenger Cars ^a	Light-Duty Trucks ^a	Heavy-Duty Vehicles	Motorcycles	Passenger Cars (CA) ^b	Light-Duty Trucks (CA) ^b
1990	1,227.0	491.1	30.2	8.7	168.5	67.4
1991	1,186.3	556.7	32.1	8.8	159.9	75.0
1992	1,200.9	607.0	31.0	9.1	158.9	80.3
1993	1,205.1	640.6	30.3	9.3	158.2	84.1
1994	1,234.2	657.1	30.6	9.5	160.9	85.7
1995	1,264.3	679.9	30.5	9.8	162.8	87.5
1996	1,295.9	704.3	30.6	9.9	163.2	88.7
1997	1,325.5	733.7	30.6	10.1	166.3	92.1
1998	1,371.8	751.1	31.1	10.3	167.6	91.7
1999	1,385.4	776.9	31.2	10.6	173.9	97.5
2000	1,414.8	796.5	30.2	10.5	177.6	100.0

^a Excludes California^b California VMT for passenger cars and light-duty trucks were disaggregated from national VMT using data from FHWA (1996 through 2001)

Source: Derived from FHWA (1996 through 2001).

Table D-2: Vehicle Miles Traveled for Diesel Highway Vehicles (10⁹ Miles)

Year	Passenger Cars	Light-Duty Trucks	Heavy-Duty Vehicles
1990	13.6	16.0	121.7
1991	12.4	17.6	123.2
1992	12.3	19.5	128.1
1993	12.1	21.1	135.8
1994	11.7	21.8	146.0
1995	11.2	22.6	154.0
1996	10.8	23.5	158.9
1997	10.7	25.0	167.7
1998	10.2	25.4	172.3
1999	9.9	26.6	179.1
2000	9.5	27.5	183.2

Source: Derived from FHWA (1996 through 2001).

Table D-3: Age Distribution by Vehicle/Fuel Type for Highway Vehicles

Vehicle Age	LDGV	LDGT	HDTV	LDDV	LDDT	HDDV	MC
1	5.3%	5.8%	4.9%	5.3%	5.9%	4.2%	14.4%
2	7.1%	7.6%	8.9%	7.1%	7.4%	7.8%	16.8%
3	7.1%	7.5%	8.1%	7.1%	6.9%	7.2%	13.5%
4	7.1%	7.3%	7.4%	7.1%	6.4%	6.7%	10.9%
5	7.0%	7.1%	6.8%	7.0%	6.0%	6.2%	8.8%
6	7.0%	6.8%	6.2%	7.0%	5.6%	5.8%	7.0%
7	6.9%	6.5%	5.6%	6.9%	5.2%	5.3%	5.6%
8	6.8%	6.1%	5.1%	6.8%	4.8%	5.0%	4.5%
9	6.6%	5.7%	4.7%	6.6%	4.5%	4.6%	3.6%
10	6.3%	5.2%	4.3%	6.3%	4.2%	4.3%	2.9%
11	5.9%	4.7%	3.9%	5.9%	3.9%	4.0%	2.3%
12	5.4%	4.2%	3.6%	5.4%	3.6%	3.7%	9.7%
13	4.6%	3.6%	3.3%	4.6%	3.4%	3.4%	0.0%
14	3.6%	3.1%	3.0%	3.6%	3.2%	3.2%	0.0%
15	2.9%	2.6%	2.7%	2.9%	2.9%	2.9%	0.0%
16	2.3%	2.2%	2.5%	2.3%	2.7%	2.7%	0.0%

17	1.8%	1.8%	2.3%	1.8%	2.5%	2.5%	0.0%
18	1.4%	1.4%	2.1%	1.4%	2.4%	2.4%	0.0%
19	1.1%	1.2%	1.9%	1.1%	2.2%	2.2%	0.0%
20	0.9%	1.1%	1.7%	0.9%	2.1%	2.0%	0.0%
21	0.7%	1.1%	1.6%	0.7%	1.9%	1.9%	0.0%
22	0.6%	1.0%	1.5%	0.6%	1.8%	1.8%	0.0%
23	0.4%	1.0%	1.3%	0.4%	1.7%	1.6%	0.0%
24	0.4%	0.9%	1.2%	0.4%	1.6%	1.5%	0.0%
25	1.0%	4.6%	5.4%	1.0%	7.3%	7.2%	0.0%
Total	100.0%	100.0%	100.0%	100.0%	100.0%	100.0%	100.0%

LDGV (gasoline passenger cars, also referred to as light-duty gas vehicles)

LDGT (light-duty gas trucks)

HDGV (heavy-duty gas vehicles)

LDDV (diesel passenger cars, also referred to as light-duty diesel vehicles)

LDDT (light-duty diesel trucks)

HDDV (heavy-duty diesel vehicles)

MC (motorcycles)

Note: Based on vehicle registrations provided by EPA (2000).

Table D-4: Annual Age-specific Vehicle Mileage Accumulation of U.S. Vehicles (Miles)

Vehicle Age	LDGV	LDGT	HDGV	LDDV	LDDT	HDDV	MC
1	14,910	19,906	20,218	14,910	26,371	28,787	4,786
2	14,174	18,707	18,935	14,174	24,137	26,304	4,475
3	13,475	17,559	17,100	13,475	22,095	24,038	4,164
4	12,810	16,462	16,611	12,810	20,228	21,968	3,853
5	12,178	15,413	15,560	12,178	18,521	20,078	3,543
6	11,577	14,411	14,576	11,577	16,960	18,351	3,232
7	11,006	13,454	13,655	11,006	15,533	16,775	2,921
8	10,463	12,541	12,793	10,463	14,227	15,334	2,611
9	9,947	11,671	11,987	9,947	13,032	14,019	2,300
10	9,456	10,843	11,231	9,456	11,939	12,817	1,989
11	8,989	10,055	10,524	8,989	10,939	11,719	1,678
12	8,546	9,306	9,863	8,546	10,024	10,716	1,368
13	8,124	8,597	9,243	8,124	9,186	9,799	1,368
14	7,723	7,925	8,662	7,723	8,420	8,962	1,368
15	7,342	7,290	8,028	7,342	7,718	8,196	1,368
16	6,980	6,690	7,610	6,980	7,075	7,497	1,368
17	6,636	6,127	7,133	6,636	6,487	6,857	1,368
18	6,308	5,598	6,687	6,308	5,948	6,273	1,368
19	5,997	5,103	6,269	5,997	5,454	5,739	1,368
20	5,701	4,642	5,877	5,701	5,002	5,250	1,368
21	5,420	4,214	5,510	5,420	4,588	4,804	1,368
22	5,152	3,818	5,166	5,152	4,209	4,396	1,368
23	4,898	3,455	4,844	4,898	3,861	4,023	1,368
24	4,656	3,123	4,542	4,656	3,542	3,681	1,368
25	4,427	2,822	4,259	4,427	3,250	3,369	1,368

Source: EPA (2000).

Table D-5: VMT Distribution by Vehicle Age and Vehicle/Fuel Type

Vehicle Age	LDGV	LDGT	HDGV	LDDV	LDDT	HDDV	MC
1	7.51%	9.41%	7.89%	7.51%	11.50%	8.27%	19.39%
2	9.52%	11.56%	13.48%	9.52%	13.07%	14.00%	21.15%
3	9.05%	10.62%	11.11%	9.05%	11.15%	11.86%	15.82%
4	8.59%	9.70%	9.85%	8.59%	9.51%	10.05%	11.82%

5	8.14%	8.80%	8.43%	8.14%	8.11%	8.52%	8.77%
6	7.68%	7.92%	7.21%	7.68%	6.92%	7.22%	6.37%
7	7.22%	7.04%	6.16%	7.22%	5.90%	6.13%	4.60%
8	6.72%	6.19%	5.27%	6.72%	5.04%	5.20%	3.31%
9	6.20%	5.36%	4.51%	6.20%	4.30%	4.41%	2.33%
10	5.64%	4.57%	3.86%	5.64%	3.67%	3.74%	1.62%
11	5.03%	3.82%	3.31%	5.03%	3.13%	3.18%	1.09%
12	4.38%	3.14%	2.83%	4.38%	2.67%	2.70%	3.73%
13	3.54%	2.52%	2.42%	3.54%	2.28%	2.29%	0.00%
14	2.67%	1.99%	2.07%	2.67%	1.95%	1.94%	0.00%
15	2.01%	1.54%	1.76%	2.01%	1.66%	1.65%	0.00%
16	1.52%	1.16%	1.52%	1.52%	1.42%	1.40%	0.00%
17	1.14%	0.87%	1.30%	1.14%	1.21%	1.19%	0.00%
18	0.86%	0.64%	1.12%	0.86%	1.04%	1.01%	0.00%
19	0.65%	0.50%	0.96%	0.65%	0.89%	0.86%	0.00%
20	0.49%	0.43%	0.82%	0.49%	0.76%	0.73%	0.00%
21	0.37%	0.37%	0.70%	0.37%	0.65%	0.62%	0.00%
22	0.28%	0.32%	0.60%	0.28%	0.55%	0.53%	0.00%
23	0.21%	0.27%	0.52%	0.21%	0.47%	0.45%	0.00%
24	0.16%	0.23%	0.44%	0.16%	0.40%	0.38%	0.00%
25	0.43%	1.04%	1.85%	0.43%	1.75%	1.65%	0.00%
Total	100.0%	100.0%	100.0%	100.0%	100.0%	100.0%	100.0%

Note: Estimated by weighting data in Table D-3 by data in Table D-4.

Table D-6: Fuel Consumption for Non-Highway Vehicles by Fuel Type (gallons)

Vehicle Type/Year	Residual	Diesel	Jet Fuel	Gasoline ^a
Aircraft				
1990	0	0	18,280,476,364	374,216,115
1991	0	0	17,511,325,335	347,126,395
1992	0	0	17,281,746,858	341,582,453
1993	0	0	17,421,015,955	319,448,684
1994	0	0	18,270,975,739	317,306,704
1995	0	0	17,806,704,239	329,318,581
1996	0	0	18,746,820,369	310,796,773
1997	0	0	18,601,073,081	330,284,570
1998	0	0	19,057,517,441	295,344,794
1999	0	0	19,423,591,168	325,912,623
2000	0	0	20,474,811,494	301,892,666
Ships and Boats				
1990	1,521,437,386	1,697,600,270	0	1,300,400,000
1991	1,486,167,178	1,693,361,391	0	1,709,700,000
1992	2,347,064,583	1,706,143,771	0	1,316,170,000
1993	2,758,924,466	1,546,310,902	0	873,687,000
1994	2,499,868,472	1,630,092,618	0	896,700,000
1995	2,994,692,916	1,518,608,116	0	1,060,394,000
1996	2,286,349,693	1,839,335,006	0	993,671,000
1997	1,011,486,526	1,801,798,270	0	987,193,000
1998	727,907,222	1,597,011,188	0	956,232,000
1999	2,388,334,968	1,855,327,478	0	956,232,001
2000	4,580,188,492	1,889,097,816	0	1,124,269,000
Construction Equip.				
1990	0	1,581,500,000	0	318,200,000
1991	0	1,492,000,000	0	287,200,000
1992	0	1,514,205,000	0	272,900,000

1993	0	1,526,043,000	0	245,299,000
1994	0	1,531,300,000	0	272,852,000
1995	0	1,472,827,000	0	280,046,000
1996	0	1,645,647,000	0	283,911,000
1997	0	1,678,482,000	0	300,491,000
1998	0	1,749,317,000	0	234,705,000
1999	0	1,723,597,000	0	177,758,000
2000	0	1,899,837,000	0	191,516,000
Farm Equipment				
1990	0	3,164,200,000	0	812,800,000
1991	0	3,144,200,000	0	776,200,000
1992	0	3,274,811,000	0	805,500,000
1993	0	3,077,122,000	0	845,320,000
1994	0	3,062,436,000	0	911,996,000
1995	0	3,093,224,000	0	926,732,000
1996	0	3,225,029,000	0	918,085,000
1997	0	3,206,359,000	0	984,450,000
1998	0	2,965,006,000	0	906,941,000
1999	0	2,805,157,000	0	702,700,000
2000	0	3,079,664,000	0	652,256,000
Locomotives				
1990	25,422	3,210,111,000	0	0
1991	6,845	3,026,292,000	0	0
1992	8,343	3,217,231,000	0	0
1993	4,065	2,906,998,000	0	0
1994	5,956	3,063,441,000	0	0
1995	6,498	3,191,023,000	0	0
1996	9,309	3,266,861,000	0	0
1997	3,431	3,067,400,000	0	0
1998	2,587	2,833,276,000	0	0
1999	3,540	2,789,926,000	0	0
2000	7,158	3,070,766,000	0	0
Other^b				
1990	0	926,800,000	0	1,205,400,000
1991	0	955,400,000	0	1,097,700,000
1992	0	773,437,000	0	1,219,300,000
1993	0	797,140,000	0	1,025,087,667
1994	0	905,842,000	0	1,039,309,667
1995	0	800,335,000	0	1,071,596,667
1996	0	741,326,000	0	1,081,639,667
1997	0	706,754,000	0	1,097,257,667
1998	0	682,865,000	0	1,139,228,667
1999	0	685,634,000	0	1,021,835,667
2000	0	610,078,000	0	1,040,137,667

- Not applicable

^a For aircraft, this is aviation gasoline. For all other categories, this is motor gasoline.

^b Other includes snowmobiles and industrial fuel consumption.

Table D-7: Control Technology Assignments for Gasoline Passenger Cars (Percent of VMT)*

Model Years	Non-catalyst	Oxidation	Tier 0	Tier 1
1973-1974	100%	-	-	-
1975	20%	80%	-	-
1976-1977	15%	85%	-	-
1978-1979	10%	90%	-	-

1980	5%	88%	7%	-
1981	-	15%	85%	-
1982	-	14%	86%	-
1983	-	12%	88%	-
1984-1993	-	-	100%	-
1994	-	-	60%	40%
1995	-	-	20%	80%
1996-2000	-	-	-	100%

* Excluding California VMT

- Not applicable

Table D-8: Control Technology Assignments for Gasoline Light-Duty Trucks (Percent of VMT)*

Model Years	Non-catalyst	Oxidation	Tier 0	Tier 1
1973-1974	100%	-	-	-
1975	30%	70%	-	-
1976	20%	80%	-	-
1977-1978	25%	75%	-	-
1979-1980	20%	80%	-	-
1981	-	95%	5%	-
1982	-	90%	10%	-
1983	-	80%	20%	-
1984	-	70%	30%	-
1985	-	60%	40%	-
1986	-	50%	50%	-
1987-1993	-	5%	95%	-
1994	-	-	60%	40%
1995	-	-	20%	80%
1996-2000	-	-	-	100%

* Excluding California VMT

- Not applicable

Table D-9: Control Technology Assignments for California Gasoline Passenger Cars and Light-Duty Trucks (Percent of VMT)

Model Years	Non-catalyst	Oxidation	Tier 0	Tier 1	LEV
1973-1974	100%	-	-	-	-
1975-1979	-	100%	-	-	-
1980-1981	-	15%	85%	-	-
1982	-	14%	86%	-	-
1983	-	12%	88%	-	-
1984-1991	-	-	100%	-	-
1992	-	-	60%	40%	-
1993	-	-	20%	80%	-
1994	-	-	-	90%	10%
1995	-	-	-	85%	15%
1996-2000	-	-	-	80%	20%

* Excluding California VMT

- Not applicable

Table D-10: Control Technology Assignments for Gasoline Heavy-Duty Vehicles (Percent of VMT)

Model Years	Uncontrolled	Non-catalyst	Oxidation	Tier 0
≤1981	100%	-	-	-
1982-1984	95%	-	5%	-
1985-1986	-	95%	5%	-

1987	-	70%	15%	15%
1988-1989	-	60%	25%	15%
1990-2000	-	45%	30%	25%

* Excluding California VMT

- Not applicable

Table D-11: Control Technology Assignments for Diesel Highway and Motorcycle VMT

Vehicle Type/Control Technology	Model Years
Diesel Passenger Cars and Light-Duty Trucks	
Uncontrolled	1966-1982
Moderate control	1983-1995
Advanced control	1996-2000
Heavy-Duty Diesel Vehicles	
Uncontrolled	1966-1972
Moderate control	1983-1995
Advanced control	1996-2000
Motorcycles	
Uncontrolled	1966-1995
Non-catalyst controls	1996-2000

Table D-12: Emission Factors (g/km) for CH₄ and N₂O and Fuel Economy (miles per gallon) for Highway Mobile Combustion

Vehicle Type/Control Technology	N ₂ O	CH ₄	MPG
Gasoline Passenger Cars			21.99
Low Emission Vehicles ^a	0.0176	0.025	
Tier 1 ^b	0.0288	0.030	
Tier 0 ^b	0.0507	0.040	
Oxidation Catalyst	0.0322	0.070	
Non-Catalyst	0.0103	0.120	
Uncontrolled	0.0103	0.135	
Gasoline Light-Duty Trucks			17.56
Low Emission Vehicles ^a	0.0220	0.030	
Tier 1 ^b	0.0361	0.035	
Tier 0 ^b	0.0635	0.070	
Oxidation Catalyst	0.0403	0.090	
Non-Catalyst	0.0129	0.140	
Uncontrolled	0.0129	0.135	
Gasoline Heavy-Duty Vehicles			7.60
Tier 0 ^b	0.1085	0.075	
Oxidation Catalyst ^c	0.0689	0.090	
Non-Catalyst Control	0.0220	0.125	
Uncontrolled	0.0220	0.270	
Diesel Passenger Cars			19.38
Advanced	0.0100	0.01	
Moderate	0.0100	0.01	
Uncontrolled	0.0100	0.01	
Diesel Light Trucks			15.48
Advanced	0.0200	0.01	
Moderate	0.0200	0.01	
Uncontrolled	0.0200	0.01	
Diesel Heavy-Duty Vehicles			5.66
Advanced	0.0300	0.04	
Moderate	0.0300	0.05	
Uncontrolled	0.0300	0.06	

Motorcycles		50.00
Non-Catalyst Control	0.0044	0.13
Uncontrolled	0.0044	0.26

^a Applied to California VMT only.

^b The categories “Tier 0” and “Tier 1” were substituted for the early three-way catalyst and advanced three-way catalyst categories, respectively, as defined in the Revised 1996 IPCC Guidelines.

^c Methane emission factor assumed based on light-duty trucks oxidation catalyst value.

Table D-13: Emission Factors for CH₄ and N₂O Emissions from Non-Highway Mobile Combustion (g gas/kg fuel)

Vehicle Type/Fuel Type	N ₂ O	CH ₄
Ships and Boats		
Residual	0.08	0.23
Distillate	0.08	0.23
Gasoline	0.08	0.23
Locomotives		
Residual	0.08	0.25
Diesel	0.08	0.25
Coal	0.08	0.25
Farm Equipment		
Gas/Tractor	0.08	0.45
Other Gas	0.08	0.45
Diesel/Tractor	0.08	0.45
Other Diesel	0.08	0.45
Construction		
Gas Construction	0.08	0.18
Diesel Construction	0.08	0.18
Other Non-Highway		
Gas Snowmobile	0.08	0.18
Gas Small Utility	0.08	0.18
Gas HD Utility	0.08	0.18
Diesel HD Utility	0.08	0.18
Aircraft		
Jet Fuel	0.1	0.087
Aviation Gasoline	0.04	2.64

Table D-14: CH₄ Emissions from Mobile Combustion (Tg CO₂ Eq.)

Fuel Type/Vehicle Type	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000
Gasoline Highway	4.2	4.2	4.2	4.2	4.1	4.1	3.9	3.8	3.8	3.7	3.6
Passenger Cars	2.4	2.2	2.1	2.1	2.0	2.0	2.0	1.9	1.9	1.9	1.9
Light-Duty Trucks	1.6	1.7	1.8	1.9	1.9	1.8	1.7	1.7	1.6	1.6	1.5
Heavy-Duty Vehicles	0.2	0.2	0.2	0.2	0.1	0.1	0.1	0.1	0.1	0.1	0.1
Motorcycles	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1
Diesel Highway	0.2	0.2	0.2	0.2	0.3	0.3	0.3	0.3	0.3	0.3	0.3
Passenger Cars	+	+	+	+	+	+	+	+	+	+	+
Light-Duty Trucks	+	+	+	+	+	+	+	+	+	+	+
Heavy-Duty Vehicles	0.2	0.2	0.2	0.2	0.3	0.3	0.3	0.3	0.3	0.3	0.3
Non-Highway	0.4	0.4	0.4	0.4	0.4	0.4	0.4	0.4	0.4	0.4	0.5
Ships and Boats	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1
Locomotives	0.1	0.1	0.1	+	0.1	0.1	0.1	0.1	+	+	0.1
Farm Equipment	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1
Construction Equipment	+	+	+	+	+	+	+	+	+	+	+
Aircraft	0.2	0.1	0.1	0.1	0.1	0.1	0.1	0.2	0.1	0.2	0.2
Other*	+	+	+	+	+	+	+	+	+	+	+
Total	4.9	4.9	4.9	4.9	4.8	4.8	4.7	4.6	4.5	4.4	4.4

ANNEX E: Methodology for Estimating CH₄ Emissions from Coal Mining

The methodology for estimating methane emissions from coal mining consists of two distinct steps. The first step addresses emissions from underground mines. For these mines, emissions are estimated on a mine-by-mine basis and then are summed to determine total emissions. The second step of the analysis involves estimating methane emissions for surface mines and post-mining activities. In contrast to the methodology for underground mines, which uses mine-specific data, the surface mine and post-mining activities analysis consists of multiplying basin-specific coal production by basin-specific emission factors.

Step 1: Estimate Methane Liberated and Methane Emitted from Underground Mines

Underground mines generate methane from ventilation systems and from degasification systems. Some mines recover and use methane generated from degasification systems, thereby reducing emissions to the atmosphere. Total methane emitted from underground mines equals the methane liberated from ventilation systems, plus the methane liberated from degasification systems, minus methane recovered and used.

Step 1.1: Estimate Methane Liberated from Ventilation Systems

All coal mines with detectable methane emissions¹ use ventilation systems to ensure that methane levels remain within safe concentrations. Many coal mines do not have detectable levels of methane, while others emit several million cubic feet per day (MMCFD) from their ventilation systems. On a quarterly basis, the U.S. Mine Safety and Health Administration (MSHA) measures methane emissions levels at underground mines. MSHA maintains a database of measurement data from all underground mines with detectable levels of methane in their ventilation air. Based on the four quarterly measurements, MSHA estimates average daily methane liberated at each of the underground mines with detectable emissions.

For the years 1990 through 1996 and 1998 through 2000, MSHA emissions data were obtained for a large but incomplete subset of all mines with detectable emissions. This subset includes mines emitting at least 0.1 MMCFD for some years and at least 0.5 MMCFD for other years, as shown in Table E-1. Well over 90 percent of all ventilation emissions were concentrated in these subsets. For 1997, the complete MSHA database for all 586 mines with detectable methane emissions was obtained. These mines were assumed to account for 100 percent of methane liberated from underground mines. Using the complete database from 1997, the proportion of total emissions accounted for by mines emitting less than 0.1 MMCFD or 0.5 MMCFD was estimated (see Table E-1). The proportion was then applied to the years 1990 through 2000 to account for the less than 10 percent of ventilation emissions coming from mines without MSHA data.

For 1990-1999, average daily methane emissions were multiplied by 365 to determine the annual emissions for each mine. For 2000, MSHA provided quarterly emissions. The average daily methane emissions were multiplied by the number of days corresponding to the number of quarters the mine vent was operating. For example, if the mine vent was operational in one out of the four quarters, the average daily methane emissions were multiplied by 92 days. Total ventilation emissions for a particular year were estimated by summing emissions from individual mines.

Table E-1: Mine-Specific Data Used to Estimate Ventilation Emissions

Year	Individual Mine Data Used
1990	All Mines Emitting at Least 0.1 MMCFD (Assumed to Account for 97.8% of Total)*
1991	1990 Emissions Factors Used Instead of Mine-Specific Data
1992	1990 Emissions Factors Used Instead of Mine-Specific Data

¹ MSHA records coal mine methane readings with concentrations of greater than 50 ppm (parts per million) methane. Readings below this threshold are considered non-detectable.

1993	All Mines Emitting at Least 0.1 MMCFD (Assumed to Account for 97.8% of Total)*
1994	All Mines Emitting at Least 0.1 MMCFD (Assumed to Account for 97.8% of Total)*
1995	All Mines Emitting at Least 0.5 MMCFD (Assumed to Account for 94.1% of Total)*
1996	All Mines Emitting at Least 0.5 MMCFD (Assumed to Account for 94.1% of Total)*
1997	All Mines with Detectable Emissions (Assumed to Account for 100% of Total)
1998	All Mines Emitting at Least 0.1 MMCFD (Assumed to Account for 97.8% of Total)*
1999	All Mines Emitting at Least 0.1 MMCFD (Assumed to Account for 97.8% of Total)*
2000	All Mines Emitting at Least 0.1 MMCFD (Assumed to Account for 97.8% of Total)*

* Factor derived from a complete set of individual mine data collected for 1997.

Step 1.2: Estimate Methane Liberated from Degasification Systems

Coal mines use several different types of degasification systems to remove methane, including vertical wells and horizontal boreholes to recover methane prior to mining of the coal seam. Gob wells and cross-measure boreholes recover methane from the overburden (i.e., GOB area) after mining of the seam (primarily in longwall mines).

MSHA collects information about the presence and type of degasification systems in some mines, but does not collect quantitative data on the amount of methane liberated. Thus, the methodology estimated degasification emissions on a mine-by-mine basis based on other sources of available data. Many of the coal mines employing degasification systems have provided EPA with information regarding methane liberated from their degasification systems. For these mines, this reported information was used as the estimate. In other cases in which mines sell methane recovered from degasification systems to a pipeline, gas sales were used to estimate methane liberated from degasification systems (see Step 1.3). Finally, for those mines that do not sell methane to a pipeline and have not provided information to EPA, methane liberated from degasification systems was estimated based on the type of system employed. For example, for coal mines employing gob wells and horizontal boreholes, the methodology assumes that degasification emissions account for 40 percent of total methane liberated from the mine.

Step 1.3: Estimate Methane Recovered from Degasification Systems and Used (Emissions Avoided)

In 2000, ten active coal mines had methane recovery and use projects and sold the recovered methane to a pipeline. One coal mine also used some recovered methane in a thermal dryer in addition to selling gas to a pipeline. In order to calculate emissions avoided from pipeline sales, information was needed regarding the amount of gas recovered and the number of years in advance of mining that wells were drilled. Several state agencies provided gas sales data, which were used to estimate emissions avoided for these projects. Additionally, coal mine operators provided information on gas sales and/or the number of years in advance of mining. Emissions avoided were attributed to the year in which the coal seam was mined. For example, if a coal mine recovered and sold methane using a vertical well drilled five years in advance of mining, the emissions avoided associated with those gas sales (cumulative production) were attributed to the well up to the time it was mined through (e.g., five years of gas production). Where individual well data is not available, estimated percentages of the operator's annual gas sales within the field around the coal mine are attributed to emissions avoidance. For some mines, individual well data were used to assign gas sales to the appropriate emissions avoided year. In most cases, coal mine operators provided this information, which was then used to estimate emissions avoided for a particular year. Additionally, several state agencies provided production data for individual wells.

Step 2: Estimate Methane Emitted from Surface Mines and Post-Mining Activities

Mine-specific data were not available for estimating methane emissions from surface coal mines or for post-mining activities. For surface mines and post-mining activities, basin-specific coal production was multiplied by a basin-specific emission factor to determine methane emissions.

Step 2.1: Define the Geographic Resolution of the Analysis and Collect Coal Production Data

The first step in estimating methane emissions from surface mining and post-mining activities was to define the geographic resolution of the analysis and to collect coal production data at that level of resolution. The analysis was conducted by coal basin as defined in Table E-2, which presents coal basin definitions by basin and by state.

The Energy Information Agency's (EIA) Coal Industry Annual reports state- and county-specific underground and surface coal production by year. To calculate production by basin, the state level data were grouped into coal basins using the basin definitions listed in Table E-2. For two states—West Virginia and Kentucky—county-level production data was used for the basin assignments because coal production occurred from geologically distinct coal basins within these states. Table E-3 presents the coal production data aggregated by basin.

Step 2.2: Estimate Emissions Factors for Each Emissions Type

Emission factors for surface mined coal were developed from the *in situ* methane content of the surface coal in each basin. Based on an analysis presented in EPA (1993), surface mining emission factors were estimated to be from 1 to 3 times the average *in situ* methane content in the basin. For this analysis, the surface mining emission factor was determined to be twice the *in situ* methane content in the basin. Furthermore, the post-mining emission factors used were estimated to be 25 to 40 percent of the average *in situ* methane content in the basin. For this analysis, the post-mining emission factor was determined to be 32.5 percent of the *in situ* methane content in the basin. Table E-4 presents the average *in situ* content for each basin, along with the resulting emission factor estimates.

Step 2.3: Estimate Methane Emitted

The total amount of methane emitted was calculated by multiplying the coal production in each basin by the appropriate emission factors.

Total annual methane emissions is equal to the sum of underground mine emissions plus surface mine emissions plus post-mining emissions. Table E-5 and Table E-6 present estimates of methane liberated, used, and emitted for 1990 through 2000. Table E-7 provides emissions by state.

Table E-2: Coal Basin Definitions by Basin and by State

Basin	States
Northern Appalachian Basin	Maryland, Ohio, Pennsylvania, West Virginia North
Central Appalachian Basin	Kentucky East, Tennessee, Virginia, West Virginia South
Warrior Basin	Alabama, Mississippi
Illinois Basin	Illinois, Indiana, Kentucky West
South West and Rockies Basin	Arizona, California, Colorado, New Mexico, Utah
North Great Plains Basin	Montana, North Dakota, Wyoming
West Interior Basin	Arkansas, Iowa, Kansas, Louisiana, Missouri, Oklahoma, Texas
Northwest Basin	Alaska, Washington
State	Basin
Alabama	Warrior Basin
Alaska	Northwest Basin
Arizona	South West and Rockies Basin
Arkansas	West Interior Basin
California	South West and Rockies Basin
Colorado	South West and Rockies Basin
Illinois	Illinois Basin
Indiana	Illinois Basin
Iowa	West Interior Basin
Kansas	West Interior Basin
Kentucky East	Central Appalachian Basin
Kentucky West	Illinois Basin
Louisiana	West Interior Basin
Maryland	Northern Appalachian Basin

Basin	Surface Average <i>in situ</i> Content	Underground Average <i>In situ</i> Content	Surface Mine Factors	Post-Mining Surface Factors	Post Mining Underground
Northern Appalachia	49.3	171.7	98.6	16.0	55.8
Central Appalachia	49.3	330.7	98.6	16.0	107.5
Warrior	49.3	318.0	98.6	16.0	103.4
Illinois	39.0	57.20	78.0	12.7	18.6
S. West/Rockies	15.3	225.8	30.6	5.0	73.4
N. Great Plains	3.2	41.67	6.4	1.0	13.5
West Interior	3.2	41.67	6.4	1.0	13.5
Northwest	3.2	41.67	6.4	1.0	13.5

Source: EPA (1993), Anthropogenic Methane Emissions in the United States: Estimates for 1990, Report to Congress, U.S. Environmental Protection Agency, Air and Radiation, April.

Table E-5: Underground Coal Mining Methane Emissions (Billion Cubic Feet)

Activity	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000
Ventilation Output	112	NA	NA	95	96	102	90	96	94	92	87
Adjustment Factor for Mine Data *	97.8%	NA	NA	97.8%	97.8%	91.4%	91.4%	100%	97.8%	97.8%	97.8%
Adjusted Ventilation Output	114	NA	NA	97	98	111	99	96	96	94	89
Degasification System Liberated	54	NA	NA	45	46	46	50	42	49	41	45
Total Underground Liberated	167	164	162	142	144	157	149	138	146	135	134
Recovered & Used	(14)	(15)	(17)	(23)	(27)	(30)	(36)	(28)	(35)	(32)	(36)
Total	154	149	144	119	117	127	113	110	110	103	98

* Refer to Table E-1

Note: Totals may not sum due to independent rounding.

Table E-6: Total Coal Mining Methane Emissions (Billion Cubic Feet)

Activity	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000
Underground Mining	154	149	144	119	117	127	113	110	110	103	98
Surface Mining	25	23	23	23	24	22	23	24	23	22	22
Post-Mining (Underground)	33	31	30	27	30	30	31	32	31	29	28
Post-Mining (Surface)	4	4	4	4	4	4	4	4	4	4	4
Total	216	209	201	173	175	183	172	170	163	161	150

Note: Totals may not sum due to independent rounding.

Table E-7: Total Coal Mining Methane Emissions by State (Million Cubic Feet)

State	1990	1993	1994	1995	1996	1997	1998	1999	2000
Alabama	33,175	26,694	30,283	39,334	29,928	26,440	27,058	26,209	23,997
Alaska	13	12	12	13	11	11	10	12	12
Arizona	402	433	464	425	371	417	403	419	466
Arkansas	0	0	0	0	0	0	0	0	0
California	2	-	-	-	-	-	0	0	0
Colorado	10,117	7,038	9,029	8,541	5,795	9,057	9,057	9,296	10,677
Illinois	10,643	8,737	10,624	11,106	10,890	8,571	7,859	7,812	8,531
Indiana	3,149	2,623	2,791	2,106	2,480	3,088	3,239	2,980	2,492
Iowa	3	1	0	-	-	-	0	0	0
Kansas	5	3	2	2	2	3	3	3	1
Kentucky	21,229	19,823	21,037	19,103	18,292	20,089	19,240	18,255	16,910
Louisiana	24	23	26	28	24	26	24	22	27
Maryland	510	245	256	259	287	296	282	260	345
Mississippi	-	-	-	-	-	-	0	2	92
Missouri	20	5	6	4	5	3	3	3	3
Montana	280	267	310	294	283	305	319	306	285
New Mexico	905	1,186	1,223	980	856	961	1,026	1,042	972
North Dakota	217	238	240	224	222	220	223	232	233
Ohio	4,710	4,110	4,377	3,900	3,992	4,313	4,244	3,820	3,443
Oklahoma	13	14	52	14	14	132	137	209	208
Pennsylvania	22,573	26,437	24,026	27,086	26,567	30,339	29,853	24,088	25,352
Tennessee	800	350	338	366	418	390	309	349	306
Texas	415	406	389	392	410	397	391	395	361
Utah	4,562	4,512	3,696	3,541	4,061	4,807	5,060	4,851	4,045
Virginia	45,883	30,457	26,765	19,893	19,847	16,972	14,087	13,539	12,179
Washington	37	35	36	36	34	33	35	31	32
West Virginia	55,280	37,803	36,854	42,992	42,870	40,197	43,511	41,500	37,490
Wyoming	1,382	1,578	1,782	1,977	2,090	2,122	2,351	2,520	2,533
Total	216,350	173,029	174,622	182,616	169,750	169,190	168,725	158,153	151,397

+ Does not exceed 0.5 Million Cubic Feet

Note: The emission estimates provided above are inclusive of emissions from underground mines, surface mines and post-mining activities. The following states have neither underground nor surface mining and thus report no emissions as a result of coal mining: Connecticut, Delaware, Florida, Georgia, Hawaii, Idaho, Maine, Massachusetts, Michigan, Minnesota, Nebraska, Nevada, New Hampshire, New Jersey, New York, North Carolina, Oregon, Rhode Island, South Carolina, South Dakota, Vermont, and Wisconsin. Emission estimates are not given for 1991 and 1992 because underground mine data was not available for those years.

ANNEX F: Methodology for Estimating CH₄ Emissions from Natural Gas Systems

The following steps were used to estimate methane emissions from natural gas systems.

Step 1: Calculate Emission Estimates for Base Year 1992 Using GRI/EPA Study

The first step in estimating methane emissions from natural gas systems was to develop a detailed base year estimate of emissions. The study by GRI/EPA (1996) divides the industry into four stages to construct a detailed emission inventory for the year 1992. These stages include: field production, processing, transmission and storage (i.e., both underground and liquefied gas storage), and distribution. This study produced emission factors and activity data for over 100 different emission sources within the natural gas system. Emissions for 1992 were estimated by multiplying activity levels by emission factors for each system component and then summing by stage. Since publication, the EPA has updated 1992 activity data for some of the components in the system. Table F-2 displays the 1992 GRI/EPA activity levels and emission factors for venting and flaring from the field production stage, and the current EPA activity levels and emission factors. These data are shown to illustrate the kind of data used to calculate emissions from all stages.

Step 2: Collect Aggregate Statistics on Main Driver Variables

As detailed data on each of the over 100 sources were not available for the period 1990 through 1999, activity levels were estimated using aggregate statistics on key drivers, including: number of producing wells (IPAA 1990 through 2000, EIA 2000), number of gas plants (AGA 1990 through 1998; PennWell Corporation 1999, 2000, 2001), miles of transmission pipeline (OPS 1998, 1999, 2000, 2001), miles of distribution pipeline (OPS, 2001), miles of distribution services (OPS, 2001), and energy consumption (EIA 1998, 1999, 2000, 2001). Data on the distribution of gas mains and services by material type was not available for 1990 through 1992 from OPS. For those years, the distribution by type was back calculated from 1993 using compound growth rates determined for the years 1993 through 2000. Table F-3 provides the activity levels of some of the key drivers in the natural gas analysis.

Step 3: Estimate Emission Factor Changes Over Time

For the period 1990 through 1995, the emission factors were held constant, based on 1992 values. However, the natural gas industry is experiencing ongoing broad based technology improvements, which are expected to continue. These improvements have increased operating efficiency, thus reducing methane emissions. An assumed improvement in technology and practices was estimated to reduce emission factors by 5 percent by the year 2020. This assumption, annualized, amounts to a 0.2 percent decline in the 1996 emission factor, a 0.4 percent decline in the 1997 emission factor, a 0.6 percent decline in the 1998 emission factor and a 0.8 percent decline in the 1999 emission factor, all relative to 1995 emission factors.

Step 4: Estimate Emissions for Each Year and Stage

Emissions from each stage of the natural gas industry were estimated by multiplying the activity factors by the appropriate emission factors, summing all sources for each stage, and then subtracting the Natural Gas STAR emission reductions.¹ Methane reductions from the Natural Gas STAR program for the years 1994 through 2000 are presented in Table F- 1. Emission reductions by project are reported by industry partners using actual measurement data or equipment-specific emission factors. Before incorporating the reductions into the Inventory, quality assurance and quality control checks are undertaken to identify errors, inconsistencies, or irregular data. Total emissions were estimated by adding the emission estimates from each stage. Table F-4 illustrates emission estimates for venting and flaring emissions from the field production stage using this methodology.

¹ It is assumed that the 5 percent decline in the emissions factor from 1995 to 2020 does not reflect emission reductions attributed to Natural Gas STAR. The emission factor decline accounts for regular technology improvements only. This assumption is being investigated for future inventories.

Table F- 1. Methane reductions from the Natural Gas STAR program (Tg)

Process	1994	1995	1996	1997	1998	1999	2000
Production	0.10	0.12	0.19	0.25	0.29	0.38	0.48
Processing	-	-	-	-	-	-	-
Transmission and Storage	0.07	0.09	0.08	0.11	0.15	0.18	0.23
Distribution	0.02	0.02	0.01	0.03	0.02	0.02	0.02

Table F-2: 1992 Data and Emissions (Mg) for Venting and Flaring from Natural Gas Field Production Stage

Activity	GRI/EPA Values			EPA Adjusted Values		
	Activity Data	Emission Factor	Emissions	Activity Data	Emission Factor	Emissions
Drilling and Well Completion						
Completion Flaring	844 compl/yr	733 Scf/comp	12	400 compl/yr	733 scf/comp	6
Normal Operations						
Pneumatic Device Vents	249,111 controllers	345 Scfd/device	602,291	249,111 controllers	345 scfd/device	602,291
Chemical Injection Pumps	16,971 active pumps	248 Scfd/pump	29,501	16,971 active pumps	248 scfd/pump	29,502
Kimray Pumps	11,050,000 MMscf/yr	368 Scf/MMscf	78,024	7,380,194 MMscf/yr	992 scf/MMscf	140,566
Dehydrator Vents	12,400,000 MMscf/yr	276 Scf/MMscf	65,608	8,200,215 MMscf/yr	276 scf/MMscf	43,387
Compressor Exhaust Vented						
Gas Engines	27,460 MMHPHr	0.24 Scf/HPHr	126,536	27,460 MMHPHr	0.24 scf/HPHr	126,535
Routine Maintenance						
Well Workovers						
Gas Wells	9,392 w.o./yr	2,454 scfy/w.o.	443	9,392 w.o./yr	2,454 scfy/w.o.	443
Well Clean Ups (LP Gas Wells)	114,139 LP gas wells	49,570 scfy/LP well	108,631	114,139 LP gas wells	49,570 scfy/LP well	108,631
Blowdowns						
Vessel BD	255,996 vessels	78 scfy/vessel	383	242,306 vessels	78 scfy/vessel	363
Pipeline BD	340,000 miles (gath)	309 scfy/mile	2,017	340,200 miles (gath)	309 scfy/mile	2,018
Compressor BD	17,112 compressors	3,774 scfy/comp	1,240	17,112 compressors	3,774 scfy/comp	1,240
Compressor Starts	17,112 compressors	8,443 scfy/comp	2,774	17,112 compressors	8,443 scfy/comp	2,774
Upsets						
Pressure Relief Valves	529,440 PRV	34.0 scfy/PRV	346	529,440 PRV	34.0 scfy/PRV	346
ESD	1,115 platforms	256,888 scfy/plat	5,499	1,372 platforms	256,888 scfy/plat	6,767
Mishaps	340,000 miles	669 scfy/mile	4,367	340,200 miles	669 scfy/mile	4,370

Table F-3: Activity Factors for Key Drivers

Variable	Units	1990	1995	1996	1997	1998	1999	2000
Transmission Pipelines Length	miles	291,990	296,947	292,186	294,304	302,706	296,581	293,774
Wells								
GSAM Appalachia Wells*	# wells	120,162	123,092	122,700	120,037	117,878	118,723	118,723
GSAM N Central Associated Wells*	# wells	3780	3641	3417	3409	3,361	3,361	3,361
GSAM N Central Non-Associated Wells*	# wells	3,105	6,323	7,073	6,701	8,664	8,548	8,548
GSAM Rest of US Wells*	# wells	145,100	164,750	173,928	173,550	190,387	180,178	180,178
GSAM Rest of US Associated Wells*	# wells	277,301	264,837	264,807	264,185	254,848	254,848	254,848
Appalach. + N. Central Non-Assoc. + Rest of US Platforms	# wells	268,367	294,165	303,701	300,288	316,929	307,449	307,449
Gulf of Mexico Off-shore Platforms	# platforms	3,798	3,868	3,846	3,846	3,963	3,975	4019
Rest of U.S. (offshore platforms)	# platforms	24	23	24	23	23	23	23
N. Central Non-Assoc. + Rest of US Wells	# platforms	148,205	171,073	181,001	180,251	199,051	188,726	188,726
Gas Plants								
Number of Gas Plants	# gas plants	761	675	623	615	558	581	585
Distribution Services								
Steel – Unprotected	# of services	5,500,993	6,151,653	5,775,613	5,518,795	5,463,253	5,751,250	5,676,582
Steel - Protected	# of services	19,916,202	21,002,455	18,593,770	19,078,467	18,478,344	18,310,719	17,775,878
Plastic	# of services	16,269,414	26,044,545	26,187,530	27,800,407	28,629,384	28,796,959	31,644,018
Copper	# of services	1,379,237	1,445,380	1,519,625	1,498,050	1,464,019	1,458,518	1,434,091
Total	# of services	43,065,846	54,644,033	52,076,544	53,895,713	54,035,000	54,317,433	56,530,565
Distribution Mains								
Steel – Unprotected	miles	91,267	94,058	88,412	85,166	86,639	84,534	82,817
Steel – Protected	miles	491,120	503,288	484,526	479,278	484,963	459,298	468,932
Cast Iron	miles	52,644	50,625	51,542	47,669	47,587	45,865	44,736
Plastic	miles	202,269	353,735	350,699	385,373	400,627	415,210	446,554
Total	miles	837,300	1,001,706	975,179	997,486	1,019,816	1,004,907	1,043,039

* GSAM (Gas Systems Analysis Model) is a natural gas supply, demand, and transportation model used by the Federal Energy Technology Center of the U.S. Department of Energy (GSAM 1997).

Table F-4: CH₄ Emission Estimates for Venting and Flaring from the Field Production Stage (Mg)

Activity	1990	1995	1996	1997	1998	1999	2000
Drilling and Well Completion							
Completion Flaring	5.4	5.9	6.1	6.0	6.3	6.1	6.1
Normal Operations							
Pneumatic Device Vents	567,778	655,386	692,033	687,785	757,995	717,231	715,785
Chemical Injection Pumps	36,449	43,111	45,666	45,256	50,352	47,769	47,673
Kimray Pumps	134,247	147,191	151,572	149,492	157,342	152,492	152,184
Dehydrator Vents	41,436	45,432	46,784	46,142	48,565	47,068	46,973
Compressor Exhaust Vented Gas Engines	119,284	137,690	145,389	144,497	159,247	150,683	150,379
Routine Maintenance							
Well Workovers Gas Wells	531	582	600	591	622	603	602
Well Clean Ups (LP Gas Wells)	101,118	110,868	114,168	112,601	118,514	114,861	114,629
Blowdowns							
Vessel BD	256	292	306	303	329	314	314
Pipeline BD	1,710	1,852	1,908	1,894	1,997	1,932	1,928
Compressor BD	1,548	1,730	1,802	1,786	1,920	1,837	1,833
Compressor Starts	3,462	3,871	4,031	3,995	4,296	4,110	4,102
Upsets							
Pressure Relief Valves	326	376	397	395	435	412	411
ESD	6,764	6,882	6,834	6,816	7,006	7,013	7,075
Mishaps	925	1,003	1,033	1,025	1,081	1,045	1,043

Table F-5: U.S. Total Natural Gas Production (Trillion Ft³/yr) and Associated CH₄ Emissions (Gg)

Activity	1990	1995	1996	1997	1998	1999	2000
Production	17.8	18.6	18.9	18.9	18.7	18.6	19.1
CH ₄ Emissions from Production	1,404	1,470	1,466	1,397	1,496	1,329	1,226

ANNEX G: Methodology for Estimating CH₄ Emissions from Petroleum Systems

The methodology for estimating methane emissions from petroleum systems is based on the 1999 EPA draft report, *Estimates of Methane Emissions from the U.S. Oil Industry* (EPA 1999) and Radian's Study, *Methane Emissions from the U.S. Petroleum Industry* (Radian 1996). Seventy activities that emit methane from petroleum systems were examined for these reports. Most of the activities analyzed involve crude oil production field operations, which accounted for 97 percent of total oil industry emissions. Crude transportation and refining accounted for the remaining emissions at about one and two percent each, respectively.

The following steps were taken to estimate methane emissions from petroleum systems.

Step 1: Determine Emission Factors for all activities

The emission factors for 1995 are taken from the 1999 EPA draft report, which contains the most recent and comprehensive determination of emission factors for the seventy methane emitting activities in the oil industry. The emission factors determined for 1995 are assumed to be representative of emissions from each source type over the period 1990 through 2000. Therefore, the same emission factors are used for each year throughout this period.

Step 2: Determine Activity Levels for Each Year

Activity levels change from year to year. Some factors change in proportion to crude oil rates: production, transportation, or refinery runs. Some change in proportion to the number of facilities: oil wells or petroleum refineries. Some factors change proportional to both rate and number of facilities.

For fifty-seven of the seventy activities, activity levels for 1995 are taken from EPA. For the remaining thirteen activities, the activity level for 1993 is taken from Radian (1996). These thirteen activity levels were derived from field data collected in 1993, along with 1993 crude oil production and number of wells.

For both sets of data, a determination was made on a case-by-case basis, as to which measure of petroleum industry activity best reflects the change in annual activity relative to the base years (1993 and 1995). Publicly reported data from the Minerals Management Service (MMS), Energy Information Administration (EIA), American Petroleum Institute (API) and the Oil & Gas Journal (O&GJ) were used to extrapolate the activity levels from the base year to each year between 1990 to 2000. Data used include total domestic crude oil production, number of domestic crude oil wells, total imports and exports of crude oil, and total petroleum refinery crude runs. For a small number of sources, 2000 data were not yet available. In these cases, the 2000 activity factors were used. In the few cases where no data was located, activity data based on oil industry expert judgment were used.

Step 3: Estimate Methane Emissions for Each Activity for Each Year

Annual emissions from each of the 70 petroleum system activities analyzed were estimated by multiplying the activity data for each year by the corresponding emission factor. These annual emissions for each activity are then summed to estimate the total annual methane emissions. Table G-1, Table G-2, and Table G-3 provide the 2000 activity factors, emission factors, and emission estimates. Table G-4 provides a summary of emission estimates for the years 1990 through 2000.

Table G-1: 2000 CH₄ Emissions from Petroleum Production Field Operations

Activity/Equipment	Emission		Activity		Emissions (Bcf/yr)
	Factor	Units	Factor	Units	
Vented Emissions					
Oil Tanks	18 scf of CH ₄ /bbl crude		1,234	MMbbl/yr (non stripper wells)	47.754
Pneumatic Devices, High Bleed	345 scfd CH ₄ /device		142,840	No. of high-bleed devices	22.086
Pneumatic Devices, Low Bleed	35 scfd CH ₄ /device		265,275	No. of low-bleed devices	17.998
Chemical Injection Pumps	248 scfd CH ₄ /pump		28,595	No. of pumps	3.342
Vessel Blowdowns	78 scfy CH ₄ /vessel		186,502	No. of vessels	2.589
Compressor Blowdowns	3,775 scf/yr of CH ₄ /compressor		2,531	No. of compressors	0.010
Compressor Starts	8,443 scf/yr. of CH ₄ /compressor		2,531	No. of compressors	0.021
Stripper wells	2,345 scf/yr of CH ₄ /stripper well		348,867	No. of stripper wells vented	0.818
Well Completion Venting	733 scf/completion		4,731	Oil well completions	0.003
Well Workovers	96 scf CH ₄ /workover		40,050	Oil well workovers	0.004
Pipeline Pigging	2.40 scfd of CH ₄ /pig station		0	No. of crude pig stations	0.000
Offshore Platforms, Gulf of Mexico	1,283 scfd CH ₄ /platform		1,832	No. of oil platforms	0.858
Offshore Platforms, Other U.S. Areas	1,283 scfd CH ₄ /platform		23	No. of oil platforms	0.011
Fugitive Emissions					
Offshore Platforms, Gulf of Mexico	56 scfd CH ₄ /platform		1,832	No. of oil platforms	0.037
Offshore Platforms, Other U.S. Areas	56 scfd CH ₄ /platform		23	No. of oil platforms	0.000
Oil Wellheads (heavy crude)	0.13 scfd/well		13,052	No. of hvy. crude wells *	0.001
Oil Wellheads (light crude)	16.6 scfd/well		172,081	No. of lt. crude wells *	1.045
Separators (heavy crude)	0.15 scfd CH ₄ /separator		10,970	No. of hvy. crude seps.	0.001
Separators (light crude)	14 scfd CH ₄ /separator		99,837	No. of lt. crude seps.	0.505
Heater/Treaters (light crude)	19 scfd CH ₄ /heater		75,695	No. of heater treaters	0.530
Headers (heavy crude)	0.08 scfd CH ₄ /header		13,929	No. of hvy. crude hdrs.	0.000
Headers (light crude)	11 scfd CH ₄ /header		43,183	No. of lt. crude hdrs.	0.171
Floating Roof Tanks	338,306 scf CH ₄ /floating roof tank/yr.		24	No. of floating roof tanks	0.008
Compressors	100 scfd CH ₄ /compressor		2,531	No. of compressors	0.092
Large Compressors	16,360 scfd CH ₄ /compressor		0	No. of large comprs.	0.000
Sales Areas	41 scf CH ₄ /loading		1,760,539	Loadings/year	0.071
Pipelines	0 scfd of CH ₄ /mile of pipeline		30,467	Miles of gathering line	0.000
Well Drilling	0 scfd of CH ₄ /oil well drilled		7,437	No. of oil wells drilled	0.000
Battery Pumps	0.24 scfd of CH ₄ /pump		160,200	No. of battery pumps	0.014

Table G-1: 2000 CH₄ Emissions from Petroleum Production Field Operations (Continued)

Combustion Emissions									
Gas Engines	0.08 scf CH ₄ /HP-hr	15,945 MMHP-hr							1.630
Heaters	0.52 scf CH ₄ /bbl	2113.0 MBbl/yr							1.276
Well Drilling	2,453 scf CH ₄ /well drilled	7,437 Oil wells drilled, 1995							0.001
Flares	20 scf CH ₄ /per Mcf flared	485,883 Mcf flared/yr							0.018
Offshore Platforms, Gulf of Mexico	481 scfd CH ₄ /platform	1,832 No. of oil platforms							0.010
Offshore Platforms, Other U.S. Areas	481 scfd CH ₄ /platform	23 No. of oil platforms							0.321
Process Upset Emissions									0.548
Platform Emergency Shutdowns	256,888 scfy/platform	1,855 No. of platforms							0.477
Pressure Relief Valves	35 scf/yr/PR valve	176,498 No. of PR valves							0.006
Well Blowouts Offshore	5.0 MMscf/blowout	2.25 No. of blowouts/yr							0.011
Well Blowouts Onshore	2.5 MMscf/blowout	24.8 No. of blowouts/yr							0.062
Total (excluding stripper wells)									52.42

Table G-2: 2000 CH₄ Emissions from Petroleum Transportation

Activity/Equipment	Emission Factor	Units	Activity Factor	Units	Emissions (Bcf/yr)
Vented Emissions					0.233
Tanks	0.021 scf CH ₄ /yr/bbl of crude delivered to refineries		5,514 MMbbl crude feed/yr		0.114
Truck Loading	0.520 scf CH ₄ /yr/bbl of crude transported by truck		50.6 MMbbl crude feed/yr		0.026
Marine Loading	2.544 scf CH ₄ /1000 gal. crude marine loadings		29,581,881 1,000 gal./yr loaded		0.075
Rail Loading	0.520 scf CH ₄ /yr/bbl of crude transported by rail		7.3 MMbbl. Crude by rail/yr		0.004
Pump Station Maintenance	36.80 scf CH ₄ /station/yr		490 No. of pump stations		0.000
Pipeline Pigging	39 scfd of CH ₄ /pig station		980 No. of pig stations		0.014
Fugitive Emissions					0.050
Pump Stations	25 scfCH ₄ /mile/yr.		48,990 No. of miles of crude p/1		0.001
Pipelines	0 scf CH ₄ /bbl crude transported by pipeline		7,551 MM bbl crude piped		0.000
Floating Roof Tanks	58,965 scf CH ₄ /floating roof tank/yr.		824 No. of floating roof tanks		0.049
Combustion Emissions					0.000
Pump Engine Drivers	0.24 scf CH ₄ /hp-hr		NA No. of hp-hrs		NA
Heaters	0.521 scf CH ₄ /bbl.burned		NA No. of bbl. Burned		NA
Total					0.283

Table G-3: 2000 CH₄ Emissions from Petroleum Refining

Activity/Equipment	Emission		Activity		Emissions (Bcf/yr)
	Factor	Units	Factor	Units	
Vented Emissions					1,260
Tanks	20.6 scfCH ₄ /Mbbbl		1,938 Mbbbl/cd heavy crude feed		0.015
System Blowdowns	137 scfCH ₄ /Mbbbl		15,108 Mbbbl/cd refinery feed		0.754
Asphalt Blowing	2,555 scfCH ₄ /Mbbbl		527 Mbbbl/cd production		0.491
Fugitive Emissions					0.093
Fuel Gas System	439 McfCH ₄ /refinery/yr		158 Refineries		0.069
Floating Roof Tanks	587 scf CH ₄ /floating roof tank/yr.		767 No. of floating roof tanks		0.000
Wastewater Treating	1.88 scfCH ₄ /Mbbbl		15,108 Mbbbl/cd refinery feed		0.010
Cooling Towers	2.36 scfCH ₄ /Mbbbl		15,108 Mbbbl/cd refinery feed		0.013
Combustion Emissions					0.092
Atmospheric Distillation	3.61 scfCH ₄ /Mbbbl		15,108 Mbbbl/cd refinery feed		0.020
Vacuum Distillation	3.61 scfCH ₄ /Mbbbl		6,916 Mbbbl/cd feed		0.009
Thermal Operations	6.02 scfCH ₄ /Mbbbl		2,004 Mbbbl/cd feed		0.004
Catalytic Cracking	5.17 scfCH ₄ /Mbbbl		5,175 Mbbbl/cd feed		0.010
Catalytic Reforming	7.22 scfCH ₄ /Mbbbl		3,296 Mbbbl/cd feed		0.009
Catalytic Hydrocracking	7.22 scfCH ₄ /Mbbbl		1,334 Mbbbl/cd feed		0.004
Hydrorefining	2.17 scfCH ₄ /Mbbbl		1,591 Mbbbl/cd feed		0.001
Hydrotreating	6.50 scfCH ₄ /Mbbbl		8,462 Mbbbl/cd feed		0.020
Alkylation/Polymerization	12.6 scfCH ₄ /Mbbbl		1,147 Mbbbl/cd feed		0.005
Aromatics/Isomeration	1.80 scfCH ₄ /Mbbbl		929 Mbbbl/cd feed		0.001
Lube Oil Processing	0.00 scfCH ₄ /Mbbbl		163 Mbbbl/cd feed		0.000
Engines	0.006 scfCH ₄ /hp-hr		1,467 MMhp-hr/yr		0.008
Flares	0.189 scfCH ₄ /Mbbbl		15,108 Mbbbl/cd refinery feed		0.001
Total					1,445

Table G-4: Summary of CH₄ Emissions from Petroleum Systems (Gg)

Activity	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000
Production Field Operations	1,227	1,245	1,201	1,159	1,139	1,122	1,114	1,112	1,081	1,028	1,008
Tank venting	564	570	548	519	502	493	485	484	466	433	425
Pneumatic device venting	525	535	517	507	504	497	496	495	485	470	460
Wellhead fugitives	25	26	25	24	25	25	25	24	23	21	20
Combustion & process upsets	47	48	46	45	45	44	45	45	44	42	42
Misc. venting & fugitives	66	66	65	64	64	63	63	63	63	62	61
Crude Oil Transportation	7	6	6	6	6	6	6	6	6	6	5
Refining	25	24	24	25	25	25	26	27	27	27	28
Total	1,258	1,276	1,231	1,190	1,170	1,154	1,145	1,144	1,114	1,061	1,041

Note: Totals may not sum due to independent rounding.

ANNEX H: Methodology for Estimating CO₂ Emissions from Municipal Solid Waste Combustion

Emissions of CO₂ from MSW combustion include CO₂ generated by the combustion of plastics, synthetic rubber and synthetic fibers in MSW, and combustion of synthetic rubber and carbon black in tires. Combustion of MSW also results in emissions of N₂O. The methodology for calculating emissions from each of these waste combustion sources is described in this Annex.

CO₂ from Plastics Combustion

In the report, *Characterization of Municipal Solid Waste in the United States* (EPA 2000c), the flows of plastics in the U.S. waste stream are reported for seven resin categories. The 1998 quantity generated, recovered, and discarded for each resin is shown in Table H-1. The EPA report does not provide estimates for individual materials landfilled and combusted, although it does provide such an estimate for the waste stream as a whole. To estimate the quantity of plastics landfilled and combusted, total discards were apportioned based on the proportions of landfilling and combustion for the entire U.S. waste stream in 1998 (76 percent and 24 percent, respectively). Emissions for 1990 through 1997 were calculated using the same approach. Figures for 1999 and 2000 are calculated using 1998 resin ratios and forecasted from 1998 figures for generation and recovery using a 2.85 percent annual growth rate for plastics generation and a 5.4 percent growth rate for the plastics recovery rate, based on reported trends (EPA 2000c).

Table H-1: 1998 Plastics in the Municipal Solid Waste Stream by Resin (Gg)

Waste Pathway	PET	HDPE	LDPE/		PP	PS	Other	Total
			PVC	LLDPE				
Generation	2,023	4,500	1,243	4,844	2,576	1,969	3,139	20,294
Recovery	354	399	0	127	154	18	45	1,098
Discard	1,669	4,101	1,243	4,717	2,422	1,950	3,094	19,196
Landfill	1,269	3,116	945	3,585	1,841	1,482	2,351	14,589
Combustion	401	984	298	1,132	581	468	742	4,607
Recovery*	17%	9%	0%	3%	6%	1%	1%	5%
Discard*	83%	91%	100%	97%	94%	99%	99%	95%
Landfill*	63%	69%	76%	74%	71%	75%	75%	72%
Combustion*	20%	22%	24%	23%	23%	24%	24%	23%

*As a percent of waste generation.

Note: Totals may not sum due to independent rounding. Abbreviations: PET (polyethylene terephthalate), HDPE (high density polyethylene), PVC (polyvinyl chloride), LDPE/LLDPE (linear low density polyethylene), PP (polypropylene), PS (polystyrene).

Fossil fuel-based CO₂ emissions for 1998 were calculated as the product of plastic combusted, carbon content, and fraction oxidized (see Table H-2). The carbon content of each of the six types of plastics is listed, with the value for "other plastics" assumed equal to the weighted average of the six categories. The fraction oxidized was assumed to be 98 percent.

Table H-2: 1998 Plastics Combusted (Gg), Carbon Content (%), and Carbon Combusted (Gg)

Factor	PET	HDPE	LDPE/		PP	PS	Other	Total
			PVC	LLDPE				
Quantity Combusted	401	984	298	1,132	581	468	742	4,607
Carbon Content of Resin	63%	86%	38%	86%	86%	92%	66% ^a	-
Carbon in Resin Combusted	250	844	115	970	498	432	489	3,598
Emissions (Tg CO₂ Eq.)^b	0.9	3.0	0.4	3.5	1.8	1.6	1.8	12.9

^a Weighted average of other plastics produced in 1998 production.

^b Assumes a fraction oxidized of 98 percent.

CO₂ from Combustion of Synthetic Rubber and Carbon Black in Tires

Emissions from tire combustion require two pieces of information: the amount of tires combusted and the carbon content of the tires. The *Scrap Tire Use/Disposal Study 1998/1999 Update* (STMC 1999) reports that 114 million of the 270 million scrap tires generated in 1998 (approximately 42 percent of generation) were used for fuel purposes. Using STMC estimates of average tire composition and weight, the weight of synthetic rubber and carbon black in scrap tires was determined. Synthetic rubber in tires was estimated to be 90 percent carbon by weight, based on the weighted average carbon contents of the major elastomers used in new tire consumption (see Table H-3).¹ Carbon black is 100 percent carbon. Multiplying the proportion of scrap tires combusted by the total carbon content of the synthetic rubber and carbon black portion of scrap tires yielded CO₂ emissions, as shown in Table H-4. Note that the disposal rate of rubber in tires (0.4 Tg/yr) is smaller than the consumption rate for tires based on summing the elastomers listed in Table H-3 (1.3 Tg/yr); this is due to the fact that much of the rubber is lost through tire wear during the product's lifetime and due to the lag time between consumption and disposal of tires. Tire production and fuel use for 1999 and 2000 were extrapolated from trend data for 1994 through 1998.

Table H-3: Elastomers Consumed in 1998 (Gg)

Elastomer	Consumed	Carbon Content	Carbon Equivalent
Styrene butadiene rubber solid	908	91%	828
For Tires	743	91%	677
For Other Products*	165	91%	151
Polybutadiene	561	89%	499
For Tires	404	89%	359
For Other Products	157	89%	140
Ethylene Propylene	320	86%	274
For Tires	10	86%	8
For Other Products	310	86%	266
Polychloroprene	69	59%	40
For Tires	0	59%	0
For Other Products	69	59%	40
Nitrile butadiene rubber solid	87	77%	67
For Tires	1	77%	1
For Other Products	86	77%	67
Polyisoprene	78	88%	69
For Tires	65	88%	57
For Other Products	13	88%	12
Others	369	88%	324
For Tires	63	88%	56
For Other Products	306	88%	268
Total	2,392	-	2,101

*Used to calculate carbon content of non-tire rubber products in municipal solid waste.

- Not applicable

Table H-4: Scrap Tire Constituents and CO₂ Emissions from Scrap Tire Combustion in 1998

Material	Weight of Material (Tg)	Carbon Content	Percent Combusted	Emissions (Tg CO ₂ Eq.)*
Synthetic Rubber	0.4	90%	42%	1.1
Carbon Black	0.5	100%	42%	1.7
Total	0.8	-	-	2.9

¹ The carbon content of tires (1,158,000 Tg) divided by for the mass of tires (1,285,000 Tg) equals 90 percent.

* Assumes a fraction oxidized of 98 percent.
 - Not applicable

CO₂ from Combustion of Synthetic Rubber in Municipal Solid Waste

Similar to the methodology for scrap tires, CO₂ emissions from synthetic rubber in MSW were estimated by multiplying the amount of rubber combusted by an average rubber carbon content. The amount of rubber in the MSW stream was estimated from data provided in the *Characterization of Municipal Solid Waste in the United States* (EPA 2000c). The report divides rubber found in MSW into three product categories: other durables (not including tires), non-durables (which includes clothing and footwear and other non-durables), and containers and packaging. Since there was negligible recovery for these product types, all the waste generated can be considered discarded. Similar to the plastics method, discards were apportioned based on the proportions of landfilling and combustion for the entire U.S. waste stream (76 percent and 24 percent, respectively). The report aggregates rubber and leather in the MSW stream; an assumed synthetic rubber content was assigned to each product type, as shown in Table H-5.² A carbon content of 85 percent was assigned to synthetic rubber for all product types, according to the weighted average carbon content of rubber consumed for non-tire uses (see Table H-4). For 1999 and 2000, waste generation values were not available so an average annual rate of increase in generation of synthetic rubber in MSW of 1.85 percent (based on statistics in EPA 2000c) was used to project generation in these years.

Table H-5: Rubber and Leather in Municipal Solid Waste in 1998

Product Type	Generation (Gg)	Synthetic Rubber (%)	Carbon Content (%)	Emissions (Tg CO ₂ Eq.)*
Durables (not Tires)	2,141	100%	85%	1.6
Non-Durables	744	100%	85%	0.2
Clothing and Footwear	526	25%	85%	0.1
Other Non-Durables	218	75%	85%	0.1
Containers and Packaging	18	100%	85%	+
Total	2,903	-	-	1.8

* Assumes a fraction oxidized of 98 percent.
 + Less than 0.05 Tg CO₂ Eq.
 - Not applicable

CO₂ from Combustion of Synthetic Fibers

Carbon dioxide emissions from synthetic fibers were estimated as the product of the amount of synthetic fiber discarded annually and the average carbon content of synthetic fiber. Fiber in the MSW stream was estimated from data provided in the *Characterization of Municipal Solid Waste in the United States* (EPA 2000c) for textiles. The amount of synthetic fiber in MSW was estimated by subtracting (a) the amount recovered from (b) the waste generated (see Table H-6). As with the other materials in the MSW stream, discards were apportioned based on the proportions of landfilling and combustion for the entire U.S. waste stream (76 percent and 24 percent, respectively). It was assumed that approximately 55 percent of the fiber was synthetic in origin, based on information received from the Fiber Economics Bureau (DeZan 2000). An average carbon content of 70 percent was assigned to synthetic fiber using the production-weighted average of the carbon contents of the four major fiber types (polyester, nylon, olefin, and acrylic) produced in 1998 (see Table H-7). The equation relating CO₂ emissions to the amount of textiles combusted is shown below. Since 1999 and 2000 values were not provided in the *Characterization* report, generation and recovery rates were forecast by applying their respective average annual growth rates for 1990 through 1998 to the 1998 values.

$$\text{CO}_2 \text{ Emissions from the Combustion of Synthetic Fibers} = \text{Annual Textile Combustion (Gg)} \times (\text{Percent of Total Fiber that is Synthetic}) \times (\text{Average Carbon Content of Synthetic Fiber}) \times$$

² As a biogenic material, the combustion of leather is assumed to have no net carbon dioxide emissions.

(44g CO₂/12 g C)

Table H-6: Textiles in MSW (Gg)

Year	Generation	Recovery	Discards	Combustion
1990	2,884	328	2,557	614
1991	3,064	341	2,723	654
1992	3,225	354	2,901	696
1993	3,458	368	3,090	742
1994	3,674	382	3,291	790
1995	3,674	447	3,227	774
1996	3,832	472	3,361	807
1997	4,090	526	3,564	855
1998	4,269	551	3,781	892
1999*	4,372	566	3,805	913
2000*	4,490	584	3,906	937

*Projected using 1998 data and the 1997 to 2000 Average Annual Growth Rate for Generation (EPA 2000c).

Table H-7: Synthetic Fiber Production in 1998

Fiber	Production (Tg)	Carbon Content	Carbon Equivalent (Tg CO ₂ Eq.)
Polyester	1.8	63%	4.1
Nylon	1.3	64%	3.0
Olefin	1.3	86%	4.1
Acrylic	0.2	68%	0.5
Total	4.6	-	11.7

- Not applicable

N₂O from Municipal Solid Waste Combustion

Estimates of N₂O emissions from MSW combustion in the United States are based on the methodology outlined in the EPA's Compilation of Air Pollutant Emission Factors (EPA 1997). According to this methodology, emissions of N₂O from MSW combustion are the product of the mass of MSW combusted, an emission factor of N₂O emitted per unit mass of waste combusted, and an N₂O emissions control removal efficiency. For MSW combustion in the United States, an emission factor of 30 g N₂O/metric ton MSW and an estimated emissions control removal efficiency of zero percent were used. No information was available on the mass of waste combusted in 2000. It was assumed for the purposes of this calculation that the mass of waste combusted in 2000 was the same as estimated for 1999.

ANNEX I: Methodology for Estimating Emissions from International Bunker Fuels used by the U.S. Military

Bunker fuel emissions estimates for the Department of Defense (DoD) were developed using data primarily generated by the Defense Energy Support Center for aviation and naval fuels (DESC 2001). The Defense Energy Support Center (DESC) of the Defense Logistics Agency (DLA) prepared a special report based on data in the Defense Fuels Automated Management System (DFAMS). DFAMS contains data for 1995 through 2000, but the data set was not complete for years prior to 1995. Fuel quantities for 1990 to 1994 were estimated based on a back-calculation of the 1995 DFAMS values using DLA aviation and marine fuel procurement data. The back-calculation was refined last year to better account for the jet fuel conversion from JP4 to JP8 that occurred within the DoD between 1992 and 1995. Data for marine fuel consumption in 2000 were obtained from the Naval Operations Navy Strategic Mobility/Combat Logistics Division (N42 2001).

Gasoline and diesel fuel totals presented in Table I-1 were estimated using data provided by the military services. The 1991 through 1995 data points were interpolated from the inventory data. The 1997 through 1999 motor gasoline and diesel fuel data were extrapolated from the 1996 inventory data. Growth factors used for other diesel and gasoline were 5.2 percent and -21.1 percent, respectively. Data sets for other diesel and gasoline consumed by the military in 2000 were estimated based on Air Force ground fuels consumption trends. This method produced a result that was more consistent with expected consumption for 2000.

Step 1: Omit Extra-Territorial Fuel Deliveries

Beginning with the complete DFAMS data set for each year, the first step in the development of DoD related emissions from international bunker fuels was to identify data that would be representative of international bunker fuel consumption as that term is defined by decisions of the UNFCCC (i.e., fuel sold to a vessel, aircraft, or installation within the United States or its territories and used in international maritime or aviation transport). Therefore, fuel data were categorized by the location of fuel delivery in order to identify and omit all extra-territorial fuel transactions/deliveries (i.e., sales abroad). Table I-1 displays the fuels that remain at the completion of Step 1, summarized by fuel type.

Step 2: Omit Fuel Transactions Received by Military Services that are not Considered to be International Bunker Fuels

Next, fuel transaction/delivery records were sorted by Military Service. The following assumptions were made regarding bunker fuel use by Service, leaving only the Navy and Air Force as users of military international bunker fuels.

- Only fuel delivered to a ship, aircraft, or installation in the United States can be a potential international bunker fuel. Fuel consumed in international aviation or marine transport should be included in the bunker fuel estimate of the country where the ship or aircraft was fueled. Fuel consumed entirely within a country's borders is not bunker fuel.
- Based on discussions with the Army staff, only an extremely small percentage of Army aviation emissions, and none of its watercraft emissions, qualified as bunker fuel emissions. The magnitude of these emissions was judged to be insignificant when compared to Air Force and Navy emissions. Based on this, Army bunker fuel emissions are assumed to be zero.
- Marine Corps aircraft operating while embarked consume fuel reported as delivered to the Navy. Bunker fuel emissions from embarked Marine Corps aircraft are reported in the Navy bunker fuel estimates. Bunker fuel emissions from other Marine Corps operations and training are assumed to be zero.
- Bunker fuel emissions from other DoD and non-DoD activities (i.e., other federal agencies) that purchase fuel from DESC are assumed to be zero.

Step 3: Omit Land-Based Fuels

Navy and Air Force land-based fuel consumption (i.e., fuel not used by ships or aircraft) were also omitted. The remaining fuels, listed below, were potential military international bunker fuels.

- Marine: naval distillate fuel (F76), marine gas oil (MGO), and intermediate fuel oil (IFO).
- Aviation: jet fuels (JP8, JP5, JP4, JAA, JA1, and JAB).

At the completion of step 3 of the 2000 estimate, it was apparent that the Navy maritime data provided by DESC were abnormal compared to those data for each year from 1995 to 1999. The Navy fuels and logistics office identified a separate data set that was used to compare the DESC 2000 data for Navy maritime fuel consumption. After comparing the 2000 data from DESC to that from the Navy, it was determined that the Navy data (N42 2001) should be used as the source of the 2000 maritime values. However, DoD will continue to investigate the 2000 maritime data from DESC, which Navy fuels experts consider an anomaly.

Step 4: Determine Bunker Fuel Percentages

Next it was necessary to determine what percent of the marine and aviation fuels were used as international bunker fuels. Military aviation bunkers include international operations (i.e., sorties that originate in the United States and terminate in a foreign country), operations conducted from naval vessels at sea, and operations conducted from U.S. installations principally over international water in direct support of military operations at sea (e.g., anti-submarine warfare flights). For the Air Force, a bunker fuel weighted average was calculated based on flying hours by major command. International flights were weighted by an adjustment factor to reflect the fact that they typically last longer than domestic flights. In addition, a fuel use correction factor was used to account for the fact that transport aircraft burn more fuel per hour of flight than most tactical aircraft. The Air Force bunker fuel percentage was determined to be 13.2 percent. This percentage was multiplied by total annual Air Force aviation fuel delivered for U.S. activities, producing an estimate for international bunker fuel consumed by the U.S. Air Force. The naval aviation bunker fuel percentage of total fuel was calculated using flying hour data from *Chief of Naval Operations Flying Hour Projection System Budget Analysis Report for FY 1998* (N45 1998), and estimates of bunker fuel percent of flights provided by the fleet. The naval aviation bunker fuel percentage, determined to be 40.4 percent, was multiplied by total annual Navy aviation fuel delivered for U.S. activities, yielding total Navy aviation bunker fuel consumed.

For marine bunkers, fuels consumed while ships were underway were assumed to be bunker fuels. For 2000, the Navy reported that 79 percent of vessel operations were underway, while the remaining 21 percent of operations occurred in port (i.e., pierside). Therefore, the Navy maritime bunker fuel percentage was determined to be 79 percent for 2000. The percentage of time underway may vary from year-to-year and the 2000 value represents a change from previous years, for which the bunker fuel percentage of 87 percent was applied. Table I-2 and Table I-3 display DoD bunker fuel totals for the Navy and Air Force.

Step 5: Calculate Emissions from Military International Bunker Fuels

Bunker fuel totals were multiplied by appropriate emission factors to determine greenhouse gas emissions (see Table I-4 and Table I-5).

The rows labeled 'U.S. Military' and 'U.S. Military Naval Fuels' within Table 2-43 and Table 2-44 in the Energy Chapter were based on the international bunker fuel totals provided in Table I-2 and Table I-3, below. Total CO₂ emissions from military bunker fuels are presented in Table I-6. Carbon dioxide emissions from aviation bunkers and distillate marine bunkers presented in Table 2-7 are the total of military plus civil aviation and civil marine bunker fuels, respectively. The military component of each total is based on fuels tallied in Table I-2 and Table I-3.

Table I-1: Transportation Fuels from Domestic Fuel Deliveries^a (Million Gallons)

Vehicle Type/Fuel	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000
Aviation	4,598.45	4,562.84	3,734.49	3,610.85	3,246.23	3,099.93	2,941.91	2,697.28	2,764.76	2,670.28	2,711.16
JP8	285.75	283.54	234.46	989.38	1,598.07	2,182.80	2,253.15	2,083.64	2,145.89	2,101.52	2,169.42
JP5	1,025.36	1,017.42	832.71	805.14	723.84	691.22	615.83	552.77	515.56	505.50	472.10
Other Jet Fuels	3,287.31	3,261.86	2,667.29	1,816.30	924.30	225.89	72.92	60.86	103.29	63.25	69.65
Aviation Gasoline	0.03	0.03	0.02	0.02	0.02	0.02	0.01	+	0.02	0.01	0.01
Marine	686.80	632.61	646.18	589.37	478.59	438.91	493.34	639.85	674.22	598.86	680.19
Middle Distillate (MGO)	+	+	+	+	+	+	38.52	47.48	51.14	49.22	97.61
Naval Distillate (F76)	686.80	632.61	646.18	589.37	478.59	438.91	448.96	583.41	608.39	542.94	549.57
Intermediate Fuel Oil (IFO) ^b	+	+	+	+	+	+	5.86	8.95	14.69	6.70	33.01
Other	717.11	590.41	491.68	415.10	356.06	310.95	276.90	251.66	233.47	220.92	201.45
Diesel	93.04	97.88	102.96	108.31	113.94	119.86	126.09	132.64	139.53	146.78	126.63
Gasoline	624.07	492.53	388.72	306.78	242.12	191.09	150.81	119.02	93.94	74.14	74.81
Total (Including Bunkers)	6,002.37	5,785.85	4,872.34	4,615.32	4,080.89	3,849.78	3,712.15	3,588.79	3,672.45	3,490.06	3,592.80

Note: Totals may not sum due to independent rounding.

^a Includes fuel consumption in United States and U.S. Territories.

^b Intermediate fuel oil (IFO 180 and IFO 380) is a blend of distillate and residual fuels. IFO is used by the Military Sealift Command.
 + Does not exceed 0.005 million gallons.

ANNEX J: Methodology for Estimating HFC, and PFC Emissions from Substitution of Ozone Depleting Substances

The Vintaging Model was developed as a tool for estimating the annual chemical emissions from industrial sectors that have historically used ozone depleting substances (ODSs) in their products. Under the terms of the Montreal Protocol and the Clean Air Act Amendments of 1990, the domestic production of ODSs, CFCs, halons, carbon tetrachloride, methyl chloroform, and hydrochlorofluorocarbons (HCFCs) – has been drastically reduced, forcing these industrial sectors to transition to more ozone friendly chemicals. As these industries have moved toward ODSs alternatives such as hydrofluorocarbons (HFCs) and perfluorocarbons (PFCs), the Vintaging Model has evolved into a tool for estimating the rise in consumption and emissions of these alternatives, and the decline of ODS consumption and emissions.

The Vintaging Model estimates emissions from six industrial sectors: refrigeration and air-conditioning, foams, aerosols, solvents, fire extinguishing, and sterilization. Within these sectors, there are over 40 independently modeled end-uses. The model requires information on the market growth for each of the end-uses, as well as a history of the market transition from ODSs to alternatives. As ODSs are phased out, a percentage of the market share originally filled by the ODSs is allocated to each of its substitutes.

The model, named for its method of tracking the emissions of annual “vintages” of new equipment that enter into service, is a “bottom-up” model. It models the consumption of chemicals based on estimates of the quantity of equipment or products sold, serviced, and retired each year, and the amount of the chemical required to manufacture and/or maintain the equipment. The Vintaging Model makes use of this market information to build an inventory of the in-use stocks of the equipment in each of the end-uses. Emissions are estimated by applying annual leak rates, service emission rates, and disposal emission rates to each population of equipment. By aggregating the emission and consumption output from the different end-uses, the model produces estimates of total annual use and emissions of each chemical.

The Vintaging Model synthesizes data from a variety of sources, including data from the ODS Tracking System maintained by the Global Programs Division and information from submissions to EPA under the Significant New Alternatives Policy (SNAP) program. Published sources include documents prepared by the United Nations Environment Programme (UNEP) Technical Options Committees, reports from the Alternative Fluorocarbons Environmental Acceptability Study (AFEAS), and proceedings from the international conferences on ozone protection technologies and Earth Technologies Forum. EPA also coordinates extensively with numerous trade associations and individual companies. For example, the Alliance for Responsible Atmospheric Policy, the Air-Conditioning and Refrigeration Institute, the Association of Home Appliance Manufacturers, the American Automobile Manufacturers Association, and many of their member companies, have provided valuable information over the years. In some instances the unpublished information that the EPA uses in the model is classified as Confidential Business Information (CBI). The annual emissions inventories of chemicals are aggregated in such a way that CBI cannot be inferred. Full public disclosure of the inputs to the Vintaging Model would jeopardize the security of the CBI that has been entrusted to the EPA.

The following sections discuss the forms of the emission estimating equations used in the Vintaging Model for each broad end-use category. These equations are applied separately for each chemical used within each of approximately 40 different end-uses. In the majority of these end-uses, more than one ODS substitute chemical is used.

In general, the modeled emissions are a function of the amount of chemical consumed in each end-use market. Estimates of the consumption of ODS alternatives can be inferred by extrapolating forward in time from the amount of regulated ODSs used in the early 1990s. Using data gleaned from a variety of sources, assessments are made regarding which alternatives will likely be used, and what fraction of the ODS market in each end-use will be captured by a given alternative. By combining this with estimates of the total end-use market growth, a consumption value can be estimated for each chemical used within each end-use.

Alternatively, “top-down” information on total U.S. consumption of a given chemical is sometimes available. This data can be used by estimating the fraction of this total that is consumed within each end-use. These allocation schemes are guided by EPA’s synthesis of the data available through the aforementioned sources.

Methodology

The methodology used by the vintaging model to calculate emissions varies by end-use sector. The methodologies and specific equations used by end-use sector are presented below.

Refrigeration and Air-Conditioning

For refrigeration and air conditioning products, emission calculations are split into two categories: emissions during equipment lifetime, which arise from annual leakage and service losses, and disposal emissions, which occur at the time of discard. Two separate steps are required to calculate the lifetime emissions from leakage and service, and the emissions resulting from disposal of the equipment. These lifetime emissions and disposal emissions are summed to calculate the total emissions from refrigeration and air-conditioning. As new technologies replace older ones, it is generally assumed that there are improvements in their leak, service, and disposal emission rates.

Step 1. Calculate lifetime emissions

Lifetime emissions from any piece of equipment include both the amount of chemical leaked during equipment operation and during service recharges. Emissions from leakage and servicing can be expressed as follows:

$$Es_j = (l_a + l_s) \times \sum_{i=1 \rightarrow k} Qc_{j-i+1} \quad \text{for } i=1 \rightarrow k$$

Where,

- Es = Emissions from Equipment Serviced. Emissions in year j from normal leakage and servicing (including recharging) of equipment.
- l_a = Annual Leak Rate. Average annual leak rate during normal equipment operation (expressed as a percentage of total chemical charge).
- l_s = Service Leak Rate. Average leakage during equipment servicing (expressed as a percentage of total chemical charge).
- Qc = Quantity of Chemical in New Equipment. Total amount of a specific chemical used to charge new equipment in a given year, j , by weight.
- K = Lifetime. The average lifetime of the equipment.

Step 2. Calculate disposal emissions

The disposal emission equations assume that a certain percentage of the chemical charge will be emitted to the atmosphere when that vintage is discarded. Disposal emissions are thus a function of the quantity of chemical contained in the retiring equipment fleet and the proportion of chemical released at disposal:

$$Ed_j = Qc_{j-k+1} \times [1 - (rm \times rc)]$$

Where,

- Ed = Emissions from Equipment Disposed. Emissions in year j from the disposal of equipment.
- Qc = Quantity of Chemical in New Equipment. Total amount of a specific chemical used to charge new equipment in a given year, j , by weight.
- rm = Chemical Remaining. Amount of chemical remaining in equipment at the time of disposal (expressed as a percentage of total chemical charge)

- rc = Chemical Recovery Rate. Amount of chemical that is recovered just prior to disposal (expressed as a percentage of chemical remaining at disposal (rm))
- k = Lifetime. The average lifetime of the equipment.

Step 3. Calculate total emissions

Finally, lifetime and disposal emissions are summed to provide an estimate of total emissions.

$$E_j = Es_j + Ed_j$$

Where,

- E = Total Emissions. Emissions from refrigeration and air conditioning equipment in year j .
- Es = Emissions from Equipment Serviced. Emissions in year j from normal leakage and servicing (recharging) of equipment.
- Ed = Emissions from Equipment Disposed. Emissions in year j from the disposal of equipment.

Aerosols

All HFCs and PFCs used in aerosols are assumed to be emitted in the year of manufacture. Since there is currently no aerosol recycling, it is assumed that all of the annual production of aerosol propellants is released to the atmosphere. The following equation describes the emissions from the aerosols sector.

$$E_j = Qc_j$$

Where,

- E = Emissions. Total emissions of a specific chemical in year j from use in aerosol products, by weight.
- Qc = Quantity of Chemical. Total quantity of a specific chemical contained in aerosol products sold in year j , by weight.

Solvents

Generally, most solvents are assumed to remain in the liquid phase and are not emitted as gas. Thus, emissions are considered “incomplete,” and are a fixed percentage of the amount of solvent consumed in a year. The remainder of the consumed solvent is assumed to be reused or disposed without being released to the atmosphere. The following equation calculates emissions from solvent applications.

$$E_j = l \times Qc_j$$

Where,

- E = Emissions. Total emissions of a specific chemical in year j from use in solvent applications, by weight.
- L = Percent Leakage. The percentage of the total chemical that is leaked to the atmosphere.
- Qc = Quantity of Chemical. Total quantity of a specific chemical sold for use in solvent applications in the year j , by weight.

Fire Extinguishing

Total emissions from fire extinguishing are assumed, in aggregate, to equal a percentage of the total quantity of chemical in operation at a given time. For modeling purposes, it is assumed that fire extinguishing equipment leaks at a constant rate for an average equipment lifetime. This percentage varies for streaming and flooding equipment.

Step 1. Calculate emissions from streaming equipment

$$E_j = l_{se} \times \sum Qc_{j-i+1} \quad \text{for } i=1 \rightarrow k$$

Where,

E = Emissions. Total emissions of a specific chemical in year j for streaming fire extinguishing equipment, by weight.

l_{se} = Percent Leakage. The percentage of the total chemical in operation that is leaked to the atmosphere.

Qc = Quantity of Chemical from streaming equipment. Total amount of a specific chemical used in new streaming fire extinguishing equipment in a given year, j , by weight.

k = Lifetime. The average lifetime of the equipment.

Step 2. Calculate emissions from flooding equipment

$$E_j = l_{fe} \times \sum Qc_{j-i+1} \quad \text{for } i=1 \rightarrow k$$

Where,

E = Emissions. Total emissions of a specific chemical in year j for streaming fire extinguishing equipment, by weight.

l_f = Percent Leakage. The percentage of the total chemical in operation that is leaked to the atmosphere.

Qc = Quantity of Chemical from flooding equipment. Total amount of a specific chemical used in new streaming fire extinguishing equipment in a given year, j , by weight.

k = Lifetime. The average lifetime of the equipment.

Foam Blowing

Foams are given emission profiles depending on the foam type (open cell or closed cell). Open cell foams are assumed to be 100 percent emissive in the year of manufacture. Closed cell foams are assumed to emit a portion of their total HFC or PFC content upon manufacture, a portion at a constant rate over the lifetime of the foam, and a portion at disposal.

Step 1. Calculate emissions from open-cell foam

Emissions from open-cell foams are calculated using the following equation.

$$E_j = Qc_j$$

Where,

E = Emissions. Total emissions of a specific chemical in year j used for open-cell foam blowing, by weight.

Qc = Quantity of Chemical. Total amount of a specific chemical used for open-cell foam blowing in year j , by weight.

Step 2. Calculate emissions from closed-cell foam

Emissions from closed-cell foams are calculated using the following equation.

$$E_j = \sum (ef_i \times Qc_{j-i+1}) \quad \text{for } i=1 \rightarrow k$$

Where,

E = Emissions. Total emissions of a specific chemical in year j for closed-cell foam blowing, by weight.

ef = Emission Factor. Percent of foam's original charge emitted in each year ($1 \rightarrow k$). This emission factor is generally variable, including a rate for manufacturing emissions (occurs in the first year of foam life), annual emissions (every year throughout the foam lifetime), and disposal emissions (occurs during the final year of foam life).

Qc = Quantity of Chemical. Total amount of a specific chemical used in closed-cell foams in year j .

k = Lifetime. Average lifetime of foam product.

Sterilization

For sterilization applications, all chemicals that are used in the equipment in any given year are assumed to be emitted in that year, as shown in the following equation.

$$E_j = Qc_j$$

Where,

E = Emissions. Total emissions of a specific chemical in year j from use in sterilization equipment, by weight.

Qc = Quantity of Chemical. Total quantity of a specific chemical used in sterilization equipment in year j , by weight.

Model Output

By repeating these calculations for each year, the Vintaging Model creates annual profiles of use and emissions for ODS and ODS substitutes. The results can be shown for each year in two ways: 1) on a chemical-by-chemical basis, summed across the end-uses, or 2) on an end-use basis. Values for use and emissions are calculated both in metric tons and in teragrams of carbon dioxide equivalents (Tg CO₂ Eq.). The conversion of metric tons of chemical to Tg CO₂ Eq. is accomplished through a linear scaling of tonnage by the global warming potential (GWP) of each chemical.

Throughout its development, the Vintaging Model has undergone annual modifications. As new or more accurate information becomes available, the model is adjusted in such a way that both past and future emission estimates are often altered.

ANNEX K: Methodology for Estimating CH₄ Emissions from Enteric Fermentation

Methane emissions from enteric fermentation were estimated for five livestock categories: cattle, horses, sheep, swine and goats. Emissions from cattle represent the majority of U.S. emissions, consequently, the more detailed IPCC Tier 2 methodology was used to estimate emissions from cattle and the IPCC Tier 1 methodology was used to estimate emissions from the other types of livestock.

Estimate Methane Emissions from Cattle

This section describes the process used to estimate methane emissions from cattle enteric fermentation. A model based on recommendations provided in IPCC/UNEP/OECD/IEA (1997) and IPCC (2000) was developed that uses information on population, energy requirements, digestible energy, and methane conversion rates to estimate methane emissions. The emission methodology consists of the following three steps: (1) characterize the cattle population to account for animal population categories with different emissions profiles; (2) characterize cattle diets to generate information needed to estimate emissions factors; and (3) estimate emissions using these data and the IPCC Tier 2 equations.

Step 1. Characterize U.S. Cattle Population

Each stage in the cattle lifecycle was modeled to simulate the cattle population from birth to slaughter. This level of detail accounts for the variability in methane emissions associated with each life stage. Given that the time in which cattle can be in a stage can be less than one year (e.g., beef calves are weaned at 7 months), the stages are modeled on a per month basis. The type of cattle use also impacts methane emissions (e.g., beef versus dairy). Consequently, cattle life stages were modeled for several categories of dairy and beef cattle. These categories are listed in Table K-1.

Table K-1: Cattle Population Categories Used for Estimating Methane Emissions

Dairy Cattle	Beef Cattle
Calves	Calves
Heifer Replacements	Heifer Replacements
Cows	Heifer and Steer Stockers
	Animals in Feedlots
	Cows
	Bulls

The key variables tracked for each of these cattle population categories (except bulls¹) are as follows:

- *Calving rates:* The number of animals born on a monthly basis was used to initiate monthly cohorts and to determine population age structure. The number of calves born each month was obtained by multiplying annual births by the percentage of births by month. Annual birth information was taken from USDA (2001a). Average percentage of births by month for beef from USDA (USDA/APHIS/VS 1998, 1994, 1993) were used for 1990 through 2000. For dairy animals, birth rates were assumed constant throughout the year. Whether calves were born to dairy or beef cows was estimated using the dairy cow calving rate and the total dairy cow population to determine the percent of births attributable to dairy cows, with the remainder assumed to be attributable to beef cows.

¹ Only end-of-year census population statistics and a national emission factors are used to estimate methane emissions from the bull population.

- *Average weights and weight gains:* Average weights were tracked for each monthly age group using starting weight and monthly weight gain estimates. Weight gain (i.e., pounds per month) was estimated based on weight gain needed to reach a set target weight, divided by the number of months remaining before target weight was achieved. Birth weight was assumed to be 88 pounds for both beef and dairy animals. Weaning weights were estimated to range from 480 to 575 pounds. Other reported target weights were available for 12, 15, 24, and 36 month-old animals. Live slaughter weights were derived from dressed slaughter weight data (USDA 2001c, 2000c). Live slaughter weight was estimated as dressed weight divided by 0.63.
- *Feedlot placements:* Feedlot placement statistics were available that specify placement of animals from the stocker population into feedlots on a monthly basis by weight class. The model used these data to shift a sufficient number of animals from the stocker cohorts into the feedlot populations to match the reported data. After animals are placed in feedlots they progress through two steps. First, animals spend time on a step-up diet to become acclimated to the new feed type. Animals are then switched to a finishing diet for a period of time before they are slaughtered. The length of time an animal spends in a feedlot depends on the start weight (i.e., placement weight), the rate of weight gain during the start-up and finishing phase of diet, and the end weight (as determined by weights at slaughter). Weight gain during start-up diets is estimated to be 2.8 to 3 pounds per day. Weight gain during finishing diets is estimated to be 3 to 3.3 pounds per day (Johnson 1999). All animals are estimated to spend 25 days in the step-up diet phase (Johnson 1999). Length of time finishing can be calculated based on start weight, weight gain per day, and target slaughter weight.
- *Pregnancy and lactation:* Energy requirements and hence, composition of diets, level of intake, and emissions for particular animals, are greatly influenced by whether the animal is pregnant or lactating. Information is therefore needed on the percentage of all mature animals that are pregnant each month, as well as milk production, to estimate methane emissions. A weighted average percent of pregnant cows each month was estimated using information on births by month and average pregnancy term. For beef cattle, a weighted average total milk production per animal per month was estimated using information on typical lactation cycles and amounts (NRC 1999), and data on births by month. This results in a range of weighted monthly lactation estimates expressed as lbs/animal/month. The monthly estimates from January to December are 3.33, 5.06, 8.70, 12.01, 13.58, 13.32, 11.67, 9.34, 6.88, 4.45, 3.04, and 2.77. Monthly estimates for dairy cattle were taken from USDA monthly milk production statistics.
- *Death rates:* This factor is applied to all heifer and steer cohorts to account for death loss within the model on a monthly basis. The death rates are estimated by determining the death rate that results in model estimates of the end-of-year population for cows that match the published end-of-year population census statistics.

Number of animals per category each month: The population of animals per category is calculated based on number of births (or graduates) into the monthly age group minus those animals that die or are slaughtered and those that graduate to next category (including feedlot placements). These monthly age groups are tracked in the enteric fermentation model to estimate emissions by animal type on a regional basis. Table K-2 provides the cattle population estimates as output from the enteric fermentation model from 1990 through 2000. This table includes the population categories used in the model to estimate total emissions, including tracking emissions that occur the following year for feedlot animals placed late in the year. Dairy lactation estimates for 1990 through 2000 are shown in Table K-3. Table K-4 provides the target weights used to track average weights of cattle by animal type. Table K-5 provides a summary of the reported feedlot placement statistics for 2000.

Cattle population data were taken from U.S. Department of Agriculture (USDA) National Agricultural Statistics Service (NASS) reports. The USDA publishes monthly, annual, and multi-year livestock population and production estimates. Multi-year reports include revisions to earlier published data. Cattle and calf populations, feedlot placement statistics (e.g., number of animals placed in feedlots by weight class), slaughter numbers, and lactation data were obtained from the USDA (1990-2001). Beef calf birth percentages were obtained from the National Animal Health Monitoring System (NAHMS) (USDA/APHIS/VS 1998, 1994, 1993). Estimates of the number of animals in different population categories of the model differ from the reported national population statistics. This difference is due to model output indicating the average number of animals in that category for the year rather than the end of year population census.

Step 2. Characterize U.S. Cattle Population Diets

To support development of digestible energy (DE, the percent of gross energy intake digestible to the animal) and methane conversion rate (Y_m , the fraction of gross energy converted to methane) values for each of the cattle population categories, data were collected on diets considered representative of different regions. For both grazing animals and animals being fed mixed rations, representative regional diets were estimated using information collected from state livestock specialists and from USDA (1996a). The data for each of the diets (e.g., proportions of different feed constituents, such as hay or grains) were used to determine chemical composition for use in estimating DE and Y_m for each animal type. Additional detail on the regional diet characterization is provided in EPA (2000).

DE and Y_m were used to estimate methane emissions from enteric fermentation and vary by diet and animal type. The IPCC recommends Y_m values of 3.5 to 4.5 percent for feedlot cattle and 5.5 to 6.5 percent for all other cattle. Given the availability of detailed diet information for different regions and animal types in the United States, DE and Y_m values unique to the United States² were developed. Table K-6 shows the regional DE, the Y_m , and percent of total U.S. cattle population in each region based on 2000 data.

DE and Y_m values were estimated for each cattle population category based on physiological modeling and expert opinion. DE and Y_m values for dairy cows and most grazing animals were estimated using a model (Donovan and Baldwin 1999) that represents physiological processes in the ruminant animals. The three major categories of input required by the model are animal description (e.g., cattle type, mature weight), animal performance (e.g., initial and final weight, age at start of period), and feed characteristics (e.g., chemical composition, habitat, grain or forage). Data used to simulate ruminant digestion is provided for a particular animal that is then used to represent a group of animals with similar characteristics. The model accounts for differing diets (i.e., grain-based, forage-based, range-based), so that Y_m values for the variable feeding characteristics within the U.S. cattle population can be estimated.

For feedlot animals, DE and Y_m values were taken from Johnson (1999). In response to peer reviewer comments (Johnson 2000), values for dairy replacement heifers are based on EPA (1993).

Step 3. Estimate Methane Emissions from Cattle

Emissions were estimated in three steps: a) determine gross energy intake using the IPCC (2000) equations, b) determine an emissions factor using the GE values and other factors, and c) sum the daily emissions for each animal type. The necessary data values include:

- Body Weight (kg)
- Weight Gain (kg/day)
- Net Energy for Activity (C_a)³
- Standard Reference Weight⁴ (Dairy = 1,324 kg; Beef = 1,195 kg)
- Milk Production (kg/day)
- Milk Fat (percent of fat in milk = 4)
- Pregnancy (percent of population that is pregnant)
- DE (percent of gross energy intake digestible)
- Y_m (the fraction of gross energy converted to methane)

² In some cases, the Y_m values used for this analysis extend beyond the range provided by the IPCC. However, EPA believes that these values are representative for the U.S. due to the research conducted to characterize the diets of U.S. cattle and to assess the Y_m values associated with different animal performance and feed characteristics in the United States.

³ Zero for feedlot conditions, 0.17 for grazing conditions, 0.37 for high quality grazing conditions. C_a factor for dairy cows is weighted to account for the fraction of the population in the region that grazes during the year.

⁴ Standard Reference Weight is used in the model to account for breed potential.

Step 3a: Gross Energy, GE:

As shown in the following equation, Gross Energy (GE) is derived based on the net energy estimates and the feed characteristics. Only variables relevant to each animal category are used (e.g., estimates for feedlot animals do not require the NE_l factor). All net energy equations are provided in IPCC (2000).

$$GE = [((NE_m + NE_{mobilized} + NE_a + NE_l + NE_p) / \{NE_{ma}/DE\}) + (NE_g / \{NE_{ga}/DE\})] / (DE / 100)$$

where,

GE = gross energy (MJ/day)

NE_m = net energy required by the animal for maintenance (MJ/day)

NE_{mobilized} = net energy due to weight loss (mobilized) (MJ/day)

NE_a = net energy for animal activity (MJ/day)

NE_l = net energy for lactation (MJ/day)

NE_p = net energy required for pregnancy (MJ/day)

{NE_{ma}/DE} = ratio of net energy available in a diet for maintenance to digestible energy consumed

NE_g = net energy needed for growth (MJ/day)

{NE_{ga}/DE} = ratio of net energy available for growth in a diet to digestible energy consumed

DE = digestible energy expressed as a percentage of gross energy (percent)

Step 3b: Emission Factor

The emissions factor (DayEmit) was determined using the GE value and the methane conversion factor (Y_m) for each category. This is shown in the following equation:

$$\text{DayEmit} = [GE \times Y_m] / [55.65 \text{ MJ/kg CH}_4]$$

where,

DayEmit = emission factor (kg CH₄/head/day)

GE = gross energy intake (MJ/head/day)

Y_m = methane conversion rate which is the fraction of gross energy in feed converted to methane (percent)

The daily emission factors were estimated for each animal type, weight and region. The implied national annual average emission factors for each of the animal categories are shown in Table K-7. The implied factors are not used in the inventory, but are provided for comparative purposes only.

Step 3c: Estimate Total Emissions

Emissions were summed for each month and for each population category using the daily emission factor for a representative animal and the number of animals in the category. The following equation was used:

$$\text{Emissions} = \text{DayEmit} \times \text{Days/Month} \times \text{SubPop}$$

where,

DayEmit = the emission factor for the subcategory (kg CH₄/head/day)

Days/Month = the number of days in the month

SubPop = the number of animals in the subcategory during the month

This process was repeated for each month, and the totals for each subcategory were summed to achieve an emissions estimate for the entire year. The estimates for each of the 10 subcategories of cattle are listed in Table K-8. The

emissions for each subcategory were then summed to estimate total emissions from beef cattle and dairy cattle for the entire year. The total emissions from 1990 through 2000 are shown in Table K-9.

Emission Estimates from Other Livestock

All livestock population data, except for horses, were taken from U.S. Department of Agriculture (USDA) National Agricultural Statistics Service (NASS) reports. For each animal category, the USDA publishes monthly, annual, and multi-year livestock population and production estimates. Multi-year reports include revisions to earlier published data. Recent reports were obtained from the USDA Economics and Statistics System, while historical data were downloaded from the USDA-NASS. The Food and Agriculture Organization (FAO) publishes horse population data. These data were accessed from the FAOSTAT database at <http://apps.fao.org/>. Methane emissions from sheep, goats, swine, and horses were estimated by multiplying published national population estimates by the national emission factor for each year. Table K-10 shows the populations used for these other livestock from 1990 to 2000 and Table K-11 shows the emission factors used for these other livestock.

A complete time series of enteric fermentation emissions from livestock is shown in Table K-12 (Tg CO₂ Eq.) and Table K-13 (Gg).

Table K-2: Estimates of Average Annual Populations of U.S. Cattle (Thousand Head)^a

Livestock Type	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000
Calves 0-6 months	22,561	22,531	22,707	23,004	23,346	23,468	23,255	22,810	22,674	22,655	22,564
Dairy											
Cows	10,015	9,965	9,728	9,658	9,528	9,487	9,416	9,309	9,191	9,133	9,190
Replacements 7-11 months	1,214	1,219	1,232	1,230	1,228	1,220	1,205	1,182	1,194	1,188	1,194
Replacements 12-23 months	2,915	2,874	2,901	2,926	2,907	2,905	2,877	2,838	2,797	2,846	2,811
Beef											
Cows	32,454	32,520	33,007	33,365	34,650	35,156	35,228	34,271	33,683	33,745	33,569
Replacements 7-11 months	1,269	1,315	1,402	1,465	1,529	1,492	1,462	1,378	1,321	1,303	1,313
Replacements 12-23 months	2,967	3,063	3,182	3,393	3,592	3,647	3,526	3,391	3,212	3,110	3,097
Steer Stockers	7,864	7,362	8,458	8,786	7,900	8,700	8,322	7,839	7,647	7,203	6,798
Heifer Stockers	4,806	4,572	4,809	4,944	4,715	5,083	4,933	4,987	4,838	4,671	4,407
Total Adjusted Feedlot ^b	10,574	10,461	10,409	9,901	10,775	11,344	11,353	11,528	11,517	12,858	12,774
Bulls	2,180	2,198	2,220	2,239	2,306	2,392	2,392	2,325	2,235	2,241	2,197
Total Placements ^{c,d}	25,587	25,396	25,348	25,586	26,615	27,623	27,580	28,518	27,111	29,424	27,921

Source: Enteric Fermentation Model.

^a Populations represent the average of each population category throughout the year.

^b Total Adjusted Feedlot = Average number in feedlots accounting for current year plus the population carried over from the previous year (e.g., the “next year” population numbers from this table are added into the following years “adjusted numbers”).

^c Placements represent a flow of animals from backgrounding situations to feedlots rather than an average annual population estimate.

^d Reported placements from USDA are adjusted using a scaling factor based on the slaughter to placement ratio.

Table K-3: Dairy Lactation by Region (lbs· year/cow)*

Year	Northern						
	California	West	Great Plains	Southcentral	Northeast	Midwest	Southeast
1990	18,800	16,769	13,502	12,397	14,058	14,218	12,943
1991	18,771	16,631	13,316	12,389	14,560	14,555	12,850
1992	19,072	17,838	13,597	12,710	15,135	15,028	13,292
1993	18,852	17,347	14,109	13,034	14,937	15,203	13,873
1994	20,203	17,890	14,496	13,236	15,024	15,374	14,200
1995	19,573	17,724	14,650	13,228	15,398	15,728	14,384
1996	19,161	18,116	14,872	13,215	15,454	15,596	14,244
1997	19,829	18,248	15,013	13,212	15,928	16,027	14,548
1998	19,442	18,377	15,489	13,580	16,305	16,494	14,525
1999	20,788	19,330	15,910	13,476	16,571	16,655	14,930

2000	21,169	20,828	17,196	15,323	17,544	17,255	15,201
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Source: USDA (2001d).

* Beef lactation data were developed using the methodology described in the text.

Table K-4: Target Weights for Use in Estimating Average Weights and Weight Gains (lbs)

Cattle Type	Typical Weights
Beef Replacement Heifer Data	
Replacement Weight at 15 months	715
Replacement Weight at 24 months	1,078
Mature Weight at 36 months	1,172
Dairy Replacement Heifer Data	
Replacement Weight at 15 months	800
Replacement Weight at 24 months	1,225
Mature Weight at 36 months	1,350
Stockers Data – Grazing/Forage Based Only	
Steer Weight Gain/Month to 12 months	45
Steer Weight Gain/Month to 24 months	35
Heifer Weight Gain/Month to 12 months	35
Heifer Weight Gain/Month to 24 months	30

Source: Feedstuffs (1998), Western Dairyman (1998), Johnson (1999), NRC (1999).

Table K-5: Feedlot Placements in the United States for 2000* (Number of animals placed in Thousand Head)

Weight When Placed	Jan	Feb	Mar	Apr	May	Jun	Jul	Aug	Sep	Oct	Nov	Dec	Total
< 600 lbs	489	351	333	301	382	347	424	573	775	1066	757	504	6,302
600 - 700 lbs	691	476	411	310	471	380	386	504	612	755	559	516	6,071
700 - 800 lbs	654	596	717	577	794	498	592	691	681	531	405	406	7,142
> 800 lbs	382	457	570	519	658	439	505	672	618	477	293	273	5,863
Total	2,216	1,880	2,031	1,707	2,305	1,664	1,907	2,440	2,686	2,829	2,014	1,699	25,378

Source: USDA (2001g).

Note: Totals may not sum due to independent rounding.

* Data were available for 1996 through 2000. Data for 1990 to 1995 were based on the average of monthly placements from the 1996 to 1998 reported figures.

Table K-6: Regional Digestible Energy (DE), Methane Conversion Rates (Y_m), and population percentages for Cattle in 2000

Animal Type	Data	California	West	Northern Great Plains	Southcentral	Northeast	Midwest	Southeast
Beef Repl. Heif. ^a	DE ^b	67	66	68	66	64	68	68
	Ym ^c	8.0%	7.4%	8.0%	8.3%	8.4%	8.0%	7.8%
	Pop. ^d	3%	11%	31%	24%	2%	13%	16%
Dairy Repl. Heif. ^a	DE	66	66	66	64	68	66	66
	Ym	5.9%	5.9%	5.6%	6.4%	6.3%	5.6%	6.9%
	Pop.	18%	11%	5%	4%	19%	36%	7%
Steer Stockers ^a	DE	67	66	68	66	64	68	68
	Ym	8.0%	7.4%	8.0%	8.3%	8.4%	8.0%	7.8%
	Pop.	4%	8%	41%	23%	2%	18%	4%
Heifer Stockers ^a	DE	67	66	68	66	64	68	68
	Ym	8.0%	7.4%	8.0%	8.3%	8.4%	8.0%	7.8%
	Pop.	2%	7%	50%	22%	1%	15%	4%
Steer Feedlot ^e	DE	85	85	85	85	85	85	85
	Ym	3.0%	3.0%	3.0%	3.0%	3.0%	3.0%	3.0%

ANNEX L: Methodology for Estimating CH₄ and N₂O Emissions from Manure Management

This annex presents a discussion of the methodology used to calculate methane and nitrous oxide emissions from manure management systems. More detailed discussions of selected topics may be found in supplemental memoranda in the supporting docket to this inventory.

The following steps were used to estimate methane and nitrous oxide emissions from the management of livestock manure. Nitrous oxide emissions associated with pasture, range, or paddock systems and daily spread systems are included in the emissions estimates for Agricultural Soil Management.

Step 1: Livestock Population Characterization Data

Annual animal population data for 1990 through 2000 for all livestock types, except horses and goats, were obtained from the USDA National Agricultural Statistics Service (USDA, 1994a-b, 1995a-b, 1998a-b, 1999a-c, 2000a-g, 2001a-f). The actual population data used in the emissions calculations for cattle and swine were downloaded from the USDA National Agricultural Statistics Service Population Estimates Data Base (<<http://www.nass.usda.gov:81/ipedb/>>). Horse population data were obtained from the FAOSTAT database (FAO 2001). Goat population data for 1992 and 1997 were obtained from the Census of Agriculture (USDA 1999d). Information regarding poultry turnover (i.e., slaughter) rate was obtained from state Natural Resource Conservation Service personnel (Lange 2000).

A summary of the livestock population characterization data used to calculate methane and nitrous oxide emissions is presented in Table L-1.

Dairy Cattle: The total annual dairy cow and heifer state population data for 1990 through 2000 are provided in various USDA National Agricultural Statistics Service reports (USDA 1995a, 1999a, 2000a-b, 2001a-b). The actual total annual dairy cow and heifer state population data used in the emissions calculations were downloaded from the U.S. Department of Agriculture National Agricultural Statistics Service Published Estimates Database (<<http://www.nass.usda.gov:81/ipedb/>>) for Cattle and Calves. The specific data used to estimate dairy cattle populations are “Cows That Calved – Milk” and “Heifers 500+ Lbs – Milk Repl.”

Beef Cattle: The total annual beef cattle population data for each state for 1990 through 2000 are provided in various USDA National Agricultural Statistics Service reports (USDA 1995a, 1999a, 2000a-b, 2001a-b). The actual data used in the emissions calculations were downloaded from the U.S. Department of Agriculture National Agricultural Statistics Service Published Estimates Database (<<http://www.nass.usda.gov:81/ipedb/>>), Cattle and Calves. The specific data used to estimate beef cattle populations are: “Cows That Calved—Beef,” “Heifers 500+ Lbs—Beef Repl,” “Heifers 500+ Lbs—Other,” and “Steers 500+ Lbs.” Additional information regarding the percent of beef steer and heifers on feedlots was obtained from contacts with the national USDA office (Milton 2000).

For all beef cattle groups (cows, heifers, steer, bulls, and calves), the USDA data provide cattle inventories from January and July of each year. Cattle inventory changes over the course of the year, sometimes significantly, as new calves are born and as fattened cattle are slaughtered; therefore, to develop the best estimate for the annual animal population, the average inventory of cattle by state was calculated. USDA provides January inventory data for each state; however, July inventory data is only presented as a total for the United States. In order to estimate average annual populations by state, a “scaling factor” was developed that adjusts the January state-level data to reflect July inventory changes. This factor equals the average of the US January and July data divided by the January data. The scaling factor is derived for each cattle group and is then applied to the January state-level data to arrive at the state-level annual population estimates.

Swine: The total annual swine population data for each state for 1990 through 2000 are provided in various USDA National Agricultural Statistics Service reports (USDA 1994a, 1998a, 2000c, 2001c). The USDA data provides

quarterly data for each swine subcategory: breeding, market under 60 pounds (less than 27 kg), market 60 to 119 pounds (27 to 54 kg), market 120 to 179 pounds (54 to 81 kg), and market 180 pounds and over (greater than 82 kg). The average of the quarterly data was used in the emissions calculations. For states where only December inventory is reported, the December data were used directly. The actual data used in the emissions calculations were downloaded from the U.S. Department of Agriculture National Agricultural Statistics Service Published Estimates Database (<<http://www.nass.usda.gov:81/ipedb/>>), Hogs and Pigs.

Sheep: The total annual sheep population data for each state for 1990 through 2000 were obtained from USDA National Agricultural Statistics Service (USDA 1994b, 1999c, 2000f, 2001f). Population data for lamb and sheep on feed are not available after 1993. The number of lamb and sheep on feed for 1994 through 2000 were calculated using the average of the percent of lamb and sheep on feed from 1990 through 1993. In addition, all of the sheep and lamb “on feed” are not necessarily on “feedlots”; they may be on pasture/crop residue supplemented by feed. Data for those animals on feed that are on feedlots versus pasture/crop residue were provided only for lamb in 1993. To calculate the populations of sheep and lamb on feedlots for all years, it was assumed that the percentage of sheep and lamb on feed that are on feedlots versus pasture/crop residue is the same as that for lambs in 1993 (Anderson 2000).

Goats: Annual goat population data by state were available for only 1992 and 1997 (USDA 1999d). The data for 1992 were used for 1990 through 1992 and the data for 1997 were used for 1997 through 2000. Data for 1993 through 1996 were extrapolated using the 1992 and 1997 data.

Poultry: Annual poultry population data by state for the various animal categories (hens 1 year and older, pullets of laying age, pullets 3 months old and older not of laying age, pullets under 3 months of age, other chickens, broilers, and turkeys) were obtained from USDA National Agricultural Statistics Service (USDA 1995b, 1998b, 1999b, 2000d-e, 2000g, 2001d-e). The annual population data for boilers and turkeys were adjusted for turnover (i.e., slaughter) rate (Lange 2000).

Horses: The Food and Agriculture Organization (FAO) publishes annual horse population data, which were accessed from the FAOSTAT database at <<http://apps.fao.org/>> (FAO 2001).

Step 2: Waste Characteristics Data

Methane and nitrous oxide emissions calculations are based on the following animal characteristics for each relevant livestock population:

- Volatile solids excretion rate (VS)
- Maximum methane producing capacity (B_0) for U.S. animal waste
- Nitrogen excretion rate (N_{ex})
- Typical animal mass (TAM)
- Annual state-specific milk production rate

Published sources were reviewed for U.S.-specific livestock waste characterization data that would be consistent with the animal population data discussed in Step 1. Data from the National Engineering Handbook, Agricultural Waste Management Field Handbook (USDA 1996a) were chosen as the primary source of waste characteristics. In some cases, data from the American Society of Agricultural Engineers, Standard D384.1 (ASAE 1999) were used to supplement the USDA data. The volatile solids and nitrogen excretion data for breeding swine are a combination of the types of animals that make up this animal group, namely gestating and farrowing swine and boars. It is assumed that a group of breeding swine is typically broken out as 80 percent gestating sows, 15 percent farrowing swine, and 5 percent boars (Safley 2000). Table L-2 presents a summary of the waste characteristics used in the emissions estimates.

The method for calculating volatile solids production from dairy cows is based on the relationship between milk production and volatile solids production. Cows that produce more milk per year also produce more volatile solids in their manure due to their increased feed. Figure 4-1 in the *Agricultural Waste Management Field Handbook*

(USDA 1996a) was used to determine the mathematical relationship between volatile solids production and milk production for a 1,400-pound dairy cow. The resulting best fit equation is logarithmic, shown in Figure L-1.

Figure L-1: Volatile Solids Production

Annual milk production data, published by USDA's National Agricultural Statistics Service (USDA 2001g), was accessed for each state and for each year of the inventory. State-specific volatile solids production rates were then calculated for each year of the inventory and used instead of a single national volatile solids excretion rate constant. Table L-3 presents the volatile solids production rates used for 2000.

Step 3: Waste Management System Usage Data

Estimates were made of the distribution of wastes by management system and animal type using the following sources of information:

- State contacts to estimate the breakout of dairy cows on pasture, range, or paddock, and the percent of wastes managed by daily spread systems (Deal 2000, Johnson 2000, Miller 2000, Stettler 2000, Sweeten 2000, Wright 2000)
- Data collected for EPA's Office of Water, including site visits, to medium and large beef feedlot, dairy, swine, and poultry operations (EPA 2001a)
- Contacts with the national USDA office to estimate the percent of beef steer and heifers on feedlots (Milton 2000)
- Survey data collected by USDA (USDA 1998d, 2000h) and re-aggregated by farm size and geographic location, used for small operations
- Survey data collected by the United Egg Producers (UEP 1999) and USDA (2000i) and previous EPA estimates (EPA 1992) of waste distribution for layers
- Survey data collected by Cornell University on dairy manure management operations in New York (Poe 1999)
- Previous EPA estimates of waste distribution for sheep, goat, and horse operations (EPA 1992)

Beef Feedlots: Based on EPA site visits and state contacts, beef feedlot manure is almost exclusively managed in drylots. Therefore, 100 percent of the manure excreted at beef feedlots is expected to be deposited in drylots and generate emissions. In addition, a portion of the manure that is deposited in the drylot will run off the drylot during rain events and be captured in a waste storage pond. An estimate of the runoff has been made by EPA's Office of Water for various geographic regions of the United States. These runoff numbers were used to estimate emissions from runoff storage ponds located at beef feedlots (EPA 2001a).

Dairy Cows: Based on EPA site visits and state contacts, manure from dairy cows at medium (200 through 700 head) and large (greater than 700 head) operations are managed using either flush systems or scrape/slurry systems. In addition, they may have a solids separator in place prior to their storage component. Estimates of the percent of farms that use each type of system (by geographic region) were developed by EPA's Office of Water, and were used to estimate the percent of wastes managed in lagoons (flush systems), liquid/slurry systems (scrape systems), and solid storage (separated solids). (EPA 2001a). Manure management system data for small (fewer than 200 head) dairies were obtained from USDA (USDA 2000h). These operations are more likely to use liquid/slurry and solid storage management systems than anaerobic lagoon systems. The reported manure management systems were deep pit, liquid/slurry (also includes slurry tank, slurry earth-basin, and aerated lagoon), anaerobic lagoon, and solid storage (also includes manure pack, outside storage, and inside storage).

The percent of wastes by system was estimated using the USDA data broken out by geographic region and farm size. Farm-size distribution data reported in the 1992 and 1997 Census of Agriculture (USDA 1999e) were used to determine the percentage of all dairies using the various manure management systems. Due to lack of additional data for other years, it was assumed that the data provided for 1992 were the same as that for 1990 and 1991, and data provided for 1997 were the same as that for 1998, 1999, and 2000. Data for 1993 through 1996 were extrapolated using the 1992 and 1997 data.

Data regarding the use of daily spread and pasture, range, or paddock systems for dairy cattle were obtained from personal communications with personnel from several organizations. These organizations include state NRCS offices, state extension services, state universities, USDA National Agricultural Statistics Service (NASS), and other experts (Deal 2000, Johnson 2000, Miller 2000, Stettler 2000, Sweeten 2000, and Wright 2000). Contacts at Cornell University provided survey data on dairy manure management practices in New York (Poe 1999). Census of Agriculture population data for 1992 and 1997 (USDA 1999e) were used in conjunction with the state data obtained from personal communications to determine regional percentages of total dairy cattle and dairy wastes that are managed using these systems. These percentages were applied to the total annual dairy cow and heifer state population data for 1990 through 2000, which were obtained from the USDA National Agricultural Statistics Service (USDA 1995a, 1999a, 2000a-b, 2001a-b).

Of the dairies using systems other than daily spread and pasture, range, or paddock systems, some dairies reported using more than one type of manure management system. Therefore, the total percent of systems reported by USDA for a region and farm size is greater than 100 percent. Typically, this means that some of the manure at a dairy is handled in one system (e.g., a lagoon), and some of the manure is handled in another system (e.g., drylot). However, it is unlikely that the same manure is moved from one system to another. Therefore, to avoid double counting emissions, the reported percentages of systems in use were adjusted to equal a total of 100%, using the same distribution of systems. For example, if USDA reported that 65 percent of dairies use deep pits to manage manure and 55 percent of dairies use anaerobic lagoons to manage manure, it was assumed that 54 percent (i.e., 65 percent divided by 120 percent) of the manure is managed with deep pits and 46 percent (i.e., 55 percent divided by 120 percent) of the manure is managed with anaerobic lagoons (ERG 2000).

Dairy Heifers: The percent of dairy heifer operations that are pasture, range, or paddock or that operate as daily spread was estimated using the same approach as dairy cows. Similar to beef cattle, dairy heifers are housed on drylots when not pasture based. Based on data from EPA's Office of Water (EPA 2001a), it was assumed that 100% of the manure excreted by dairy heifers is deposited in drylots and generates emissions. Estimates of runoff have been made by EPA's Office of Water for various geographic regions of the US (EPA 2001a).

Swine: Based on data collected during site visits for EPA's Office of Water (ERG 2000), manure from swine at large (greater than 2000 head) and medium (200 through 2000 head) operations are primarily managed using deep pit systems, liquid/slurry systems, or anaerobic lagoons. Manure management system data were obtained from USDA (USDA 1998d). It was assumed those operations with less than 200 head use pasture, range, or paddock systems. The percent of waste by system was estimated using the USDA data broken out by geographic region and farm size. Farm-size distribution data reported in the 1992 and 1997 Census of Agriculture (USDA 1999e) were used to determine the percentage of all swine utilizing the various manure management systems. The reported manure management systems were deep pit, liquid/slurry (also includes above- and below-ground slurry), anaerobic lagoon, and solid storage (also includes solids separated from liquids).

Some swine operations reported using more than one management system; therefore, the total percent of systems reported by USDA for a region and farm size is greater than 100 percent. Typically, this means that some of the manure at a swine operation is handled in one system (e.g., liquid system), and some of the manure is handled in another system (e.g., dry system). However, it is unlikely that the same manure is moved from one system to another. Therefore, to avoid double counting emissions, the reported percentages of systems in use were adjusted to equal a total of 100 percent, using the same distribution of systems, as explained under "Dairy Cows".

Sheep: It was assumed that all sheep wastes not deposited on feedlots were deposited on pasture, range, or paddock lands (Anderson 2000).

Goats/Horses: Estimates of manure management distribution were obtained from EPA's previous estimates (EPA 1992).

Poultry – Layers: Waste management system data for layers for 1990 were obtained from Appendix H of *Global Methane Emissions from Livestock and Poultry Manure* (EPA 1992). The percentage of layer operations using a shallow pit flush house with anaerobic lagoon or high-rise house without bedding was obtained for 1999 from United

Egg Producers, voluntary survey, 1999 (UEP 1999). These data were augmented for key poultry states (AL, AR, CA, FL, GA, IA, IN, MN, MO, NC, NE, OH, PA, TX, and WA) with USDA data (USDA 2000i). It was assumed that the change in system usage between 1990 and 1999 is proportionally distributed among those years of the inventory. It was assumed that system usage in 2000 was equal to that estimated for 1999. It was also assumed that 1 percent of poultry wastes are deposited on pasture, range, or paddock lands (EPA 1992).

Poultry - Broilers/Turkeys: The percentage of turkeys and broilers on pasture or in high-rise houses without bedding was obtained from *Global Methane Emissions from Livestock and Poultry Manure* (EPA1992). It was assumed that 1 percent of poultry wastes are deposited in pastures, range, and paddocks (EPA 1992).

Step 4: Emission Factor Calculations

Methane conversion factors (MCFs) and nitrous oxide emission factors (EFs) used in the emission calculations were determined using the methodologies shown below:

Methane Conversion Factors (MCFs)

Good Practice Guidance and Uncertainty Management in National Greenhouse Gas Inventories (IPCC 2000) for anaerobic lagoon systems published default methane conversion factors of 0 to 100 percent, which reflects the wide range in performance that may be achieved with these systems. There exist relatively few data points on which to determine country-specific MCFs for these systems. Therefore, a climate-based approach was identified to estimate MCFs for anaerobic lagoon and other liquid storage systems.

The following approach was used to develop the MCFs for liquid systems, and is based on the van't Hoff-Arrhenius equation used to forecast performance of biological reactions. One practical way of estimating MCFs for liquid manure handling systems is based on the mean ambient temperature and the van't Hoff-Arrhenius equation with a base temperature of 30°C, as shown in the following equation (Safley and Westerman 1990):

$$f = \exp \left[\frac{E(T_2 - T_1)}{RT_1T_2} \right]$$

where,

$T_1 = 303.16\text{K}$

$T_2 =$ ambient temperature (K) for climate zone (in this case, a weighted value for each state)

$E =$ activation energy constant (15,175 cal/mol)

$R =$ ideal gas constant (1.987 cal/K mol)

The factor “f” represents the proportion of volatile solids that are biologically available for conversion to methane based on the temperature of the system. The temperature is assumed equal to the ambient temperature. For colder climates, a minimum temperature of 5°C was established for uncovered anaerobic lagoons and 7.5°C for other liquid manure handling systems. For those animal populations using liquid systems (i.e., dairy cow, dairy heifer, layers, beef on feedlots, and swine) monthly average state temperatures were based on the counties where the specific animal population resides (i.e., the temperatures were weighted based on the percent of animals located in each county). The average county and state temperature data were obtained from the National Climate Data Center (NOAA 2001), and the county population data were based on 1992 and 1997 Census data (USDA 1999e). County population data for 1990 and 1991 were assumed to be the same as 1992; county population data for 1998 through 2000 were assumed to be the same as 1997; and county population data for 1993 through 1996 were extrapolated based on 1992 and 1997 data.

Annual MCFs for liquid systems are calculated as follows for each animal type, state, and year of the inventory:

- 1) Monthly temperatures are calculated by using county-level temperature and population data. The weighted-average temperature for a state is calculated using the population estimates and average monthly temperature in each county.
- 2) Monthly temperatures are used to calculate a monthly van't Hoff-Arrhenius “f” factor, using the equation presented above. A minimum temperature of 5°C is used for anaerobic lagoons and 7.5°C is used for liquid/slurry and deep pit systems.
- 3) Monthly production of volatile solids that are added to the system is estimated based on the number of animals present and, for lagoon systems, adjusted for a management and design practices factor. This factor accounts for other mechanisms by which volatile solids are removed from the management system prior to conversion to methane, such as solids being removed from the system for application to cropland. This factor, equal to 0.8, has been estimated using currently available methane measurement data from anaerobic lagoon systems in the United States (ERG 2001).
- 4) The amount of volatile solids available for conversion to methane is assumed to be equal to the amount of volatile solids produced during the month (from Step 3). For anaerobic lagoons, the amount of volatile solids available also includes volatile solids that may remain in the system from previous months.
- 5) The amount of volatile solids consumed during the month is equal to the amount available for conversion multiplied by the “f” factor.
- 6) For anaerobic lagoons, the amount of volatile solids carried over from one month to the next is equal to the amount available for conversion minus the amount consumed.
- 7) The estimated amount of methane generated during the month is equal to the monthly volatile solids consumed multiplied by the maximum methane potential of the waste (B_0).
- 8) The annual MCF is then calculated as:

$$\text{MCF (annual)} = \text{CH}_4 \text{ generated (annual)} / (\text{VS generated (annual)} * B_0)$$

In order to account for the carry over of volatile solids from the year prior to the inventory year for which estimates are calculated, it is assumed in the MCF calculation for lagoons that a portion of the volatile solids from October, November, and December of the year prior to the inventory year are available in the lagoon system starting January of the inventory year.

Following this procedure, the resulting MCF accounts for temperature variation throughout the year, residual volatile solids in a system (carryover), and management and design practices that may reduce the volatile solids available for conversion to methane. The methane conversion factors presented in Table L-4 by state and waste management system represent the average MCF for 2000 by state for all animal groups located in that state. However, in the calculation of methane emissions, specific MCFs for each animal type in the state are used.

Nitrous Oxide Emission Factors

Nitrous oxide emission factors for all manure management systems were set equal to the default IPCC factors (IPCC 2000).

Step 5: Weighted Emission Factors

For beef cattle, dairy cattle, swine, and poultry, the emission factors for both methane and nitrous oxide were weighted to incorporate the distribution of wastes by management system for each state. The following equation was used to determine the weighted MCF for a particular animal type in a particular state:

$$MCF_{animal, state} = \sum_{system} (MCF_{system, state} \times \% Manure_{animal, system, state})$$

where:

$MCF_{animal, state}$ = Weighted MCF for that animal group and state

$MCF_{system, state}$ = MCF for that system and state (see Step 4)

$\% Manure_{animal, system, state}$ = Percent of manure managed in the system for that animal group in that state (expressed as a decimal)

The weighted nitrous oxide emission factor for a particular animal type in a particular state was determined as follows:

$$EF_{animal, state} = \sum_{system} (EF_{system} \times \% Manure_{animal, system, state})$$

where,

$EF_{animal, state}$ = Weighted emission factor for that animal group and state

EF_{system} = Emission factor for that system (see Step 4)

$\% Manure_{animal, system, state}$ = Percent of manure managed in the system for that animal group in that state (expressed as a decimal)

Data for the calculated weighted factors for 1992 came from the 1992 Census of Agriculture, combined with assumptions on manure management system usage based on farm size, and were also used for 1990 and 1991. Data for the calculated weighted factors for 1997 came from the 1997 Census of Agriculture, combined with assumptions on manure management system usage based on farm size, and were also used for 1998, 1999, and 2000. Factors for 1993 through 1996 were calculated by interpolating between the two sets of factors. A summary of the weighted MCFs used to calculate beef feedlot, dairy cow and heifer, swine, and poultry emissions for 2000 are presented in Table L-5.

Step 6: Methane and Nitrous Oxide Emission Calculations

For beef feedlot cattle, dairy cows, dairy heifers, swine, and poultry, methane emissions were calculated for each animal group as follows:

$$Methane_{animal\ group} = \sum_{state} (Population \times VS \times B_o \times MCF_{animal, state} \times 0.662)$$

where:

$Methane_{animal\ group}$ = methane emissions for that animal group (kg CH₄/yr)

Population = annual average state animal population for that animal group (head)

VS = total volatile solids produced annually per animal (kg/yr/head)

B_o = maximum methane producing capacity per kilogram of VS (m³ CH₄/kg VS)

$MCF_{animal, state}$ = weighted MCF for the animal group and state (see Step 5)

0.662 = conversion factor of m³ CH₄ to kilograms CH₄ (kg CH₄/m³ CH₄)

Methane emissions from other animals (i.e., sheep, goats, and horses) were based on the 1990 methane emissions estimated using the detailed method described in *Anthropogenic Methane Emissions in the United States: Estimates for 1990, Report to Congress* (EPA 1993). This approach is based on animal-specific manure characteristics and management system data. This process was not repeated for subsequent years for these other animal types. Instead,

national populations of each of the animal types were used to scale the 1990 emissions estimates to the period 1991 through 2000.

Nitrous oxide emissions were calculated for each animal group as follows:

$$\text{Nitrous Oxide}_{\text{animal group}} = \sum_{\text{state}} (\text{Population} \times N_{\text{ex}} \times EF_{\text{animal, state}} \times 44 / 28)$$

where:

Nitrous Oxide_{animal group} = nitrous oxide emissions for that animal group (kg/yr)

Population = annual average state animal population for that animal group (head)

N_{ex} = total Kjeldahl nitrogen excreted annually per animal (kg/yr/head)

EF_{animal, state} = weighted nitrous oxide emission factor for the animal group and state, kg N₂O-N/kg N excreted (see Step 5)

44/28 = conversion factor of N₂O-N to N₂O

Emission estimates are summarized in Table L-6 and Table L-7.

Table L-3: Estimated Dairy Cow Volatile Solids Production Rate By State for 2000

State	Volatile Solids ¹ (kg/day/1000 kg)
Alabama	7.07
Alaska	7.28
Arizona	9.32
Arkansas	6.52
California	9.20
Colorado	9.30
Connecticut	8.48
Delaware	7.42
Florida	7.67
Georgia	7.93
Hawaii	7.23
Idaho	9.11
Illinois	8.22
Indiana	7.88
Iowa	8.46
Kansas	8.00
Kentucky	6.65
Louisiana	6.38
Maine	8.00
Maryland	7.80
Massachusetts	8.10
Michigan	8.65
Minnesota	8.31
Mississippi	7.49
Missouri	7.34
Montana	8.23
Nebraska	7.86
Nevada	8.66
New Hampshire	8.15
New Jersey	7.97
New Mexico	9.14
New York	8.20
North Carolina	8.01
North Dakota	7.21
Ohio	8.09
Oklahoma	7.18
Oregon	8.60
Pennsylvania	8.40
Rhode Island	7.67
South Carolina	7.79
South Dakota	7.78
Tennessee	7.38
Texas	7.93
Utah	8.25

¹ Volatile solids production estimates based on state average annual milk production rates, combined with a mathematical relationship of volatile solids to milk production (USDA 1996a).

Vermont	8.24
Virginia	7.73
Washington	9.54
West Virginia	7.65
Wisconsin	8.18
Wyoming	6.94

Table L-4: Methane Conversion Factors By State for Liquid Systems² for 2000

State	Liquid/Slurry and Deep Pit	Anaerobic Lagoon
Alabama	0.4122	0.7538
Alaska	0.1472	0.4677
Arizona	0.4919	0.7689
Arkansas	0.3823	0.7536
California	0.3440	0.7330
Colorado	0.2336	0.6705
Connecticut	0.2337	0.6642
Delaware	0.2927	0.7124
Florida	0.5193	0.7684
Georgia	0.3919	0.7411
Hawaii	0.5827	0.7869
Idaho	0.2247	0.6570
Illinois	0.2870	0.7128
Indiana	0.2714	0.6976
Iowa	0.2627	0.6981
Kansas	0.3439	0.7515
Kentucky	0.3151	0.7241
Louisiana	0.4790	0.7631
Maine	0.1917	0.6025
Maryland	0.2786	0.6999
Massachusetts	0.2243	0.6523
Michigan	0.2295	0.6576
Minnesota	0.2335	0.6675
Mississippi	0.4308	0.7584
Missouri	0.3245	0.7361
Montana	0.2073	0.6337
Nebraska	0.2856	0.7197
Nevada	0.2466	0.6787
New Hampshire	0.2007	0.6176
New Jersey	0.2605	0.6896
New Mexico	0.3272	0.7328
New York	0.2167	0.6402
North Carolina	0.3346	0.7255
North Dakota	0.2165	0.6482
Ohio	0.2573	0.6841
Oklahoma	0.3933	0.7602
Oregon	0.2112	0.6291
Pennsylvania	0.2485	0.6764
Rhode Island	0.2420	0.6765
South Carolina	0.3831	0.7401

² As defined by IPCC (IPCC, 2000).

South Dakota	0.2496	0.6911
Tennessee	0.3390	0.7367
Texas	0.4622	0.7613
Utah	0.2673	0.7029
Vermont	0.1965	0.6090
Virginia	0.2829	0.7009
Washington	0.2126	0.6329
West Virginia	0.2607	0.6850
Wisconsin	0.2278	0.6595
Wyoming	0.2184	0.6513

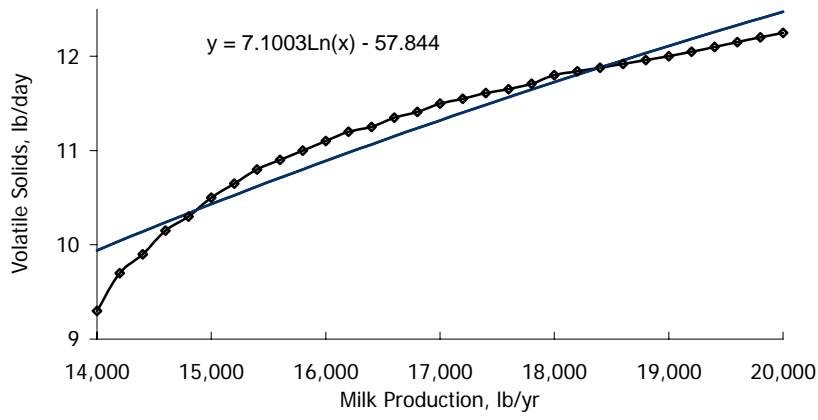


Figure L-1: Volatile Solids Production

ANNEX M: Methodology for Estimating N₂O Emissions from Agricultural Soil Management

Nitrous oxide (N₂O) emissions from agricultural soil management covers activities that add nitrogen (N) to soils, and thereby enhance natural emissions of N₂O. The IPCC methodology (IPCC/UNEP/OECD/IEA 1997, IPCC 2000), which is used here, divides this source category into three components: (1) direct N₂O emissions from managed soils; (2) direct N₂O emissions from pasture, range, and paddock livestock manure; and (3) indirect N₂O emissions from soils induced by applications of nitrogen.

There are five steps in estimating N₂O emissions from agricultural soil management. First, the activity data are derived for each of the three components. Note that some of the data used in the first component are also used in the third component. In the second, third, and fourth steps, N₂O emissions from each of the three components are estimated. In the fifth step, emissions from the three components are summed to estimate total emissions. The remainder of this annex describes these steps, and data used in these steps, in detail.

Step 1: Derive Activity Data

The activity data for this source category are annual amounts of nitrogen added to soils for each relevant activity, except for histosol cultivation, for which the activity data are annual histosol areas cultivated.¹ The activity data are derived from statistics, such as fertilizer consumption data or livestock population data, and various factors used to convert these statistics to annual amounts of nitrogen, such as fertilizer nitrogen contents or livestock excretion rates. Activity data were derived for each of the three components, as described below.

Step 1a. Direct N₂O Emissions from Managed Soils.

The activity data for this component include: a) the amount of nitrogen in synthetic and organic commercial fertilizers that are applied annually, b) the amount of nitrogen in livestock manure that is applied annually through both daily spread operations and the eventual application of manure that had been stored in manure management systems, c) the amount of nitrogen in sewage sludge that is applied annually, d) the amount of nitrogen in the aboveground biomass of nitrogen-fixing crops that are produced annually, e) the amount of nitrogen in crop residues that are retained on soils annually, and f) the area of histosols cultivated annually.

Application of synthetic and organic commercial fertilizer: Annual commercial fertilizer consumption data for the United States were taken from annual publications of synthetic and organic fertilizer statistics (TVA 1991, 1992a, 1993, 1994; AAPFCO 1995, 1996, 1997, 1998, 1999, 2000b) and a recent AAPFCO database (AAPFCO 2000a). These data were manipulated in several ways to derive the activity data needed for the inventory. First, the manure and sewage sludge portions of the organic fertilizers were subtracted from the total organic fertilizer consumption data because these nitrogen additions are accounted for under “manure application” and “sewage sludge application.”² Second, the organic fertilizer data, which are recorded in mass units of fertilizer, had to be converted to mass units of nitrogen by multiplying by the average organic fertilizer nitrogen contents provided in the annual fertilizer publications. These nitrogen contents are weighted average values, so they vary from year-to-year (ranging from 2.3 percent to 3.9 percent over the period 1990 through 2000). The synthetic fertilizer data are recorded in units of nitrogen, so these data did not need to be converted. Lastly, both the synthetic and organic fertilizer consumption data are recorded in “fertilizer year” totals (i.e., July to June); therefore, the data were converted to

¹ Histosols are soils with a high organic carbon content. All soils with more than 20 to 30 percent organic matter by weight (depending on the clay content) are classified as histosols (Brady and Weil 1999).

² Organic fertilizers included in these publications are manure, compost, dried blood, sewage sludge, tankage, and “other.” (Tankage is dried animal residue, usually freed from fat and gelatin). The manure and sewage sludge used as commercial fertilizer are accounted for elsewhere, so these were subtracted from the organic fertilizer statistics to avoid double counting.

calendar year totals. This was done by assuming that approximately 35 percent of fertilizer usage occurred from July to December, and 65 percent from January to June (TVA 1992b). July to December values were not available for calendar year 2000, so a “least squares line” statistical extrapolation using the previous ten years of data was used to arrive at an approximate value. Annual consumption of commercial fertilizers—synthetic and non-manure/non-sewage organic—in units of nitrogen and on a calendar year basis are presented in Table M-1.

Application of livestock manure: To estimate the amount of livestock manure nitrogen applied to soils, it was assumed that all of the manure produced by livestock would be applied to soils with two exceptions. These exceptions were: (1) the portion of poultry manure that is used as a feed supplement for ruminants, and (2) the manure that is deposited on soils by livestock on pasture, range, and paddock. In other words, all of the manure that is managed, except the portion of poultry manure that is used as a feed supplement, is assumed to be applied to soils. The amount of managed manure for each livestock type was calculated by determining the population of animals that were on feedlots or otherwise housed in order to collect and manage the manure. In some instances, the number of animals in managed systems was determined by subtracting the number of animals in pasture, range, and paddock from the total animal population for a particular animal type.

Annual animal population data for all livestock types, except horses and goats, were obtained from the USDA National Agricultural Statistics Service (USDA 1994b,c; 1995a,b; 1998a; 1998c; 1999a-c; 2000a-g; 2001b-g). Horse population data were obtained from the FAOSTAT database (FAO 2001). Goat population data were obtained from the Census of Agriculture (USDA 1999d). Information regarding poultry turnover (i.e., slaughter) rate was obtained from state Natural Resource Conservation Service personnel (Lange 2000). Additional population data for different farm size categories for dairy and swine were obtained from the Census of Agriculture (USDA 1999e).

Information regarding the percentage of manure handled using various manure management systems for dairy cattle, beef cattle, and sheep was obtained from communications with personnel from state Natural Resource Conservation Service offices, state universities, National Agricultural Statistics Service, and other experts (Poe et al. 1999, Anderson 2000, Deal 2000, Johnson 2000, Miller 2000, Milton 2000, Stettler 2000, Sweeten 2000, Wright 2000). Information regarding the percentage of manure handled using various manure management systems for swine, poultry, goats, and horses was obtained from Safley et al. (1992). A more detailed discussion of manure management system usage is provided in Annex L.

Once the animal populations for each livestock type and management system were estimated, these populations were then multiplied by an average animal mass constant (USDA 1996, USDA 1998d, ASAE 1999, Safley 2000) to derive total animal mass for each animal type in each management system. Total Kjeldahl nitrogen³ excreted per year for each livestock type and management system was then calculated using daily rates of nitrogen excretion per unit of animal mass (USDA 1996, ASAE 1999). The total poultry manure nitrogen in managed systems was reduced by the amount assumed to be used as a feed supplement (i.e., 4.2 percent of the managed poultry manure; Carpenter 1992). The annual amounts of Kjeldahl nitrogen were then summed over all livestock types and management systems to derive estimates of the annual manure nitrogen applied to soils (Table M-2).

Application of sewage sludge: Estimates of annual nitrogen additions from land application of sewage sludge were derived from periodic estimates of sludge generation and disposal rates that were developed by EPA. Sewage sludge is generated from the treatment of raw sewage in public or private wastewater treatment works. Based on a 1988 questionnaire returned from 600 publicly owned treatment works (POTWs), the EPA estimated that 5.4 million metric tons of dry sewage sludge were generated by POTWs in the United States in that year (EPA 1993). Of this total, 43.7 percent was applied to land, including agricultural applications, compost manufacture, forest land application, the reclamation of mining areas, and other forms of surface disposal. A subsequent EPA report (EPA, 1999) compiled data from several national studies and surveys, and estimated that approximately 6.7 and 6.9 million

³ Total Kjeldahl nitrogen is a measure of organically bound nitrogen and ammonia nitrogen in both the solid and liquid wastes.

metric tons of dry sewage sludge were generated in 1996 and 1998, respectively, and projected that approximately 7.1 million metric tons would be generated in 2000. The same study concluded that 60 percent of the sewage sludge generated in 1998 was applied to land (based on the results of a 1995 survey), and projected that 63 percent would be land applied in 2000. These EPA estimates of sludge generation and percent land applied were linearly interpolated to derive estimates for each year in the 1990 to 2000 period. To estimate annual amounts of nitrogen applied, the annual amounts of dry sewage sludge applied were multiplied by an average nitrogen content of 3.3 percent (Metcalf and Eddy, Inc. 1991). Final estimates of annual amounts of sewage sludge nitrogen applied to land are presented in Table M-1.

Production of nitrogen-fixing crops: Annual production statistics for beans, pulses, and alfalfa were taken from U.S. Department of Agriculture crop production reports (USDA 1994a, 1998b, 2000i, 2001a). Annual production statistics for the remaining nitrogen-fixing crops (i.e., the major non-alfalfa forage crops, specifically red clover, white clover, birdsfoot trefoil, arrowleaf clover, and crimson clover) were derived from information in a book on forage crops (Taylor and Smith 1995, Pederson 1995, Beuselinck and Grant 1995, Hoveland and Evers 1995), and personal communications with forage experts (Cropper 2000, Evers 2000, Gerrish 2000, Hoveland 2000, and Pederson 2000).

The production statistics for beans, pulses, and alfalfa were in tons of product, which needed to be converted to tons of aboveground biomass nitrogen. This was done by multiplying the production statistics by one plus the aboveground residue to crop product mass ratios, dry matter fractions, and nitrogen contents. The residue to crop product mass ratios for soybeans and peanuts, and the dry matter content for soybeans, were obtained from Strehler and Stützel (1987). The dry matter content for peanuts was obtained through personal communications with Ketzis (1999). The residue to crop product ratios and dry matter contents for the other beans and pulses were estimated by taking averages of the values for soybeans and peanuts. The dry matter content for alfalfa was obtained through personal communications with Karkosh (2000). The IPCC default nitrogen content of 3 percent (IPCC/UNEP/OECD/IEA 1997) was used for all beans, pulses, and alfalfa.⁴

The production statistics for the non-alfalfa forage crops were derived by multiplying estimates of areas planted by estimates of annual yields, in dry matter mass units. These derived production statistics were then converted to units of nitrogen by applying the IPCC default nitrogen content of 3 percent (IPCC/UNEP/OECD/IEA 1997).

The final estimates of annual aboveground biomass production, in units of nitrogen, are presented in Table M-3. The residue to crop product mass ratios and dry matter fractions used in these calculations are presented in Table M-6.

Retention of crop residue: It was assumed that 90 percent of residues from corn, wheat, barley, sorghum, oats, rye, millet, soybeans, peanuts, and other beans and pulses are left on the field after harvest (e.g., rolled into the soil, chopped and disked into the soil, or otherwise left behind) (Karkosh 2000).⁵ It was also assumed that 100 percent of unburned rice residue is left on the field.⁶

The derivation of residue nitrogen activity data was very similar to the derivation of nitrogen-fixing crop activity data. Crop production statistics were multiplied by aboveground residue to crop product mass ratios, residue dry

⁴ This nitrogen content may be an overestimate for the residue portion of the aboveground biomass of the beans and pulses. Also, the dry matter fractions used for beans and pulses were taken from literature on crop residues, and so may be underestimates for the product portion of the aboveground biomass.

⁵ Although the mode of residue application would likely affect the magnitude of N₂O emissions, an emission estimation methodology that accounts for this has not been developed.

⁶ Some of the rice residue may be used for other purposes, such as for biofuel or livestock bedding material. Research to obtain more detailed information regarding final disposition of rice residue, as well as the residue of other crops, will be undertaken for future inventories.

matter fractions, residue nitrogen contents, and the fraction of residues left on soils. Annual production statistics for all crops except rice in Florida were taken from U.S. Department of Agriculture reports (USDA 1994a, 1998b, 2000i, 2001a). Production statistics for rice in Florida, which are not recorded by USDA, were estimated by applying an average rice crop yield for Florida (Smith 2001) to annual Florida rice acreages (Schueneman 1999, 2001). Residue to crop product ratios for all crops were obtained from, or derived from, Strehler and Stützle (1987). Dry matter contents for wheat, rice, corn, and barley residue were obtained from Turn et al. (1997). Soybean and millet residue dry matter contents were obtained from Strehler and Stützle (1987). Peanut, sorghum, oat, and rye residue dry matter contents were obtained through personal communications with Ketzis (1999). Dry matter contents for all other beans and pulses were estimated by averaging the values for soybeans and peanuts. The residue nitrogen contents for wheat, rice, corn, and barley are from Turn et al. (1997). The nitrogen content of soybean residue is from Barnard and Kristoferson (1985), the nitrogen contents of peanut, sorghum, oat, and rye residue are from Ketzis (1999), and the nitrogen content of millet residue is from Strehler and Stützle (1987). Nitrogen contents of all other beans and pulses were estimated by averaging the values for soybeans and peanuts. Estimates of the amounts of rice residue burned annually were derived using information obtained from agricultural extension agents in each of the rice-growing states (see Agricultural Residue Burning section of Agriculture Chapter for more detail).

The final estimates of residue retained on soil, in units of nitrogen (N), are presented in Table M-4. The residue to crop product mass ratios, residue dry matter fractions, and residue nitrogen contents used in these calculations are presented in Table M-6.

Cultivation of histosols: Estimates of the areas of histosols cultivated in 1982, 1992, and 1997 were obtained from the USDA's 1997 *National Resources Inventory* (USDA 2000h, as extracted by Eve 2001, and revised by Ogle 2002).⁷ These areas were grouped by broad climatic region⁸ using temperature and precipitation estimates from Daly et al. (1994, 1998), and then further aggregated to derive a temperate total and a sub-tropical total. These final areas were then linearly interpolated to obtain estimates for 1990 through 1996, and linearly extrapolated to obtain area estimates for 1998 through 2000 (Table M-5).

Step 1b. Direct N₂O Emissions from Pasture, Range, and Paddock Livestock Manure.

Estimates of N₂O emissions from this component were based on livestock manure that is not managed in manure management systems, but instead is deposited directly on soils by animals in pasture, range, and paddock. The livestock included in this component were: dairy cattle, beef cattle, swine, sheep, goats, poultry, and horses.

Dairy Cattle: Information regarding dairy farm grazing was obtained from communications with personnel from state Natural Resource Conservation Service offices, state universities, and other experts (Poe et al. 1999, Deal 2000, Johnson 2000, Miller 2000, Stettler 2000, Sweeten 2000, Wright 2000). Because grazing operations are typically related to the number of animals on a farm, farm-size distribution data reported in the 1992 and 1997 *Census of Agriculture* (USDA 1999e) were used in conjunction with the state data obtained from personal communications to determine the percentage of total dairy cattle that graze. An overall percent of dairy waste that is deposited in pasture, range, and paddock was developed for each region of the United States. This percentage was applied to the total annual dairy cow and heifer state population data for 1990 through 2000, which were obtained from the USDA National Agricultural Statistics Service (USDA 1995a; 1999a; 2000a,b; 2001b,c).

Beef Cattle: To determine the population of beef cattle that are on pasture, range, and paddock, the following assumptions were made: 1) beef cows, bulls, and calves were not housed on feedlots; 2) a portion of heifers and steers were on feedlots; and 3) all beef cattle that were not housed on feedlots were located on pasture, range, and

⁷ These areas do not include Alaska, but Alaska's cropland area accounts for less than 0.1 percent of total U.S. cropland area, so this omission is not significant.

⁸ These climatic regions were: 1) cold temperate, dry, 2) cold temperate, moist, 3) sub-tropical, dry, 4) sub-tropical, moist, 5) warm temperate, dry, and 6) warm temperate, moist.

paddock (i.e., total population minus population on feedlots equals population of pasture, range, and paddock) (Milton 2000). Information regarding the percentage of heifers and steers on feedlots was obtained from USDA personnel (Milton 2000) and used in conjunction with USDA National Agricultural Statistics Service population data (USDA 1995a; 1999a; 2000a,b; 2001b,c) to determine the population of steers and heifers on pasture, range, and paddock.

Swine: Based on the assumption that smaller facilities are less likely to utilize manure management systems, farm-size distribution data reported in the *1992 and 1997 Census of Agriculture* (USDA 1999e) were used to determine the percentage of all swine whose manure is not managed (i.e., the percentage on pasture, range, and paddock). These percentages were applied to the average of the quarterly USDA National Agricultural Statistics Service population data for swine (USDA 1994b, 1998a, 2000e, 2001d) to determine the population of swine on pasture, range, and paddock.

Sheep: It was assumed that all sheep and lamb manure not deposited on feedlots was deposited on pasture, range, and paddock (Anderson 2000). Sheep population data were obtained from the USDA National Agricultural Statistics Service (USDA 1994c, 1999c, 2000g, 2001f). However, population data for lamb and sheep on feed were not available after 1993. The number of lamb and sheep on feed for 1994 through 2000 were calculated using the average of the percent of lamb and sheep on feed from 1990 through 1993. In addition, all of the sheep and lamb “on feed” were not necessarily on “feedlots”; they may have been on pasture/crop residue supplemented by feed. Data for those feedlot animals versus pasture/crop residue were provided only for lamb in 1993. To calculate the populations of sheep and lamb on feedlots for all years, it was assumed that the percentage of sheep and lamb on feedlots versus pasture/crop residue is the same as that for lambs in 1993 (Anderson 2000).

Goats: It was assumed that 92 percent of goat manure was deposited on pasture, range, and paddock (Safley et al. 1992). Annual goat population data by state were available for only 1992 and 1997 (USDA 1999d). The data for 1992 were used for 1990 through 1992 and the data for 1997 were used for 1997 through 2000. Data for 1993 through 1996 were extrapolated using the 1992 and 1997 data.

Poultry: It was assumed that one percent of poultry manure was deposited on pasture, range, and paddock (Safley et al. 1992). Poultry population data were obtained from USDA National Agricultural Statistics Service (USDA 1995b, 1998a, 1999b, 2000c, 2000d, 2000f, 2001f). The annual population data for boilers and turkeys were adjusted for turnover (i.e., slaughter) rate (Lange 2000).

Horses: It was assumed that 92 percent of horse manure was deposited on pasture, range, and paddock (Safley et al. 1992). Horse population data were obtained from the FAOSTAT database (FAO 2001).

For each animal type, the population of animals within pasture, range, and paddock systems was multiplied by an average animal mass constant (USDA 1996, ASAE 1999, USDA 1998d, Safley 2000) to derive total animal mass for each animal type. Total Kjeldahl nitrogen excreted per year was then calculated for each animal type using daily rates of nitrogen excretion per unit of animal mass (USDA 1996, ASAE 1999). Annual nitrogen excretion was then summed over all animal types to yield total nitrogen in pasture, range, and paddock manure (Table M-2).

Step 1c. Indirect N₂O Emissions from Soils Induced by Applications of Nitrogen.

This component accounts for N₂O that is emitted indirectly from nitrogen applied as commercial fertilizer, sewage sludge, and livestock manure. Through volatilization, some of this nitrogen enters the atmosphere as NH₃ and NO_x, and subsequently returns to soils through atmospheric deposition, thereby enhancing N₂O production. Additional nitrogen is lost from soils through leaching and runoff, and enters groundwater and surface water systems, from which a portion is emitted as N₂O. These two indirect emission pathways are treated separately, although the activity data used, except for livestock manure, are identical. The activity data for commercial fertilizer and sewage sludge are the same as those used in the calculation of direct emissions from managed soils (Table M-1). The activity data for livestock manure are different from those used in other calculations. Here, total livestock manure (i.e., the sum of applied manure, manure in pasture, range, and paddock, and manure used as a livestock feed supplement) is used in the volatilization and deposition calculation; and livestock manure applied or deposited on soils (i.e., the sum of

+ Less than 0.5 Gg N.

Note: Totals may not sum due to independent rounding.

Table M-5: Cultivated Histosol Area (Thousand Hectares)

Year	Temperate Area	Sub-Tropical Area
1990	432	192
1991	431	193
1992	429	194
1993	431	194
1994	433	195
1995	435	195
1996	437	196
1997	439	196
1998	441	197
1999	443	197
2000	445	197

Table M-6: Key Assumptions for Nitrogen-Fixing Crop Production and Crop Residue

Crop	Residue/Crop Ratio	Residue Dry Matter Fraction	Residue Nitrogen Fraction
Soybeans	2.1	0.87	0.023
Peanuts	1.0	0.86	0.0106
Dry Edible Beans	1.55	0.87	0.0062
Dry Edible Peas	1.55	0.87	0.0062
Austrian Winter Peas	1.55	0.87	0.0062
Lentils	1.55	0.87	0.0062
Wrinkled Seed Peas	1.55	0.87	0.0062
Alfalfa	0	0.85	NA
Corn	1.0	0.91	0.0058
Wheat	1.3	0.93	0.0062
Barley	1.2	0.93	0.0077
Sorghum	1.4	0.91	0.0108
Oats	1.3	0.92	0.007
Rye	1.6	0.90	0.0048
Millet	1.4	0.89	0.007
Rice	1.4	0.91	0.0072

Note: For the derivation of activity data for nitrogen-fixing crop production, the IPCC default nitrogen content of aboveground biomass (3 percent) was used.

Step 2: Estimate Direct N₂O Emissions from Managed Soils Due to Nitrogen Additions and Cultivation of Histosols

In this step, N₂O emissions were calculated for each of two parts (direct N₂O emissions due to nitrogen additions and direct N₂O emissions due to histosol cultivation), which were then summed to yield total direct N₂O emissions from managed soils (Table M-7).

Step 2a. Direct N₂O Emissions Due to Nitrogen Additions.

To estimate these emissions, the amounts of nitrogen applied were each reduced by the IPCC default fraction of nitrogen that is assumed to volatilize, the unvolatilized amounts were then summed, and the total unvolatilized nitrogen was multiplied by the IPCC default emission factor of 0.0125 kg N₂O-N/kg Nitrogen (IPCC/UNEP/OECD/IEA 1997). The volatilization assumptions are described below.

- *Application of synthetic and organic commercial fertilizer:* The total amounts of nitrogen applied in the form of synthetic commercial fertilizers and non-manure/non-sewage organic commercial fertilizers were reduced by 10 percent and 20 percent, respectively, to account for the portion that volatilizes to NH₃ and NO_x (IPCC/UNEP/OECD/IEA 1997).
- *Application of livestock manure:* The total amount of livestock manure nitrogen applied to soils was reduced by 20 percent to account for the portion that volatilizes to NH₃ and NO_x (IPCC/UNEP/OECD/IEA 1997).
- *Application of sewage sludge:* The total amount of sewage sludge nitrogen applied to soils was reduced by 20 percent to account for the portion that volatilizes to NH₃ and NO_x (IPCC/UNEP/OECD/IEA 1997, IPCC 2000).
- *Production of nitrogen-fixing crops:* None of the nitrogen in the aboveground biomass of nitrogen-fixing crops was assumed to volatilize.
- *Retention of crop residue:* None of the nitrogen in retained crop residue was assumed to volatilize.

Step 2b. Direct N₂O Emissions Due to Cultivation of Histosols.

To estimate annual N₂O emissions from histosol cultivation, the temperate histosol area was multiplied by the IPCC default emission factor for temperate soils (8 kg N₂O-N/ha cultivated; IPCC 2000), and the sub-tropical histosol area was multiplied by the average of the temperate and tropical IPCC default emission factors (12 kg N₂O-N/ha cultivated; IPCC 2000).

Step 3: Estimate Direct N₂O Emissions from Pasture, Range, and Paddock Livestock Manure

To estimate direct N₂O emissions from soils due to the deposition of pasture, range, and paddock manure, the total nitrogen excreted by these animals was multiplied by the IPCC default emission factor (0.02 kg N₂O-N/kg N excreted) (see Table M-8).

Step 4: Estimate Indirect N₂O Emissions Induced by Applications of Nitrogen

In this step, N₂O emissions were calculated for each of two parts (indirect N₂O emissions due to volatilization of applied nitrogen and indirect N₂O emissions due to leaching and runoff of applied nitrogen), which were then summed to yield total direct N₂O emissions from managed soils.

Step 4a. Indirect Emissions Due to Volatilization.

To estimate these emissions, first the amounts of commercial fertilizer nitrogen and sewage sludge nitrogen applied, and the total amount of manure nitrogen produced, were each multiplied by the IPCC default fraction of nitrogen that is assumed to volatilize to NH₃ and NO_x (10 percent for synthetic fertilizer nitrogen; and 20 percent for nitrogen in organic fertilizer, sewage sludge, and livestock manure). Next, the volatilized amounts of nitrogen were summed, and then the total volatilized nitrogen was multiplied by the IPCC default emission factor of 0.01 kg N₂O-N/kg N (IPCC/UNEP/OECD/IEA 1997). These emission estimates are presented in (Table M-9).

Step 4b. Indirect Emissions Due to Leaching and Runoff.

To estimate these emissions, first the amounts of commercial fertilizer nitrogen and sewage sludge nitrogen applied, and the total amount of manure nitrogen applied or deposited, were each multiplied by the IPCC default fraction of nitrogen that is assumed to leach and runoff (30 percent for all nitrogen). Next, the leached/runoff amounts of nitrogen were summed, and then the total nitrogen was multiplied by the IPCC default emission factor of 0.025 kg N₂O-N/kg N (IPCC/UNEP/OECD/IEA 1997). These emission estimates are presented in (Table M-9).

Table M-7: Direct N₂O Emissions from Managed Soils (Tg CO₂ Eq.)

Activity	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000
Commercial Fertilizers*	55	56	57	59	61	59	61	61	61	62	62
Livestock Manure	13	13	13	13	13	13	13	13	14	14	14
Sewage Sludge	+	+	1	1	1	1	1	1	1	1	1
Nitrogen Fixation	58	60	61	57	66	62	64	68	69	68	69

ANNEX N: Methodology for Estimating Net Changes in Forest Carbon Stocks

This annex presents a discussion of the methodology used to calculate net changes in carbon stocks in trees, understory, forest floor, down dead wood, forest soils, and harvested wood (i.e., wood products and landfilled wood). More detailed discussions of selected topics may be found in the references cited in this annex.

The details of carbon conversion factors and step-by-step details of calculating net CO₂ flux for forests are given in three steps. In addition, the modeling projection system is briefly described.

Step 1. Estimate Forest Carbon Stocks and Net Changes in Forest Carbon Stocks

Step 1a. Obtain Forest Inventory Data

Forest survey data in the United States by broad forest type and region for 1987 and 1997 were obtained from U.S. Forest Service, Forest Inventory & Analysis (FIA) estimates of forest resources, published in Waddell et al. (1989) and Smith et al. (2001). The FIA data include: (1) growing stock volume per acre by forest type (referred to hereinafter as “growing stock volumes”); and (2) area by Timberland and other forest land, for general forest types by region (referred to hereinafter as “forest areas”). For 2001, the same variables were obtained from model results as described in Haynes et al. (2001b). (See The Forest Sector Modeling Projection System below). This information was combined with separate estimates of carbon density (carbon mass per unit area) to compile estimates of carbon stocks.

Step 1b. Estimate Carbon in Living and Standing Dead Trees

To estimate live tree biomass, equations that convert forest tree volumes to total live tree dry biomass (Smith et al. in review) were applied to the growing stock volumes by forest type and region (obtained in Step 1a). Tree biomass includes aboveground biomass and belowground biomass of coarse roots. The minimum size tree is one-inch diameter at diameter breast height (1.3 meter). Trees less than one-inch diameter are counted as carbon in understory vegetation. Biomass estimates were divided by two to obtain estimates of carbon in living trees (i.e., it was assumed that dry biomass is 50 percent carbon). Standing dead tree biomass was calculated by applying equations that estimate biomass for standing dead trees (Smith et al. in review) from growing stock volumes. Again, standing dead tree biomass was divided by two to estimate carbon in standing dead trees. Table N-1 lists the average living and standing dead tree carbon densities by forest type, as calculated by applying the equations to the 1997 data.

Table N-1: Average U.S. Carbon Densities of Forest Components* (metric tons C/ha)

Region^a/Forest Type	Live and Standing Dead Tree Carbon	Forest Floor Carbon	Soil Organic Carbon^b
Eastern			
White-red-jack pine	77.1	13.8	196.1
Spruce-fir	59.8	40.2	192.9
Longleaf-slash pine	42.4	9.2	136.3
Loblolly-shortleaf pine	49.3	9.1	91.7
Oak-pine	57.3	11.6	82.3
Oak-hickory	76.3	6.6	85.0
Oak-gum-cypress	86.0	6.0	152.2
Elm-ash-cottonwood	67.6	23.0	118.1
Maple-beech-birch	82.5	28.0	139.5
Aspen-birch	56.0	7.6	237.0
Other forest types	1.8	2.1	99.6
Nonstocked	3.7	3.5	99.6
Western			
Douglas-fir	110.8	30.7	89.6

Ponderosa pine	66.3	20.3	70.4
Western white pine	69.2	25.8	68.3
Fir-spruce	113.0	37.4	137.5
Hemlock-Sitka spruce	152.4	34.1	157.1
Larch	97.0	30.2	65.6
Lodgepole pine	67.8	23.9	62.7
Redwood	186.6	26.9	85.8
Hardwoods	89.0	9.9	79.5
Other forest types	55.4	28.2	90.1
Pinyon-juniper	20.8	21.1	56.3
Chaparral	17.5	25.7	58.7
Nonstocked	18.1	24.4	90.1

* Based on 1997 data for major forest types of the conterminous United States.

^aEastern United States is defined as states east of, and including North Dakota, South Dakota, Nebraska, Kansas, Oklahoma, and Texas. Western United States includes the remaining conterminous States.

^bSoil includes both mineral soils and organic soils (i.e., histosols); carbon densities are to a depth of 1 meter.

Step 1c. Estimate Carbon in Understory Vegetation

To estimate carbon in understory vegetation, equations based on Birdsey (1992) were applied to the database that was used to produce the compiled forest statistics in Smith et al. (2001). Understory vegetation is defined as all biomass of undergrowth plants in a forest, including woody shrubs and trees less than one-inch diameter, measured at breast height. A ratio of understory carbon to live tree carbon was calculated, and multiplied by 100 to display the percent that understory carbon is as related to live tree carbon. The average percent understory carbon to live tree carbon was calculated by region and forest type. This percent was multiplied by the live tree carbon data in 1987 and 1997 to calculate understory carbon. These percentages are given in Table N-2. This procedure was used instead of applying the Birdsey equations directly, because detailed databases are not available for inventory years prior to 1987. Using average estimates results in consistent historical carbon estimates for all survey years.

Table N-2: Ratios of Understory and Down Dead Wood Carbon to Live Tree Carbon* (percent)

Region/Forest Type	Ratio of Understory Carbon to Live Tree Carbon	Ratio of Down Dead Wood Carbon to Live Tree Carbon
Northeast		
White-red-jack pine	2.5	10.8
Spruce-fir	2.6	13.3
Longleaf-Slash pine	2.5	10.8
Loblolly-shortleaf pine	2.5	10.8
Oak-pine	2.8	12.9
Oak-hickory	2.4	10.9
Oak-gum-cypress	2.6	11.1
Elm-ash-cottonwood	2.6	11.1
Maple-beech-birch	1.9	11.1
Aspen-birch	2.7	13.6
Other Forest Types	8.9	3.8
Nonstocked	8.9	3.8
North Central		
White-red-jack pine	1.8	9.8
Spruce-fir	2.2	17.4
Longleaf-Slash pine	2.4	7.4
Loblolly-shortleaf pine	2.4	7.4
Oak-pine	1.9	7.2
Oak-hickory	2.3	9.6
Oak-gum-cypress	2.3	9.6

Elm-ash-cottonwood	2.2	10.8
Maple-beech-birch	2.2	10.8
Aspen-birch	2.8	13.3
Other Forest Types	5.5	4.1
Nonstocked	5.5	4.1
Southeast		
White-red-jack pine	6.8	23.9
Spruce-fir	6.8	23.9
Longleaf-Slash pine	6.8	23.9
Loblolly-shortleaf pine	6.8	23.9
Oak-pine	5.2	28.0
Oak-hickory	4.4	24.2
Oak-gum-cypress	2.2	21.8
Elm-ash-cottonwood	2.2	21.8
Maple-beech-birch	4.4	24.2
Aspen-birch	2.2	21.8
Other Forest Types	11.9	2.0
Nonstocked	11.9	2.0
South Central		
White-red-jack pine	5.9	18.6
Spruce-fir	5.9	18.6
Longleaf-Slash pine	5.9	18.6
Loblolly-shortleaf pine	5.9	18.6
Oak-pine	4.4	17.3
Oak-hickory	3.7	15.0
Oak-gum-cypress	2.2	15.7
Elm-ash-cottonwood	2.2	15.7
Maple-beech-birch	3.7	15.0
Aspen-birch	2.2	15.7
Other Forest Types	16.9	1.7
Nonstocked	16.9	1.7
Pacific Northwest Eastside of Cascades		
Douglas-fir	1.6	10.0
Ponderosa Pine	2.5	12.6
Western White Pine	1.6	10.0
Fir-Spruce	1.1	15.7
Hemlock-Sitka spruce	1.6	10.0
Larch	1.6	10.0
Lodgepole pine	2.6	21.3
Redwood	1.9	25.8
Other hardwoods	1.4	8.9
Unclassified & other	2.5	12.6
Pinyon-Juniper	10.7	3.7
Chaparral	9.7	2.1
Nonstocked	9.7	2.1
Pacific Northwest Westside of Cascades		
Douglas-fir	2.0	11.9
Ponderosa Pine	2.5	18.1
Western White Pine	2.5	18.1
Fir-Spruce	1.0	13.7
Hemlock-Sitka spruce	1.0	13.7
Larch	2.0	11.9
Lodgepole pine	1.7	16.4
Redwood	2.0	11.9
Other hardwoods	4.5	3.9

Unclassified & other	1.7	16.4
Pinyon-Juniper	20.2	3.7
Chaparral	14.2	3.0
Nonstocked	14.2	3.0
Rocky Mountain, Northern		
Douglas-fir	2.6	19.2
Ponderosa Pine	2.4	19.6
Western White Pine	2.2	9.7
Fir-Spruce	1.7	14.8
Hemlock-Sitka spruce	2.0	18.7
Larch	2.2	9.7
Lodgepole pine	2.4	19.6
Redwood	2.2	9.7
Other hardwoods	1.9	14.2
Unclassified & other	2.2	9.7
Pinyon-Juniper	16.1	3.2
Chaparral	16.1	3.2
Nonstocked	16.1	3.2
Rocky Mountain, Southern		
Douglas-fir	2.8	19.4
Ponderosa Pine	4.1	21.6
Western White Pine	2.8	19.4
Fir-Spruce	2.2	17.4
Hemlock-Sitka spruce	2.8	19.4
Larch	2.8	19.4
Lodgepole pine	3.1	12.8
Redwood	2.8	19.4
Other hardwoods	9.2	26.7
Unclassified & other	10.7	3.3
Pinyon-Juniper	9.8	3.9
Chaparral	9.8	3.9
Nonstocked	2.6	15.2
Pacific Southwest		
Douglas-fir	2.3	15.5
Ponderosa Pine	2.6	15.2
Western White Pine	2.2	11.5
Fir-Spruce	2.6	15.2
Hemlock-Sitka spruce	2.6	15.2
Larch	4.6	10.8
Lodgepole pine	2.6	15.2
Redwood	4.4	9.7
Other hardwoods	2.8	11.5
Unclassified & other	9.9	3.1
Pinyon-Juniper	15.3	3.5
Chaparral	15.3	3.5
Nonstocked	2.5	10.8

* Based on data from 1997. Regions are defined in Figure 6-2 of the Land-Use Change and Forestry Chapter.

Step 1d. Estimate Carbon in Forest Floor

To estimate forest floor carbon, the forest floor equations (Smith and Heath, in review) were applied to the dataset described in Step 1a. Forest floor carbon is the pool of organic carbon (litter, duff, humus, and small woody debris) above the mineral soil and includes woody fragments with diameters of up to 7.5 cm. Table N-1 shows the average forest floor carbon densities by forest type, as calculated by applying the equations to the 1997 data.

Step 1e. Estimate Carbon in Down Dead Wood

To estimate carbon in down dead wood, a procedure similar to estimating carbon in understory vegetation was used. Down dead wood is defined as dead wood pieces not attached to trees, greater than 7.5 cm diameter, including stumps and roots of harvested trees. Down dead wood was estimated in the projections by using decay rates applied to logging residue, along with equations that estimate down dead wood not related to harvesting. The ratio of down dead wood carbon to live tree carbon was calculated, and multiplied by 100 to display the ratio as a percentage. The average percentage of down dead wood carbon as compared to live tree carbon was calculated by region and forest type. The percent was multiplied by the live tree carbon data based on the dataset described in step 1a to calculate down dead wood carbon. These percentages are given in Table N-2. This procedure was used because detailed databases are not available for older data. By using average estimates, carbon estimates from historical data are consistent with carbon estimates from current FIA data.

Step 1f. Estimate Forest Soil Carbon

To estimate forest soil carbon, soil carbon estimates for 1 meter depth were obtained from the STATSGO database (USDA 1991). A forest type coverage (Powell et al. 1993) was overlaid onto the soil carbon estimates derived from STATSGO. An average soil carbon estimate was then calculated by forest type. Soil organic carbon of both mineral soils and organic soils (histosol soil order, characterized as soils that develop in wetland areas, and have greater than 20 to 30 percent organic matter by weight, depending on clay content) was included. Coarse roots were included with tree carbon estimates rather than with soils. The soil carbon estimates are given in Table N-1. These estimates were multiplied by the area of forest land in each forest type for all years. Thus, any change in soil carbon is purely a reflection of the changing forest land base.

Step 1g. Calculate Net Carbon Stock Changes

The next step was to calculate the average annual net carbon stock change for each forest carbon pool for the years from 1990 through 2000. The net annual stock change for each pool for 1987 through 1997 was derived by subtracting the 1987 stock from the 1992 stock, and dividing by the number of years between estimates (10 years). The stocks, by definition, correspond to the stock as of January 1 of the given year. The net annual stock changes for 1997 through 2000 were derived in the same way using the 1997 and 2001 stocks.

Step 2. Estimate Harvested Wood Carbon Fluxes

The first step in estimating harvested wood (i.e., wood products and landfilled wood) carbon flux estimates was to compile historical data on: the production of lumber, plywood and veneer, pulp and other products; product and log imports and exports; and fuelwood (in terms of million cubic feet of roundwood equivalent beginning in the year 1900, as described in Skog and Nicholson 1998). Data were obtained from USDA (1964), Ulrich (1989), and Howard (2001). Projected products and roundwood use were obtained from the models used for the USDA Forest Service 2000 Resource Planning Act Assessment (Haynes et al. 2001b, Ince 1994). Roundwood products include logs, bolts, and other round timber generated from harvesting trees for industrial or consumer use. The harvested wood-to-carbon conversion factors (as listed in Skog and Nicholson 1998) were applied to annual estimates and projections to produce an estimate for carbon in roundwood in products. Roundwood consumed was categorized according to product, such as lumber, railroad ties, and paper, because the time carbon remains in those products differs substantially. The dynamics of carbon loss through decay or through disposal of the product is summarized as the half-life of each product (Skog and Nicholson 1998). The resulting estimates can be applied to products to derive the net carbon change in wood products and landfills. Note that, unlike forest carbon stock estimates, carbon in harvested wood products estimates are derived as a carbon stock change. In other words, the annual roundwood production is a change variable already before it is converted to carbon.

Step 3. Sum the Results from Step 1 and Step 2 for the Total Net Flux from U.S. Forests

In the final step, net changes in forest carbon stocks are added to net changes in harvested wood carbon stocks, to obtain estimates of total net forest flux (see Table N-3).

Table N-3: Net CO₂ Flux from U.S. Forest Carbon Stocks (Tg CO₂ Eq.)

Description	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000
Forests	(773.7)	(773.7)	(773.7)	(773.7)	(773.7)	(773.7)	(773.7)	(546.3)	(546.3)	(546.3)	(546.3)
Trees	(469.3)	(469.3)	(469.3)	(469.3)	(469.3)	(469.3)	(469.3)	(447.3)	(447.3)	(447.3)	(447.3)
Understory	(11.0)	(11.0)	(11.0)	(11.0)	(11.0)	(11.0)	(11.0)	(14.7)	(14.7)	(14.7)	(14.7)
Forest Floor	(25.7)	(25.7)	(25.7)	(25.7)	(25.7)	(25.7)	(25.7)	29.3	29.3	29.3	29.3
Down Dead Wood	(55.0)	(55.0)	(55.0)	(55.0)	(55.0)	(55.0)	(55.0)	(58.7)	(58.7)	(58.7)	(58.7)
Forest Soils	(212.7)	(212.7)	(212.7)	(212.7)	(212.7)	(212.7)	(212.7)	(55.0)	(55.0)	(55.0)	(55.0)
Harvested Wood	(209.0)	(198.0)	(202.8)	(203.9)	(210.5)	(205.3)	(205.3)	(212.7)	(205.3)	(216.3)	(223.7)
Wood Products	(47.7)	(40.7)	(46.6)	(54.6)	(60.9)	(55.0)	(55.0)	(58.7)	(51.3)	(62.3)	(66.0)
Landfilled Wood	(161.3)	(157.3)	(156.2)	(149.2)	(149.6)	(150.3)	(150.3)	(154.0)	(154.0)	(154.0)	(157.7)
Total	(982.7)	(971.7)	(976.4)	(977.5)	(984.1)	(979.0)	(979.0)	(759.0)	(751.7)	(762.7)	(770.0)

Note: Parentheses indicate net carbon “sequestration” (i.e., accumulation into the carbon pool minus emissions or stock removal from the carbon pool). The sum of the net stock changes in this table (i.e., total net flux) is an estimate of the actual net flux between the total forest carbon pool and the atmosphere. Lightly shaded areas indicate values based on a combination of historical data and projections. Forest values are based on periodic measurements; harvested wood estimates are based on annual surveys. Totals may not sum due to independent rounding.

The Forest Sector Modeling Projection System

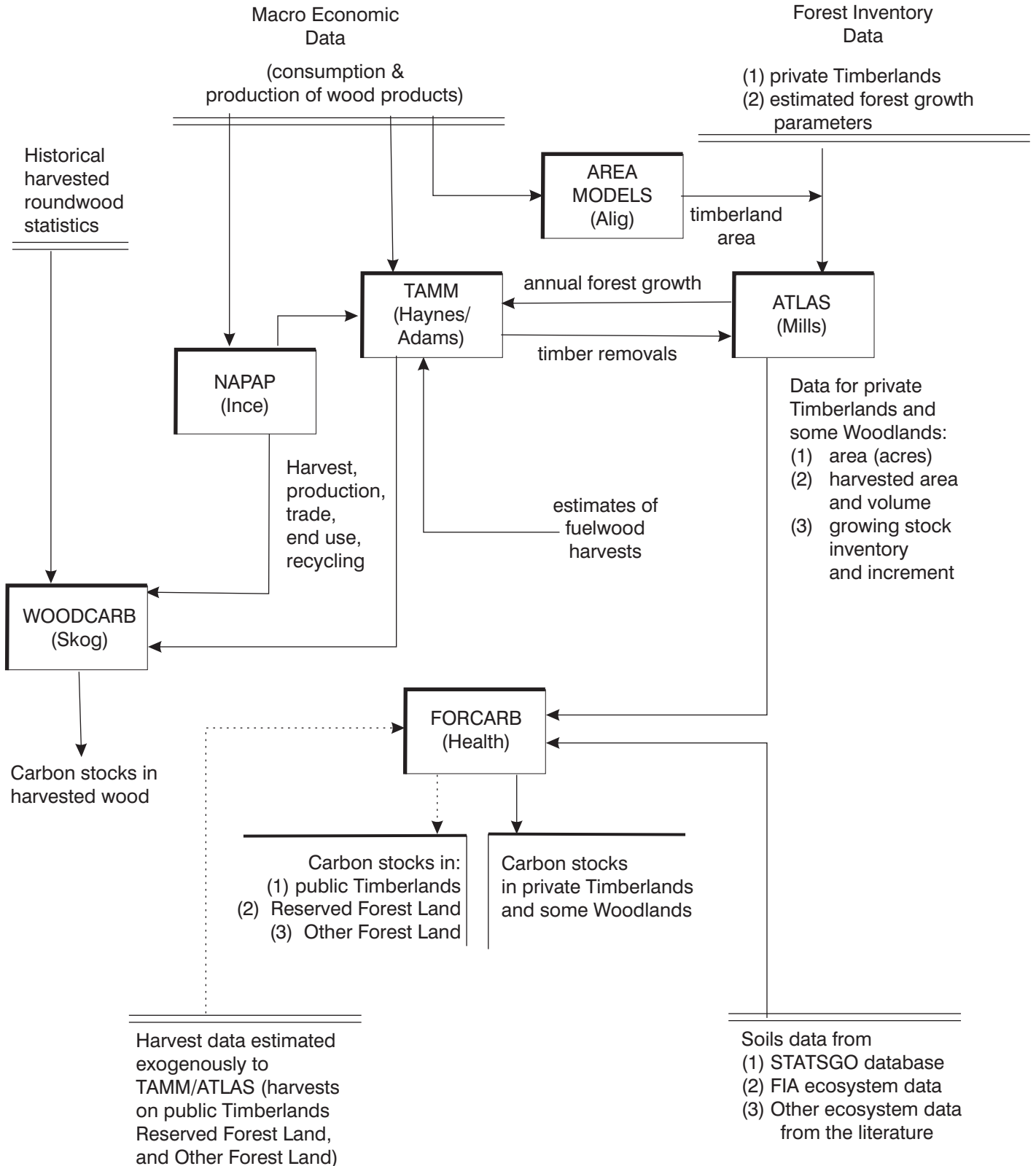
The modeling projection system is a set of models that has been used for the USDA Forest Service, Resource Planning Act Assessment since the late 1980’s (see Figure N-1). The models include an area change model (Alig 1985), a timber market model (TAMM; Adams and Haynes 1980), a pulp and paper model (NAPAP; Ince 1994) and an inventory model (ATLAS; Mills and Kincaid 1992). Many of these models are econometric models, designed to project the demand and supply and prices in the forest sector. Results of the projection include timber volume, forest areas, harvests, and primary product production. To see all the assumptions and results of the modeling system for 2001, see Haynes et al. (2001b).

The FORCARB model (Plantinga and Birdsey 1993, Heath and Birdsey 1993, and Heath et al. 1996) uses data on timber volume, forest areas, and harvests from the modeling system to estimate carbon in trees using biometrical relationships between carbon and live tree volume. FORCARB estimates carbon in all other forest ecosystem components, producing carbon density estimates similar to those in Table N-1 and Table N-2. The model WOODCARB (Skog and Nicholson 1998) uses harvested roundwood product statistics, along with end-use, decay rate, and duration information to estimate carbon in harvested wood.

Figure N-1: Forest Sector Modeling Projection System

This figure illustrates the models, data inputs, and data outputs that compose the forest sector modeling projection system. Names of model authors are in parentheses in each model box to facilitate identification of model citations. Data that are external to the models are marked with double lines.

Figure N-1: Forest Sector Modeling Projection System



ANNEX O: Methodology for Estimating CH₄ Emissions from Landfills

Landfill methane (CH₄) is produced from a complex process of waste decomposition and subsequent fermentation under anaerobic conditions. The amount and rate of methane production depends upon the characteristics of the landfilled material and the surrounding environment. To estimate the amount of methane produced in a landfill in a given year, the following information is needed: the quantity of waste in the landfill, the landfill characteristics, the residence time of the waste in the landfill, and the landfill capacity.

The amount of methane emitted from a landfill is less than the amount generated. If no measures are taken to extract the methane, a portion of it will oxidize as it travels through the top layer of the landfill cover. The portion of the methane that oxidizes turns primarily to carbon dioxide (CO₂). If the methane is extracted and combusted (e.g., flared or used for energy), then that portion of the methane produced in the landfill will not be emitted as methane, but again, would be oxidized to CO₂. In general, landfill-related CO₂ emissions are of biogenic origin and primarily result from the decomposition, either aerobic or anaerobic, of organic matter such as food or yard wastes.¹

Methane emissions are primarily driven by the quantity of waste in landfills. From an analysis of the population of municipal solid waste (MSW) landfills, landfill-specific data were extracted and used in an emissions model to estimate the amount of methane produced by municipal solid waste. Although not explicitly modeled, methane emissions from industrial landfills were assumed to be seven percent of the total methane generated from MSW at landfills. Total methane emissions were estimated by adding the methane from MSW landfills, subtracting the amount recovered or used for energy or flared, subtracting the amount oxidized in the soil, and adding emissions from industrial landfills. The steps taken to estimate emissions from U.S. landfills for the years 1990 through 2000 are discussed in greater detail below.

Step 1: Estimate Municipal Solid Waste-in-Place Contributing to Methane Emissions

First, landfills were characterized as of 1990 based on a landfill survey (EPA 1988). Each landfill was characterized in terms of its year of opening, waste acceptance during operation, year of closure, and design capacity. Following characterization of the landfill population, waste was simulated to be placed in these landfills. For 1990 through 2000, waste disposal estimates were based on annual *BioCycle* (2001) data. Landfills were simulated to open and close based on waste disposal rates. If landfills reached their design capacity, they were simulated to close. New landfills were simulated to open when a significant shortfall in disposal capacity was predicted. Simulated new landfills were assumed to be larger, on average, reflecting the trend toward fewer and more centralized facilities. The analysis updated the landfill characteristics each year, calculating the total waste-in-place and the profile of waste disposal over time. Table O-1 shows the amount of waste landfilled each year and the total estimated waste-in-place contributing to methane emissions.

Step 2: Estimate Landfill Methane Production

Emissions for each landfill were estimated by applying the emissions model (EPA 1993) to the landfill waste-in-place contributing to methane production. The model estimates that landfilled waste generates methane for 30 years after disposal. Total emissions were then calculated as the sum of emissions from all landfills, open and closed.

Step 3: Estimate Industrial Landfill Methane Production

Industrial landfills receive waste from factories, processing plants, and other manufacturing activities. Because no data were available on methane generation at industrial landfills, emissions from industrial landfills were assumed to equal seven percent of the total methane emitted from MSW landfills (EPA 1993). The EPA landfill survey

¹ Emissions and sinks of biogenic carbon are accounted for in the Land-Use Change and Forestry chapter.

contained estimates of industrial waste (EPA 1988). The organic content of industrial waste represents seven percent of the methane producing capacity of MSW. These emissions are shown in Table O-2.

Step 4: Estimate Methane Emissions Avoided

The estimate of methane emissions avoided was based on landfill-specific data on flares and landfill gas-to-energy (LFGTE) projects.

Step 4a. Estimate Methane Emissions Avoided through Flaring

The quantity of methane flared was based on data collected from flaring equipment vendors, including information on the quantity of flares, landfill gas flow rates, and year of installation. To avoid double counting, flares associated with landfills that had an LFGTE project were excluded from the flaring analysis. Total methane recovered was estimated by summing the median landfill gas flow rate for each remaining flare. However, several vendors provided information on the size of the flare rather than the landfill gas flow rate. To estimate a median flare rate for flares associated with these vendors, the size of the flare was matched with the size and corresponding flow rate provided by the other vendors.

Step 4b. Estimate Methane Emissions Avoided through Landfill gas-to-energy (LFGTE) projects

The quantity of methane avoided due to LFGTE systems was estimated based on information in a database compiled by EPA's Landfill Methane Outreach Program (LMOP). Using data on landfill gas flow and energy generation, the total direct methane emissions avoided were estimated.

To avoid double counting flares associated with LFGTE projects, the flare estimates were adjusted to account for LFGTE projects for which an associated flare could not be identified.

Step 5: Estimate Methane Oxidation

As discussed above, a portion of the methane escaping from a landfill through its cover oxidizes in the top layer of the soil. The amount of oxidation that occurs is uncertain and depends upon the characteristics of the soil and the environment. For purposes of this analysis, it was assumed that ten percent of the methane produced, minus the amount of gas recovered for flaring or LFGTE projects, was oxidized in the soil (Liptay et al. 1998).

Step 6: Estimate Total Methane Emissions

Total methane emissions were calculated by adding emissions from MSW and industrial waste, and subtracting methane recovered and oxidized, as shown in Table O-2.

Table O-1: Municipal Solid Waste (MSW) Contributing to Methane Emissions (Tg unless otherwise noted)

Description	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000
Total MSW Generated ^a	266	255	265	278	293	296	297	309	340	347	371
Percent of MSW Landfilled ^a	77%	76%	72%	71%	67%	63%	62%	61%	61%	60%	61%
Total MSW Landfilled	205	194	191	198	196	187	184	188	207	208	226
MSW Contributing to Emissions ^b	4,926	5,027	5,162	5,292	5,428	5,559	5,676	5,790	5,906	6,035	6,147

^a Source: *BioCycle* (2001). The data, originally reported in short tons, are converted to metric tons.

^b The emissions model (EPA 1993) defines all waste that has been in place for less than 30 years as contributing to methane emissions.

Table O-2: Methane Emissions from Landfills (Gg)

Activity	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000
MSW Generation	11,599	11,837	12,168	12,499	12,847	13,218	13,490	13,774	14,015	14,348	14,617
Large Landfills	4,780	4,817	4,883	4,950	5,038	5,129	5,199	5,280	5,351	5,453	5,520
Medium Landfills	5,545	5,720	5,954	6,190	6,424	6,682	6,868	7,057	7,211	7,424	7,614
Small Landfills	1,273	1,300	1,332	1,359	1,385	1,407	1,423	1,438	1,453	1,471	1,483

Industrial Generation	731	746	767	787	809	833	850	868	883	904	921
Potential Emissions	12,330	12,582	12,935	13,286	13,657	14,051	14,340	14,642	14,898	15,252	15,538
Emissions Avoided	(1,119)	(1,387)	(1,601)	(1,848)	(2,225)	(2,682)	(3,244)	(3,820)	(4,362)	(4,607)	(4,874)
Landfill Gas-to-											
Energy	(692)	(728)	(784)	(855)	(977)	(1,017)	(1,171)	(1,415)	(1,729)	(1,984)	(2,196)
Flare	(427)	(659)	(817)	(994)	(1,248)	(1,665)	(2,073)	(2,405)	(2,633)	(2,623)	(2,678)
Oxidation	(1,048)	(1,045)	(1,057)	(1,065)	(1,062)	(1,054)	(1,025)	(995)	(965)	(974)	(974)
Net Emissions	10,162	10,150	10,277	10,373	10,370	10,315	10,072	9,827	9,571	9,671	9,690

Note: Totals may not sum due to independent rounding.

Note: MSW generation in Table O-2 represents emissions before oxidation. In other tables throughout the text, MSW generation estimates account for oxidation.

ANNEX P: Key Source Analysis

This Annex provides an analysis of key sources of emissions found in this report in keeping with the IPCC's *Good Practice Guidance* (IPCC 2000). In order to ensure accuracy and reliability of inventory estimates, quality assurance and quality control (QA/QC) resources and activities should be directed to the key source categories in a given country's greenhouse gas emissions inventory. A key source category is defined as a "[source category] that is prioritized within the national inventory system because its estimate has a significant influence on a country's total inventory of direct greenhouse gases in terms of the absolute level of emissions, the trend in emissions, or both."¹ By definition, key source categories are sources that have the greatest contribution to the absolute overall level of national emissions. In addition, when an entire time series of emission estimates is prepared, a thorough investigation of key source categories must also include accounting for the influence of trends of individual source categories. Therefore, a trend assessment is also conducted based on an attempt to identify source categories for which significant uncertainty in the estimate would have considerable effects on overall emission trends. This analysis culls out source categories that diverge from the overall trend in national emissions. Finally, a qualitative evaluation of key source categories should be performed, in order to capture any key source categories that were not identified in either of the quantitative analyses.

The methodology for conducting a key source analysis, as defined by IPCC's *Good Practice Guidance* (IPCC 2000), includes:

- Tier 1 approach (including both level and trend assessments);
- Tier 2 approach (including both level and trend assessments, and incorporating uncertainty analysis); and
- Qualitative approach.

Following this introduction, the Annex will present and analyze key source categories; discuss Tier 1, Tier 2, and qualitative approaches to identifying key sources; provide level and trend assessment equations; and provide a brief statistical evaluation of IPCC's quantitative methodologies for defining key sources.

Table P-1 presents the key source categories for the United States using emissions data in this report, and ranked according to their sector and global warming potential-weighted emissions in 2000. The table also identifies the criteria used in identifying these source categories (i.e., level, trend, and/or qualitative assessments).

Table P-1: Key Source Categories for the United States (1990-2000) Based on Tier 1 Approach

IPCC Source Categories	Gas	Criteria			2000 Emissions (Tg CO ₂ Eq.)
		Level	Trend	Qual. ^a	
Energy					
CO ₂ Emissions from Stationary Combustion - Coal	CO ₂	✓	✓	C,T,Q	2,030.1
CO ₂ Emissions from Stationary Combustion - Oil	CO ₂	✓	✓		640.7
Mobile Combustion: Road & Other	CO ₂	✓	✓		1,503.2
CO ₂ Emissions from Stationary Combustion - Gas	CO ₂	✓	✓		1,162.9
Mobile Combustion: Aviation	CO ₂	✓			196.5
Fugitive Emissions from Oil & Gas Operations	CH ₄	✓	✓		138.2
Mobile Combustion: Marine	CO ₂	✓	✓		89.9
Fugitive Emissions from Coal Mining & Handling	CH ₄	✓	✓		61.0
Mobile Combustion: Road & Other	N ₂ O	✓			55.7
Indirect CO ₂ Emissions from CH ₄ Oxidation	CO ₂		✓		26.3
International Bunker Fuels ^b	Several			✓	101.2
Non-Energy Use of Fossil Fuels ^b	CO ₂			✓	409.6

¹ See chapter 7 "Methodological Choice and Recalculation" in IPCC (2000).
< <http://www.ipcc-nggip.iges.or.jp/public/gp/gpgaum.htm>>

Industrial Processes				
CO ₂ Emissions from Other Industrial Processes	CO ₂	✓	✓	107.6
Emissions from Substitutes for Ozone Depleting Substances	Several	✓	✓	57.8
CO ₂ Emissions from Cement Production	CO ₂	✓		41.1
HFC-23 Emissions from HCFC-22 Manufacture	HFCs	✓	✓	29.8
SF ₆ Emissions from Electrical Equipment	SF ₆		✓	14.4
PFC Emissions from Aluminum Production	PFCs		✓	7.9
N ₂ O Emissions from Adipic Acid Production	N ₂ O		✓	8.1
PFC, HFC, and SF ₆ Emissions from Semiconductor Manufacturing	Several		✓	7.4
Agriculture				
Direct N ₂ O Emissions from Agricultural Soils	N ₂ O	✓		217.8
CH ₄ Emissions from Enteric Fermentation in Domestic Livestock	CH ₄	✓	✓	123.9
Indirect N ₂ O Emissions from Nitrogen Used in Agriculture	N ₂ O	✓		79.8
CH ₄ Emissions from Manure Management	CH ₄	✓	✓	37.5
Waste				
CH ₄ Emissions from Solid Waste Disposal Sites	CH ₄	✓	✓	203.5
CO ₂ Emissions from Waste Incineration	CO ₂		✓	22.5
Subtotal				6,843.4
Total				7,001.2
Percent of Total				98.0%

^a Qualitative criteria.

^b Emissions from these sources not included in totals.

Notes: Sinks (e.g., LUCF, Landfill Carbon Storage) are not included in this analysis. The Tier 1 approach for identifying key source categories does not directly include assessment of uncertainty in emissions estimates.

Table P-2 provides a complete listing of source categories by IPCC sector and with additional comments on the criteria used in identifying key source categories. Specifically, the level assessment was performed for each year that inventory data was available (i.e., 1990 to 2000). As the emissions change over time, categories may fall under or over the threshold for being key. The following points should be noted regarding the key sources identified.

Due to the relative quantity of CO₂ emissions from fossil fuel combustion—particularly from mobile source and stationary combustion of coal, gas, and oil—these sources contributed most to each year's level assessment. Additionally, the following sources were the largest contributors to the level assessments for each year (listed in descending order):

- CH₄ from solid waste disposal sites;
- N₂O from agricultural soils;
- CO₂ emissions from mobile combustion in the aviation and marine sectors;
- Fugitive emissions from oil and gas operations;
- CH₄ from enteric fermentation in domestic livestock;
- CO₂ emissions from other industrial processes;
- Fugitive emissions from coal mining; and
- N₂O emissions from mobile road source emissions.

The remaining key sources identified under the level assessment varied by year. The following four source categories were determined to be key using the level assessment for only part of the complete times series:

- CO₂ emissions from cement production (1991, 1993 to 1996);
- HFC and PFC emissions from substitutes for ozone depleting substances (1997 to 2000);
- HFC-23 emissions from HCFC-22 manufacture (1990, 1992, 1996, 1998); and
- CH₄ emissions from manure management (1995).

Although other sources have fluctuated by greater percentages since 1990, by virtue of their size, CO₂ emissions from stationary combustion of coal, gas, and oil, and from mobile combustion from road vehicles are the greatest contributors to the overall trend for 2000.

Another large contributor to the overall trend is emissions of substitutes for ozone depleting substances, which are growing quickly with the Montreal Protocol phase-out of ozone depleting substances.

Fugitive emissions from coal mining and PFC emissions from aluminum manufacturing have decreased by approximately 30 and 56 percent, respectively from 1990 through 2000. Reductions in emissions from coal mining are primarily due to EPA's voluntary coalbed methane capture program and the mining of less gassy coal than in previous years. PFC emissions have decreased primarily as a result of emission reduction activities by the aluminum industry.

Six other source categories were determined to be key using the qualitative criteria. A brief discussion of the reasoning for the qualitative designation is given below:

- International bunker fuels are fuels consumed for aviation or marine international transport activities, and emissions from these fuels are reported separately from totals in accordance with IPCC guidelines. If these emissions were included in the totals, bunker fuels would qualify as a key source according to the Tier 1 approach. The amount of uncertainty associated with estimation of emissions from international bunker fuels also supports the qualification of this source category as key.
- Non-energy uses of fossil fuels represent a significant percentage of the total carbon inventory, and the idea that small changes in storage factors for these non-energy uses may result in large changes in storage and emissions qualifies this source category as key.
- Nitrous oxide emissions from adipic acid plants have been dramatically reduced due to the installation of emission control technologies on 3 of the 4 production facilities in the United States. These changes in addition to the uncertainty in this emission source category suggest that it should be treated as key, although it has also been identified using the trend assessment.
- Estimates of SF₆ emissions from electrical equipment have been made using only a limited amount of data; therefore, there is a significant degree of uncertainty associated with them. Although future inventories are expected to incorporate improvements, the current lack of data and small margin under which the category missed both the trend and level assessment thresholds suggests that it should be treated as key.
- Emissions of HFCs, PFCs and SF₆ from semiconductor manufacturing have increased significantly from 1990 through 1999, almost tripling in size. This source category's potential future growth—in addition to historical growth that has already led to list listing as key using the trend assessment—suggests that it should be treated as key.
- Estimated CH₄ emissions from manure management have been significantly revised relative to the previous greenhouse gas inventory. This revision is due to both changes in the estimation methodology and data sources. The reduction in estimated emissions for the entire time series have by approximately 50 to 60 percent, suggests that it should be treated as key, although it has also been identified using the trend assessment

Following the text of this Annex, Table P-3 through Table P-13 contain each individual year's level assessment and contain further detail on where each source falls within the analysis. Table P-14 details the trend assessment for 1990-2000.

Table P-2: U.S Greenhouse Gas Inventory Source Categories Based on Tier 1 Approach

IPCC Source Categories	Direct GHG	2000 Emissions (Tg CO ₂ Eq.)	Key Source Category Flag?	Criteria ^a	Comments
Energy					
CO ₂ Emissions from Stationary Combustion - Coal	CO ₂	2,030.1	✓	L, T	All years
CO ₂ Emissions from Stationary Combustion - Oil	CO ₂	640.7	✓	L, T	All years
CO ₂ Emissions from Stationary Combustion - Gas	CO ₂	1,162.9	✓	L, T	All years
CO ₂ Emissions from Stationary Combustion –	CO ₂	+			
Geothermal					
CO ₂ Emissions from Natural Gas Flaring	CO ₂	6.1			
Non-CO ₂ Emissions from Stationary Combustion	CH ₄	7.5			

Non-CO ₂ Emissions from Stationary Combustion	N ₂ O	14.9			
Mobile Combustion: Road & Other	CO ₂	1,503.2	✓	L, T	All years
Mobile Combustion: Road & Other	CH ₄	4.1			
Mobile Combustion: Road & Other	N ₂ O	55.7	✓	L	All years
Mobile Combustion: Aviation	CO ₂	196.5	✓	L	All years
Mobile Combustion: Aviation	CH ₄	0.2			
Mobile Combustion: Aviation	N ₂ O	1.9			
Mobile Combustion: Marine	CO ₂	89.9	✓	L, T	All years
Mobile Combustion: Marine	CH ₄	0.1			
Mobile Combustion: Marine	N ₂ O	0.6			
Fugitive Emissions from Coal Mining & Handling	CH ₄	61.0	✓	L, T	All years
Fugitive Emissions from Oil & Gas Operations	CH ₄	138.2	✓	L, T	All years
Indirect CO ₂ Emissions from CH ₄ Oxidation	CO ₂	26.3	✓	T	
International Bunker Fuels ^b	Several	101.2	✓	Q	
Non-Energy Use of Fossil Fuels ^b	CO ₂	409.6	✓	Q	
Industrial Processes					
CO ₂ Emissions from Cement Production	CO ₂	41.1	✓	L	Level in 1991, 1993 - 1997
CO ₂ Emissions from Lime Production	CO ₂	13.3			
CO ₂ Emissions from Other Industrial Processes	CO ₂	107.6	✓	L, T	All years
CH ₄ Emissions from Other Industrial Processes	CH ₄	1.7			
N ₂ O Emissions from Adipic Acid Production	N ₂ O	8.1	✓	T, Q	
N ₂ O Emissions from Nitric Acid Production	N ₂ O	19.8			
PFC Emissions from Aluminum Production	PFCs	7.9	✓	T	
SF ₆ Emissions from Magnesium Production	SF ₆	4.0			
SF ₆ Emissions from Electrical Equipment	SF ₆	14.4	✓	T, Q	
HFC, PFC, and SF ₆ Emissions from	Several	7.4	✓	Q	
Semiconductor Manufacturing					
Emissions from Substitutes for Ozone Depleting Substances	Several	57.8	✓	L, T	Level from 1997 - 2000
HFC-23 Emissions from HCFC-22 Manufacture	HFCs	29.8	✓	L, T	Level in 1990, 1992, 1996, 1998
Agriculture					
CH ₄ Emissions from Enteric Fermentation in Domestic Livestock	CH ₄	123.9	✓	L, T	All years
CH ₄ Emissions from Manure Management	CH ₄	37.5	✓	L, Q	Level in 1995
N ₂ O Emissions from Manure Management	N ₂ O	17.5			
Direct N ₂ O Emissions from Agricultural Soils	N ₂ O	217.8	✓	L	All years
Indirect N ₂ O Emissions from Nitrogen Used in Agriculture	N ₂ O	79.8	✓	L	All years
Agriculture					
CH ₄ Emissions from Rice Production	CH ₄	7.5			
CH ₄ Emissions from Agricultural Residue Burning	CH ₄	0.8			
N ₂ O Emissions from Agricultural Residue Burning	N ₂ O	0.5			
Waste					
CH ₄ Emissions from Solid Waste Disposal Sites	CH ₄	203.5	✓	L, T	All years
CH ₄ Emissions from Wastewater Handling	CH ₄	28.7			
N ₂ O Emissions from Wastewater Handling	N ₂ O	8.5			
CO ₂ Emissions from Waste Incineration	CO ₂	22.5	✓	T	
N ₂ O Emissions from Waste Incineration	N ₂ O	0.2			

+ Does not exceed 0.05 Tg CO₂ Eq.

^a Qualitative criteria.

^b Emissions from these sources not included in totals.

Notes: Sinks (e.g., LUCF, Landfill Carbon Storage) are not included in this analysis. The Tier 1 approach for identifying key source categories does not directly include assessment of uncertainty in emissions estimates.

Tier 1 Approach

The Tier 1 method for identifying key source categories assesses the impacts of all IPCC-defined source categories on the level and trend of the national emission inventory for the 1990 through 2000 time-series, but works independently of any formal uncertainty analysis. However, it is important to mention that although conducting a key source analysis can be very valuable in improving the U.S. inventory, it would be ideal to undertake a full uncertainty analysis in order to accurately identify all key sources and to be able to take into account the level of uncertainty associated with each estimate.

When using a Tier 1 approach for the *level*, a pre-determined cumulative emissions threshold is used to identify key source categories. When source categories are sorted in order of decreasing emissions, those that fall at the top of the list and cumulatively account for 95 percent of emissions are considered key source categories. The 95 percent threshold was established based on an evaluation of several inventories, and was designed to establish a general level where the key source category analysis covers 90 percent of inventory uncertainty. The Tier 1 approach for the *trend* uses a 95 percent contribution threshold of the cumulative contribution to the trend assessment metric, which was also designed to establish a general level where the key source category analysis covers 90 percent of inventory uncertainty. The Tier 1 method is completed using a simple spreadsheet analysis based on equations for both level and trend assessments that are described in detail below. It is the current approach that the United States is taking to identify key source categories of greenhouse gas emissions until a rigorous uncertainty analysis is completed.

Level and Trend Assessments of Key Source Categories

Level Assessment

A level assessment was performed for years 1990 to 2000. Key sources were identified as any source category which, when summed in descending order of magnitude for a given year, cumulatively add up to 95 percent of the total level assessment for that year. Level estimates are based upon the following equation:

$$\text{Source Category Level Assessment} = \text{Source Category} / \text{Total Estimate}$$
$$L_{x,t} = E_{x,t} / E_t$$

Where,

$L_{x,t}$ = level assessment for source x in year t
 $E_{x,t}$ = emissions estimate for source x in year t
 E_t = total emissions estimate for year t

Trend Assessment

A trend assessment was then conducted to evaluate how significantly the difference between the source category's trend and the overall inventory trend affect the overall trend. This assessment was done by multiplying the difference between the source category trend and the total inventory trend by the source category level assessment. Trend assessments were based upon the following equation:

$$\text{Source Category Trend Assessment} = (\text{Source Category Level Assessment}) \times \left| \frac{\text{Source Category Trend} - \text{Total Trend}}{\text{Trend}} \right|$$
$$T_{x,t} = L_{x,t} \times \left| \left[\frac{(E_{x,t} - E_{x,0})}{E_{x,t}} - \frac{(E_t - E_0)}{E_t} \right] \right|$$

Where,

$T_{x,t}$ = trend assessment for source x in year t
 $L_{x,t}$ = level assessment for source x in year t
 $E_{x,t}$ and $E_{x,0}$ = emissions estimates for source x in year t and year 0, respectively
 E_t and E_0 = total emissions estimate for year t and year 0, respectively

0 = base year (e.g., 1990)

The following section of this annex evaluates these key source category analyses. The remainder of the annex summarizes the key source categories identified by these analyses, and quantifies their contribution to total level and trend assessments.

Tier 2 Approach

IPCC recommends that inventory agencies use the Tier 2 method for identifying key source categories if nationally derived source-level uncertainties are measured. The Tier 2 approach is a more detailed analysis that builds on the Tier 1 approach by multiplying the results of the Tier 1 analysis by the relative uncertainty of each source category. This method is likely to reduce the number of key source categories under consideration. Using the Tier 2 approach, key source categories represent 90 percent of the quantified uncertainty contribution, as opposed to those that sum to the pre-determined cumulative emissions or trend threshold. A simple spreadsheet version accounts for the uncertainty contribution by applying the source category percentage uncertainty estimates to the Tier 1 level and trend assessments.

A detailed, more complete assessment of uncertainty uses Monte Carlo uncertainty modeling. The U.S. EPA is currently working on preparing such an analysis using procedures for gathering necessary data inputs and estimating uncertainty using a Monte Carlo model developed with @Risk[®] software. The project, which is in the initial phase of developing the uncertainty model, has as its goal developing a simulation model to estimate uncertainty for all source categories of the U.S. Inventory, and in total. The Monte Carlo model develops estimates of uncertainty for inventory source categories based on (a) mathematical models used to estimate emissions for each source category; (b) source category specific input parameters and emission estimates; and (c) the statistical properties underlying the input parameters and estimates.

Qualitative Approach

In addition to conducting a quantitative assessment like the ones described above, a variety of qualitative criteria could be applied to identify additional key source categories. The following qualitative criteria for identifying key source categories have been outlined in the *Good Practice Guidance* (IPCC 2000). A source category should be identified as a key source if:

- Mitigation techniques and technologies are being implemented to reduce emissions from the source category that are expected to be reflected in the inventory estimates;
- Significant changes in emissions (i.e., growth or decline) from the source category is expected in the future;
- High uncertainty is evident for the source category;
- Unexpectedly low or high emissions, or other order of magnitude discrepancies, are apparent for the source category; and
- Major changes in estimation methodology or data have occurred.

In many cases, the results of this qualitative approach to identifying key source categories will overlap with source categories already defined as key source categories through the quantitative analysis. However, the qualitative method may illuminate a few additional key source categories, which should then be included in the final list of key source categories. The application of such qualitative criteria are primarily intended to identify any additional source categories that were “just under” the threshold criteria for the level assessment and not for extremely minor source categories.

Six source categories are also considered key from a qualitative standpoint, these include:

- International Bunker Fuels,
- Non-Energy Use of Fossil Fuels,
- N₂O from adipic acid production,
- SF₆ from electrical equipment,
- HFCs, PFCs, and SF₆ from semiconductor manufacturing, and

- CH₄ from manure management.

Of these sources, N₂O from adipic acid production, SF₆ from electrical equipment, and CH₄ from manure management are also considered key sources in either the level or the trend assessments. Semiconductor manufacture is not considered a key source in either the level or the trend assessments, but the rapid growth rate of this industry identifies it as a qualitative key. Both international bunker fuels and non-fuel use of fossil fuels are considered unique in that they are not included as a part of the inventory totals. Therefore they require additional explanation in this analysis. International bunker fuel emissions are not included in national totals, and are not considered in the level or trend analyses mentioned above, but are considered key from a qualitative standpoint due to their unique position within the emissions accounting framework. Additionally, non-fuel use of fossil fuels is also not included in the level or trend analyses. However, due to the significant quantity of fossil fuels consumed in the United States that are not used to produce energy (generically referred to as feedstocks), it is imperative to understand their fate and to determine how much of the consumption results in emissions, versus in stored carbon.

International Bunker Fuels

International bunker fuels are fuels consumed for aviation or marine international transport activities, and emissions from these fuels are reported separately from totals in accordance with IPCC guidelines. If these emissions were included in the totals, bunker fuels would qualify as a key source according to the Tier 1 approach, as emissions for 2000 are estimated at 101.2 Tg CO₂ Eq. An additional reason to treat bunker fuels as a key source lies in the amount of uncertainty in these emission estimates. Difficulties in collecting this data and the use of varying definitions of bunker fuels are a few of the uncertainties that could have a significant effect on total emission trends.

Non-Energy Use of Fossil Fuel

Fossil fuel feedstocks including naphthas, liquefied petroleum gases, and natural gas are used in the manufacture of a wide variety of man-made chemicals and products, in addition to their use as fuels. Non-fuel feedstock uses of fossil fuels include manufacture of plastics, rubber, synthetic fibers, solvents, paints, fertilizers, pharmaceuticals, and food additives. Subsequent use or disposal of these products may result in either carbon storage or carbon emissions. Industrial coking coal, petroleum coke, distillate and residual fuel oils, and other fossil fuels are also used for non-energy uses. Because non-fuel uses of these fuel types are diverse, the carbon storage and carbon emissions from these non-fuel uses are difficult to characterize.

Non-energy uses of fossil fuels represent a significant percentage of the total carbon inventory. Potential emissions of carbon from non-energy use increased from 319.9 Tg CO₂ Eq. (87.25 Tg C) in 1990 to 409.6 Tg CO₂ Eq. (111.70 Tg C) in 2000, an increase of 28 percent. In the same time frame, carbon stored in products from non-energy use increased from 221.0 Tg CO₂ Eq. (60.26 Tg C) to 283.2 Tg CO₂ Eq. (77.23 Tg C), an increase of 28 percent. Small changes in storage factors for these non-energy uses may result in large changes in storage and emissions. Therefore, non-energy use of fossil fuels is considered to be a key source from a qualitative standpoint.

Evaluation of Key Source Identification Methodologies

Level Assessment

The Tier 1 approach for level assessment defines the source category contribution as the percentage of total inventory emissions from that source category. Only emission source categories are considered.² To determine key source categories, the level assessments are sorted in decreasing order, so that the source categories with the highest level assessments appear first. The level assessments are summed until the threshold of 95 percent is reached; all source categories that fall within that cumulative 95 percent are considered key source categories.

² The level assessment is intended to be applied to sources and to exclude sinks. Although the assessment would still be valid if sinks were included (as unsigned values), the 95 percent threshold by which sources are deemed “key” would lose significance because it is based on an analysis (Flusgrud et al. 1999) of selected inventories where sinks were excluded.

Since the Tier 1 approach does not explicitly incorporate uncertainties, the level assessment key source categories will be the largest contributors to total emissions but will not necessarily have large contributions to the total uncertainty. Focusing resources on improving the methodologies for estimating emissions from the source categories with the largest emissions is undesirable if those emissions are estimated relatively precisely using the current methodologies. Nevertheless, the analysis (reported in IPCC 2000) of several inventories that have source category uncertainties showed that about 90 percent of the total uncertainty could be covered by the source categories in the top 95 percent of emissions.

It is important to note that this key source category analysis can be very sensitive to the definitions of the source categories. If a large source category is split into many subcategories, then the subcategories may have contributions to the total inventory that are too small for those source categories to be considered key. Similarly, a collection of small, non-key source categories adding up to less than 5 percent of total emissions could become key source categories if those source categories were aggregated into a single source category. A consistent approach to addressing this issue is available in the *Good Practice Guidance*. Table 7.1 in IPCC (2000) provides guidance and a suggested list of source categories for analysis, although countries are given some discretion based upon their national circumstances.

Some important components of other source categories were not included in the list of IPCC source categories in the key source category chapter of IPCC's *Good Practice Guidance* (IPCC 2000). These source categories include fossil fuel feedstocks, international bunkers, and emissions from territories. They are potentially large source categories that often are derived from unique data sources, have a significant impact on the uncertainty of the estimates, and therefore ought to be considered as potential key source categories.

Trend Assessment

The Tier 1 approach for trend assessment is defined as the product of the source category level assessment (i.e., source category emissions as a fraction, or percentage, of total emissions) and the absolute difference between the source category trend and the total trend. In turn, the source category trend is defined as the change in source category emissions from the base year to the current year, as a percentage of current year emissions from that source category. The total trend is the percentage change in total inventory emissions from the base year to the current year. Thus, the *source category trend assessment* will be large if the source category represents a large percentage of emissions and/or has a trend that is quite different from the overall inventory trend. Only emissions source categories are considered.³ To determine key source categories, the trend assessments are sorted in decreasing order, so that the source categories with the highest trend assessments appear first. The trend assessments are summed until the threshold of 95 percent is reached; all source categories that fall within that cumulative 95 percent are considered key source categories.

It is important to note that the trend assessment calculation assumes that the base and current year source category emission uncertainties are the same. Therefore, the trend assessment is a useful measure in cases where the percentage uncertainties of the base and current year source category emission levels are thought to be the same. However, its usefulness diminishes when individual source category uncertainties are different between the base year and the current year. Such time series inconsistencies could result from changes in data quality or availability over time. As more rigorous methods to determine uncertainties in emission estimates are applied, it may be necessary to revisit the results of the trend assessments.

Another important caveat to the identification of key source categories through the trend assessment is that, while each individual source category's trend assessment provides a measure of how sensitive the overall trend in the inventory is to the trend of a particular source category, the sum of a number of trend assessments does not yield the

³ The trend assessment is intended to be applied to sources and to exclude sinks. Although the assessment would still be valid if sinks were included (as unsigned values), the 95 percent threshold by which sources are deemed "key" would lose significance because it is based on an analysis (Flusgrud et al. 1999) of selected inventories where sinks were excluded.

total sensitivity of the overall trend to changes in all of those source categories. In other words, the cumulative percentages should not be considered a measure of the percentage contributions to the trend from those source categories.

The trend assessment key source categories are also sensitive to the level of aggregation of the source categories; and the IPCC list of source categories may exclude some important, potentially key source category components.

References

Flugsrud, K., W. Irving, and K. Rypdal (1999) *Methodological Choice in Inventory Preparation. Suggestions for Good Practice Guidance*. Statistics Norway Department of Economic Statistics. 1999/19.

IPCC (2000) *Good Practice Guidance and Uncertainty Management in National Greenhouse Gas Inventories*, Intergovernmental Panel on Climate Change, National Greenhouse Gas Inventories Programme.

Table P-3: 1990 Key Source Tier 1 Analysis - Level Assessment

IPCC Source Categories	Direct Greenhouse Gas	Base Year Estimate (Tg CO ₂ Eq.)	Current Year Estimate (Tg CO ₂ Eq.)	Level Assessment	Cumulative Total
CO ₂ Emissions from Stationary Combustion - Coal	CO ₂	1,692.60	1,692.60	0.28	0.28
Mobile Combustion: Road & Other	CO ₂	1,235.49	1,235.49	0.20	0.48
CO ₂ Emissions from Stationary Combustion - Gas	CO ₂	952.76	952.76	0.16	0.63
CO ₂ Emissions from Stationary Combustion - Oil	CO ₂	662.46	662.46	0.11	0.74
CH ₄ Emissions from Solid Waste Disposal Sites	CH ₄	213.41	213.41	0.03	0.78
Direct N ₂ O Emissions from Agricultural Soils	N ₂ O	193.49	193.49	0.03	0.81
Mobile Combustion: Aviation	CO ₂	176.88	176.88	0.03	0.84
Fugitive Emissions from Oil & Gas Operations	CH ₄	147.64	147.64	0.02	0.86
CH ₄ Emissions from Enteric Fermentation in Domestic Livestock	CH ₄	127.88	127.88	0.02	0.88
CO ₂ Emissions from Other Industrial Processes	CO ₂	123.65	123.65	0.02	0.90
Fugitive Emissions from Coal Mining and Handling	CH ₄	87.12	87.12	0.01	0.92
Indirect N ₂ O Emissions from Nitrogen Used in Agriculture	N ₂ O	73.60	73.60	0.01	0.93
Mobile Combustion: Marine	CO ₂	59.43	59.43	0.01	0.94
Mobile Combustion: Road & Other	N ₂ O	48.86	48.86	0.01	0.95
HFC-23 Emissions from HCFC-22 Manufacture	HFCs	34.98	34.98	0.01	0.95
CO ₂ Emissions from Cement Production	CO ₂	33.28	33.28	0.01	0.96
SF ₆ Emissions from Electrical Equipment	SF ₆	31.23	31.23	0.01	0.96
Indirect CO ₂ Emissions from CH ₄ Oxidation	CO ₂	30.90	30.90	0.01	0.97
CH ₄ Emissions from Manure Management	CH ₄	29.19	29.19	<0.01	0.97
CH ₄ Emissions from Wastewater Handling	CH ₄	24.25	24.25	<0.01	0.98
PFC Emissions from Aluminum Production	PFCs	18.11	18.11	<0.01	0.98
N ₂ O Emissions from Nitric Acid Production	N ₂ O	17.85	17.85	<0.01	0.98
N ₂ O Emissions from Manure Management	N ₂ O	16.03	16.03	<0.01	0.98
Non-CO ₂ Emissions from Stationary Combustion	N ₂ O	12.82	12.82	<0.01	0.99
N ₂ O Emissions from Adipic Acid Production	N ₂ O	14.89	14.89	<0.01	0.99
CO ₂ Emissions from Waste Incineration	CO ₂	14.09	14.09	<0.01	0.99
CO ₂ Emissions from Lime Production	CO ₂	11.24	11.24	<0.01	0.99
Non-CO ₂ Emissions from Stationary Combustion	CH ₄	7.90	7.90	<0.01	0.99
CH ₄ Emissions from Rice Production	CH ₄	7.12	7.12	<0.01	0.99
N ₂ O Emissions from Wastewater Handling	N ₂ O	7.04	7.04	<0.01	1.00
SF ₆ Emissions from Magnesium Production	SF ₆	5.50	5.50	<0.01	1.00
CO ₂ Emissions from Natural Gas Flaring	CO ₂	5.51	5.51	<0.01	1.00
Mobile Combustion: Road & Other	CH ₄	4.67	4.67	<0.01	1.00

PFC, HFC, and SF ₆ Emissions from Semiconductor Manufacturing					
	SF ₆	2.86	2.86	<0.01	1.00
Mobile Combustion: Aviation					
	N ₂ O	1.71	1.71	<0.01	1.00
CH ₄ Emissions from Other Industrial Processes					
	CH ₄	1.19	1.19	<0.01	1.00
Emissions from Substitutes for Ozone Depleting Substances					
	Several	0.94	0.94	<0.01	1.00
CH ₄ Emissions from Agricultural Residue Burning					
	CH ₄	0.68	0.68	<0.01	1.00
N ₂ O Emissions from Agricultural Residue Burning					
	N ₂ O	0.37	0.37	<0.01	1.00
Mobile Combustion: Marine					
	N ₂ O	0.36	0.36	<0.01	1.00
N ₂ O Emissions from Waste Incineration					
	N ₂ O	0.29	0.29	<0.01	1.00
CO ₂ Emissions from Stationary Combustion - Geothermal Energy					
	CO ₂	0.22	0.22	<0.01	1.00
Mobile Combustion: Aviation					
	CH ₄	0.16	0.16	<0.01	1.00
Mobile Combustion: Marine					
	CH ₄	0.07	0.07	<0.01	1.00
TOTAL		6,130.72	6,130.72	1.00	

Note: Sinks (e.g., LUCF, Landfill Carbon Storage) are not included in this analysis.

Table P-4: 1991 Key Source Tier 1 Analysis - Level Assessment

IPCC Source Categories	Direct Greenhouse Gas	Base Year	Current	Level Assessment	Cumulative Total
		Estimate (Tg CO ₂ Eq.)	Year Estimate (Tg CO ₂ Eq.)		
CO ₂ Emissions from Stationary Combustion - Coal	CO ₂	1,692.60	1,684.05	0.28	0.28
Mobile Combustion: Road & Other	CO ₂	1,235.49	1,215.58	0.20	0.48
CO ₂ Emissions from Stationary Combustion - Gas	CO ₂	952.76	973.29	0.16	0.64
CO ₂ Emissions from Stationary Combustion - Oil	CO ₂	662.46	637.75	0.10	0.74
CH ₄ Emissions from Solid Waste Disposal Sites	CH ₄	213.41	213.16	0.04	0.78
Direct N ₂ O Emissions from Agricultural Soils	N ₂ O	193.49	195.30	0.03	0.81
Mobile Combustion: Aviation	CO ₂	176.88	169.34	0.03	0.84
Fugitive Emissions from Oil & Gas Operations	CH ₄	147.64	149.51	0.02	0.86
CH ₄ Emissions from Enteric Fermentation in Domestic Livestock	CH ₄	127.88	127.23	0.02	0.88
CO ₂ Emissions from Other Industrial Processes	CO ₂	123.65	114.51	0.02	0.90
Fugitive Emissions from Coal Mining and Handling	CH ₄	87.12	83.68	0.01	0.92
Indirect N ₂ O Emissions from Nitrogen Used in Agriculture	N ₂ O	73.60	74.76	0.01	0.93
Mobile Combustion: Marine	CO ₂	59.43	52.82	0.01	0.94
Mobile Combustion: Road & Other	N ₂ O	48.86	51.21	0.01	0.95
CO ₂ Emissions from Cement Production	CO ₂	33.28	32.54	0.01	0.95
SF ₆ Emissions from Electrical Equipment	SF ₆	31.23	32.48	0.01	0.96
HFC-23 Emissions from HCFC-22 Manufacture	HFCs	34.98	30.77	0.01	0.96
Indirect CO ₂ Emissions from CH ₄ Oxidation	CO ₂	30.90	30.70	0.01	0.97
CH ₄ Emissions from Manure Management	CH ₄	29.19	31.14	0.01	0.97
CH ₄ Emissions from Wastewater Handling	CH ₄	24.25	24.60	<0.01	0.98
N ₂ O Emissions from Nitric Acid Production	N ₂ O	17.85	17.83	<0.01	0.98
N ₂ O Emissions from Manure Management	N ₂ O	16.03	16.53	<0.01	0.98
CO ₂ Emissions from Waste Incineration	CO ₂	14.09	15.78	<0.01	0.98
PFC Emissions from Aluminum Production	PFCs	18.11	15.68	<0.01	0.99
Non-CO ₂ Emissions from Stationary Combustion	N ₂ O	12.82	12.68	<0.01	0.99
N ₂ O Emissions from Adipic Acid Production	N ₂ O	14.89	14.69	<0.01	0.99
CO ₂ Emissions from Lime Production	CO ₂	11.24	11.01	<0.01	0.99
Non-CO ₂ Emissions from Stationary Combustion	CH ₄	7.90	8.03	<0.01	0.99
N ₂ O Emissions from Wastewater Handling	N ₂ O	7.04	7.21	<0.01	0.99

CH ₄ Emissions from Rice Production	CH ₄	7.12	7.00	<0.01	1.00
CO ₂ Emissions from Natural Gas Flaring	CO ₂	5.51	5.59	<0.01	1.00
SF ₆ Emissions from Magnesium Production	SF ₆	5.50	5.50	<0.01	1.00
Mobile Combustion: Road & Other	CH ₄	4.67	4.64	<0.01	1.00
PFC, HFC, and SF ₆ Emissions from Semiconductor Manufacturing	SF ₆	2.86	2.86	<0.01	1.00
Mobile Combustion: Aviation	N ₂ O	1.71	1.64	<0.01	1.00
CH ₄ Emissions from Other Industrial Processes	CH ₄	1.19	1.21	<0.01	1.00
Emissions from Substitutes for Ozone Depleting Substances	Several	0.94	0.84	<0.01	1.00
CH ₄ Emissions from Agricultural Residue Burning	CH ₄	0.68	0.64	<0.01	1.00
Mobile Combustion: Marine	N ₂ O	0.36	0.38	<0.01	1.00
N ₂ O Emissions from Agricultural Residue Burning	N ₂ O	0.37	0.36	<0.01	1.00
N ₂ O Emissions from Waste Incineration	N ₂ O	0.29	0.24	<0.01	1.00
CO ₂ Emissions from Stationary Combustion - Geothermal Energy	CO ₂	0.22	0.21	<0.01	1.00
Mobile Combustion: Aviation	CH ₄	0.16	0.15	<0.01	1.00
Mobile Combustion: Marine	CH ₄	0.07	0.07	<0.01	1.00
TOTAL		6,130.72	6,075.18	1.00	

Note: Sinks (e.g., LUCF, Landfill Carbon Storage) are not included in this analysis.

Table P-5: 1992 Key Source Tier 1 Analysis - Level Assessment

IPCC Source Categories	Direct Greenhouse Gas	Base Year	Current	Level Assessment	Cumulative Total
		Estimate (Tg CO ₂ Eq.)	Year Estimate (Tg CO ₂ Eq.)		
CO ₂ Emissions from Stationary Combustion - Coal	CO ₂	1,692.60	1,702.18	0.27	0.27
Mobile Combustion: Road & Other	CO ₂	1,235.49	1,238.29	0.20	0.47
CO ₂ Emissions from Stationary Combustion - Gas	CO ₂	952.76	1,002.81	0.16	0.64
CO ₂ Emissions from Stationary Combustion - Oil	CO ₂	662.46	657.85	0.11	0.74
CH ₄ Emissions from Solid Waste Disposal Sites	CH ₄	213.41	215.82	0.03	0.78
Direct N ₂ O Emissions from Agricultural Soils	N ₂ O	193.49	202.46	0.03	0.81
Mobile Combustion: Aviation	CO ₂	176.88	167.01	0.03	0.84
Fugitive Emissions from Oil & Gas Operations	CH ₄	147.64	150.35	0.02	0.86
CH ₄ Emissions from Enteric Fermentation in Domestic Livestock	CH ₄	127.88	130.16	0.02	0.88
CO ₂ Emissions from Other Industrial Processes	CO ₂	123.65	113.82	0.02	0.90
Fugitive Emissions from Coal Mining and Handling	CH ₄	87.12	81.37	0.01	0.91
Indirect N ₂ O Emissions from Nitrogen Used in Agriculture	N ₂ O	73.60	75.58	0.01	0.93
Mobile Combustion: Marine	CO ₂	59.43	67.68	0.01	0.94
Mobile Combustion: Road & Other	N ₂ O	48.86	54.35	0.01	0.95
HFC-23 Emissions from HCFC-22 Manufacture	HFCs	34.98	34.87	0.01	0.95
CO ₂ Emissions from Cement Production	CO ₂	33.28	32.79	0.01	0.96
Indirect CO ₂ Emissions from CH ₄ Oxidation	CO ₂	30.90	30.51	<0.01	0.96
SF ₆ Emissions from Electrical Equipment	SF ₆	31.23	30.16	<0.01	0.97
CH ₄ Emissions from Manure Management	CH ₄	29.19	30.72	<0.01	0.97
CH ₄ Emissions from Wastewater Handling	CH ₄	24.25	25.22	<0.01	0.98
N ₂ O Emissions from Nitric Acid Production	N ₂ O	17.85	18.30	<0.01	0.98
CO ₂ Emissions from Waste Incineration	CO ₂	14.09	16.32	<0.01	0.98
N ₂ O Emissions from Manure Management	N ₂ O	16.03	16.31	<0.01	0.98
Non-CO ₂ Emissions from Stationary Combustion	N ₂ O	12.82	12.94	<0.01	0.99
PFC Emissions from Aluminum Production	PFCs	18.11	14.55	<0.01	0.99
N ₂ O Emissions from Adipic Acid Production	N ₂ O	14.89	12.63	<0.01	0.99

CO ₂ Emissions from Lime Production	CO ₂	11.24	11.39	<0.01	0.99
Non-CO ₂ Emissions from Stationary Combustion	CH ₄	7.90	8.30	<0.01	0.99
CH ₄ Emissions from Rice Production	CH ₄	7.12	7.87	<0.01	0.99
N ₂ O Emissions from Wastewater Handling	N ₂ O	7.04	7.34	<0.01	1.00
CO ₂ Emissions from Natural Gas Flaring	CO ₂	5.51	5.06	<0.01	1.00
SF ₆ Emissions from Magnesium Production	SF ₆	5.50	5.50	<0.01	1.00
Mobile Combustion: Road & Other	CH ₄	4.67	4.67	<0.01	1.00
PFC, HFC, and SF ₆ Emissions from Semiconductor Manufacturing	SF ₆	2.86	2.86	<0.01	1.00
Mobile Combustion: Aviation	N ₂ O	1.71	1.62	<0.01	1.00
Emissions from Substitutes for Ozone Depleting Substances	Several	0.94	1.52	<0.01	1.00
CH ₄ Emissions from Other Industrial Processes	CH ₄	1.19	1.28	<0.01	1.00
CH ₄ Emissions from Agricultural Residue Burning	CH ₄	0.68	0.75	<0.01	1.00
Mobile Combustion: Marine	N ₂ O	0.36	0.43	<0.01	1.00
N ₂ O Emissions from Agricultural Residue Burning	N ₂ O	0.37	0.41	<0.01	1.00
N ₂ O Emissions from Waste Incineration	N ₂ O	0.29	0.27	<0.01	1.00
CO ₂ Emissions from Stationary Combustion - Geothermal Energy	CO ₂	0.22	0.21	<0.01	1.00
Mobile Combustion: Aviation	CH ₄	0.16	0.15	<0.01	1.00
Mobile Combustion: Marine	CH ₄	0.07	0.08	<0.01	1.00
TOTAL		6,130.72	6,194.77	1.00	

Note: Sinks (e.g., LUCF, Landfill Carbon Storage) are not included in this analysis.

Table P-6: 1993 Key Source Tier 1 Analysis - Level Assessment

IPCC Source Categories	Direct Greenhouse Gas	Base Year Estimate Current Year Estimate		Level Assessment	Cumulative Total
		(Tg CO ₂ Eq.)	(Tg CO ₂ Eq.)		
CO ₂ Emissions from Stationary Combustion – Coal	CO ₂	1,692.60	1,764.15	0.28	0.28
Mobile Combustion: Road & Other	CO ₂	1,235.49	1,271.16	0.20	0.48
CO ₂ Emissions from Stationary Combustion – Gas	CO ₂	952.76	1,036.28	0.16	0.65
CO ₂ Emissions from Stationary Combustion – Oil	CO ₂	662.46	647.74	0.10	0.75
CH ₄ Emissions from Solid Waste Disposal Sites	CH ₄	213.41	217.83	0.03	0.78
Direct N ₂ O Emissions from Agricultural Soils	N ₂ O	193.49	195.55	0.03	0.81
Mobile Combustion: Aviation	CO ₂	176.88	168.04	0.03	0.84
Fugitive Emissions from Oil & Gas Operations	CH ₄	147.64	154.01	0.02	0.87
CH ₄ Emissions from Enteric Fermentation in Domestic Livestock	CH ₄	127.88	128.47	0.02	0.89
CO ₂ Emissions from Other Industrial Processes	CO ₂	123.65	106.89	0.02	0.90
Indirect N ₂ O Emissions from Nitrogen Used in Agriculture	N ₂ O	73.60	77.49	0.01	0.92
Fugitive Emissions from Coal Mining and Handling	CH ₄	87.12	69.68	0.01	0.93
Mobile Combustion: Marine	CO ₂	59.43	63.38	0.01	0.94
Mobile Combustion: Road & Other	N ₂ O	48.86	56.45	0.01	0.95
CO ₂ Emissions from Cement Production	CO ₂	33.28	34.62	0.01	0.95
SF ₆ Emissions from Electrical Equipment	SF ₆	31.23	34.09	0.01	0.96
HFC-23 Emissions from HCFC-22 Manufacture	HFCs	34.98	31.82	0.01	0.96
Indirect CO ₂ Emissions from CH ₄ Oxidation	CO ₂	30.90	29.48	<0.01	0.97
CH ₄ Emissions from Manure Management	CH ₄	29.19	31.64	0.01	0.97
CH ₄ Emissions from Wastewater Handling	CH ₄	24.25	25.59	<0.01	0.97
N ₂ O Emissions from Nitric Acid Production	N ₂ O	17.85	18.57	<0.01	0.98
CO ₂ Emissions from Waste Incineration	CO ₂	14.09	17.18	<0.01	0.98

N ₂ O Emissions from Manure Management	N ₂ O	16.03	16.73	<0.01	0.98
Non-CO ₂ Emissions from Stationary Combustion	N ₂ O	12.82	13.14	<0.01	0.99
N ₂ O Emissions from Adipic Acid Production	N ₂ O	14.89	13.92	<0.01	0.99
PFC Emissions from Aluminum Production	PFCs	18.11	13.86	<0.01	0.99
CO ₂ Emissions from Lime Production	CO ₂	11.24	11.64	<0.01	0.99
CO ₂ Emissions from Natural Gas Flaring	CO ₂	5.51	6.55	<0.01	0.99
Non-CO ₂ Emissions from Stationary Combustion	CH ₄	7.90	7.82	<0.01	0.99
N ₂ O Emissions from Wastewater Handling	N ₂ O	7.04	7.45	<0.01	1.00
CH ₄ Emissions from Rice Production	CH ₄	7.12	7.02	<0.01	1.00
SF ₆ Emissions from Magnesium Production	SF ₆	5.50	5.37	<0.01	1.00
Emissions from Substitutes for Ozone Depleting Substances	Several	0.94	5.24	<0.01	1.00
Mobile Combustion: Road & Other	CH ₄	4.67	4.65	<0.01	1.00
PFC, HFC, and SF ₆ Emissions from Semiconductor Manufacturing	SF ₆	2.86	3.58	<0.01	1.00
Mobile Combustion: Aviation	N ₂ O	1.71	1.63	<0.01	1.00
CH ₄ Emissions from Other Industrial Processes	CH ₄	1.19	1.40	<0.01	1.00
CH ₄ Emissions from Agricultural Residue Burning	CH ₄	0.68	0.60	<0.01	1.00
Mobile Combustion: Marine	N ₂ O	0.36	0.43	<0.01	1.00
N ₂ O Emissions from Agricultural Residue Burning	N ₂ O	0.37	0.34	<0.01	1.00
N ₂ O Emissions from Waste Incineration	N ₂ O	0.29	0.26	<0.01	1.00
CO ₂ Emissions from Stationary Combustion - Geothermal Energy	CO ₂	0.22	0.19	<0.01	1.00
Mobile Combustion: Aviation	CH ₄	0.16	0.14	<0.01	1.00
Mobile Combustion: Marine	CH ₄	0.07	0.08	<0.01	1.00
TOTAL		6,130.72	6,302.16	1.00	

Note: Sinks (e.g., LUCF, Landfill Carbon Storage) are not included in this analysis.

Table P-7: 1994 Key Source Tier 1 Analysis - Level Assessment

IPCC Source Categories	Direct Greenhouse Gas	Base Year		Level Assessment	Cumulative Total
		Estimate (Tg CO ₂ Eq.)	Current Year Estimate (Tg CO ₂ Eq.)		
CO ₂ Emissions from Stationary Combustion - Coal	CO ₂	1,692.60	1,782.55	0.28	0.28
Mobile Combustion: Road & Other	CO ₂	1,235.49	1,313.34	0.20	0.48
CO ₂ Emissions from Stationary Combustion - Gas	CO ₂	952.76	1,056.23	0.16	0.65
CO ₂ Emissions from Stationary Combustion - Oil	CO ₂	662.46	656.85	0.10	0.75
CH ₄ Emissions from Solid Waste Disposal Sites	CH ₄	213.41	217.77	0.03	0.78
Direct N ₂ O Emissions from Agricultural Soils	N ₂ O	193.49	215.16	0.03	0.81
Mobile Combustion: Aviation	CO ₂	176.88	175.85	0.03	0.84
Fugitive Emissions from Oil & Gas Operations	CH ₄	147.64	151.91	0.02	0.87
CH ₄ Emissions from Enteric Fermentation in Domestic Livestock	CH ₄	127.88	130.12	0.02	0.89
CO ₂ Emissions from Other Industrial Processes	CO ₂	123.65	111.86	0.02	0.90
Indirect N ₂ O Emissions from Nitrogen Used in Agriculture	N ₂ O	73.60	79.95	0.01	0.92
Fugitive Emissions from Coal Mining and Handling	CH ₄	87.12	70.32	0.01	0.93
Mobile Combustion: Marine	CO ₂	59.43	62.18	0.01	0.94
Mobile Combustion: Road & Other	N ₂ O	48.86	57.88	0.01	0.95
CO ₂ Emissions from Cement Production	CO ₂	33.28	36.09	0.01	0.95
HFC-23 Emissions from HCFC-22 Manufacture	HFCs	34.98	31.59	<0.01	0.96
SF ₆ Emissions from Electrical Equipment	SF ₆	31.23	31.42	<0.01	0.96
CH ₄ Emissions from Manure Management	CH ₄	29.19	33.82	0.01	0.97
Indirect CO ₂ Emissions from CH ₄ Oxidation	CO ₂	30.90	29.30	<0.01	0.97

CH ₄ Emissions from Wastewater Handling	CH ₄	24.25	26.21	<0.01	0.97
N ₂ O Emissions from Nitric Acid Production	N ₂ O	17.85	19.60	<0.01	0.98
CO ₂ Emissions from Waste Incineration	CO ₂	14.09	17.87	<0.01	0.98
N ₂ O Emissions from Manure Management	N ₂ O	16.03	16.72	<0.01	0.98
Non-CO ₂ Emissions from Stationary Combustion	N ₂ O	12.82	13.38	<0.01	0.98
N ₂ O Emissions from Adipic Acid Production	N ₂ O	14.89	15.39	<0.01	0.99
PFC Emissions from Aluminum Production	PFCs	18.11	12.17	<0.01	0.99
CO ₂ Emissions from Lime Production	CO ₂	11.24	12.05	<0.01	0.99
CO ₂ Emissions from Natural Gas Flaring	CO ₂	5.51	6.62	<0.01	0.99
Emissions from Substitutes for Ozone Depleting Substances	Several	0.94	8.45	<0.01	0.99
CH ₄ Emissions from Rice Production	CH ₄	7.12	8.21	<0.01	0.99
Non-CO ₂ Emissions from Stationary Combustion	CH ₄	7.90	7.82	<0.01	1.00
N ₂ O Emissions from Wastewater Handling	N ₂ O	7.04	7.74	<0.01	1.00
SF ₆ Emissions from Magnesium Production	SF ₆	5.50	5.19	<0.01	1.00
Mobile Combustion: Road & Other	CH ₄	4.67	4.61	<0.01	1.00
PFC, HFC, and SF ₆ Emissions from Semiconductor Manufacturing	SF ₆	2.86	3.93	<0.01	1.00
Mobile Combustion: Aviation	N ₂ O	1.71	1.71	<0.01	1.00
CH ₄ Emissions from Other Industrial Processes	CH ₄	1.19	1.50	<0.01	1.00
CH ₄ Emissions from Agricultural Residue Burning	CH ₄	0.68	0.81	<0.01	1.00
N ₂ O Emissions from Agricultural Residue Burning	N ₂ O	0.37	0.45	<0.01	1.00
Mobile Combustion: Marine	N ₂ O	0.36	0.41	<0.01	1.00
N ₂ O Emissions from Waste Incineration	N ₂ O	0.29	0.27	<0.01	1.00
CO ₂ Emissions from Stationary Combustion - Geothermal Energy	CO ₂	0.22	0.18	<0.01	1.00
Mobile Combustion: Aviation	CH ₄	0.16	0.15	<0.01	1.00
Mobile Combustion: Marine	CH ₄	0.07	0.08	<0.01	1.00
TOTAL		6,130.72	6,435.71	1.00	

Note: Sinks (e.g., LUCF, Landfill Carbon Storage) are not included in this analysis.

Table P-8: 1995 Key Source Tier 1 Analysis - Level Assessment

IPCC Source Categories	Direct Greenhouse Gas	Base Year	Current	Level Assessment	Cumulative Total
		Estimate (Tg CO ₂ Eq.)	Year Estimate (Tg CO ₂ Eq.)		
CO ₂ Emissions from Stationary Combustion - Coal	CO ₂	1,692.60	1,792.70	0.28	0.28
Mobile Combustion: Road & Other	CO ₂	1,235.49	1,340.99	0.21	0.48
CO ₂ Emissions from Stationary Combustion - Gas	CO ₂	952.76	1,102.94	0.17	0.65
CO ₂ Emissions from Stationary Combustion - Oil	CO ₂	662.46	609.93	0.09	0.75
CH ₄ Emissions from Solid Waste Disposal Sites	CH ₄	213.41	216.62	0.03	0.78
Direct N ₂ O Emissions from Agricultural Soils	N ₂ O	193.49	204.77	0.03	0.81
Mobile Combustion: Aviation	CO ₂	176.88	171.45	0.03	0.84
Fugitive Emissions from Oil & Gas Operations	CH ₄	147.64	149.89	0.02	0.86
CH ₄ Emissions from Enteric Fermentation in Domestic Livestock	CH ₄	127.88	133.18	0.02	0.88
CO ₂ Emissions from Other Industrial Processes	CO ₂	123.65	114.41	0.02	0.90
Indirect N ₂ O Emissions from Nitrogen Used in Agriculture	N ₂ O	73.60	78.64	0.01	0.91
Fugitive Emissions from Coal Mining and Handling	CH ₄	87.12	73.54	0.01	0.92
Mobile Combustion: Marine	CO ₂	59.43	66.92	0.01	0.93
Mobile Combustion: Road & Other	N ₂ O	48.86	58.28	0.01	0.94
CO ₂ Emissions from Cement Production	CO ₂	33.28	36.85	0.01	0.95
CH ₄ Emissions from Manure Management	CH ₄	29.19	34.79	0.01	0.95

Indirect CO ₂ Emissions from CH ₄ Oxidation	CO ₂	30.90	29.46	<0.01	0.96
HFC-23 Emissions from HCFC-22 Manufacture	HFCs	34.98	27.03	<0.01	0.96
CH ₄ Emissions from Wastewater Handling	CH ₄	24.25	26.79	<0.01	0.97
SF ₆ Emissions from Electrical Equipment	SF ₆	31.23	26.49	<0.01	0.97
Emissions from Substitutes for Ozone Depleting Substances	Several	0.94	21.82	<0.01	0.97
N ₂ O Emissions from Nitric Acid Production	N ₂ O	17.85	19.89	<0.01	0.98
CO ₂ Emissions from Waste Incineration	CO ₂	14.09	18.61	<0.01	0.98
N ₂ O Emissions from Adipic Acid Production	N ₂ O	14.89	17.88	<0.01	0.98
N ₂ O Emissions from Manure Management	N ₂ O	16.03	16.37	<0.01	0.99
Non-CO ₂ Emissions from Stationary Combustion	N ₂ O	12.82	13.48	<0.01	0.99
CO ₂ Emissions from Natural Gas Flaring	CO ₂	5.51	8.73	<0.01	0.99
CO ₂ Emissions from Lime Production	CO ₂	11.24	12.80	<0.01	0.99
PFC Emissions from Aluminum Production	PFCs	18.11	11.81	<0.01	0.99
Non-CO ₂ Emissions from Stationary Combustion	CH ₄	7.90	8.23	<0.01	0.99
N ₂ O Emissions from Wastewater Handling	N ₂ O	7.04	7.69	<0.01	1.00
CH ₄ Emissions from Rice Production	CH ₄	7.12	7.62	<0.01	1.00
PFC, HFC, and SF ₆ Emissions from Semiconductor Manufacturing	SF ₆	2.86	5.90	<0.01	1.00
SF ₆ Emissions from Magnesium Production	SF ₆	5.50	5.49	<0.01	1.00
Mobile Combustion: Road & Other	CH ₄	4.67	4.54	<0.01	1.00
Mobile Combustion: Aviation	N ₂ O	1.71	1.67	<0.01	1.00
CH ₄ Emissions from Other Industrial Processes	CH ₄	1.19	1.53	<0.01	1.00
CH ₄ Emissions from Agricultural Residue Burning	CH ₄	0.68	0.66	<0.01	1.00
Mobile Combustion: Marine	N ₂ O	0.36	0.46	<0.01	1.00
N ₂ O Emissions from Agricultural Residue Burning	N ₂ O	0.37	0.38	<0.01	1.00
N ₂ O Emissions from Waste Incineration	N ₂ O	0.29	0.28	<0.01	1.00
Mobile Combustion: Aviation	CH ₄	0.16	0.15	<0.01	1.00
CO ₂ Emissions from Stationary Combustion - Geothermal Energy	CO ₂	0.22	0.12	<0.01	1.00
Mobile Combustion: Marine	CH ₄	0.07	0.09	<0.01	1.00
TOTAL		6,130.72	6,481.81	1.00	

Note: Sinks (e.g., LUCF, Landfill Carbon Storage) are not included in this analysis.

Table P-9: 1996 Key Source Tier 1 Analysis - Level Assessment

IPCC Source Categories	Direct Greenhouse Gas	Base Year	Current Year	Level Assessment	Cumulative Total
		Estimate (Tg CO ₂ Eq.)	Estimate (Tg CO ₂ Eq.)		
CO ₂ Emissions from Stationary Combustion - Coal	CO ₂	1,692.60	1,878.38	0.28	0.28
Mobile Combustion: Road & Other	CO ₂	1,235.49	1,374.75	0.21	0.49
CO ₂ Emissions from Stationary Combustion - Gas	CO ₂	952.76	1,123.44	0.17	0.66
CO ₂ Emissions from Stationary Combustion - Oil	CO ₂	662.46	645.98	0.10	0.75
Direct N ₂ O Emissions from Agricultural Soils	N ₂ O	193.49	212.29	0.03	0.78
CH ₄ Emissions from Solid Waste Disposal Sites	CH ₄	213.41	211.50	0.03	0.82
Mobile Combustion: Aviation	CO ₂	176.88	180.16	0.03	0.84
Fugitive Emissions from Oil & Gas Operations	CH ₄	147.64	150.67	0.02	0.87
CH ₄ Emissions from Enteric Fermentation in Domestic Livestock	CH ₄	127.88	129.58	0.02	0.89
CO ₂ Emissions from Other Industrial Processes	CO ₂	123.65	109.78	0.02	0.90
Indirect N ₂ O Emissions from Nitrogen Used in Agriculture	N ₂ O	73.60	80.28	0.01	0.91
Fugitive Emissions from Coal Mining and Handling	CH ₄	87.12	68.36	0.01	0.92

Mobile Combustion: Marine	CO ₂	59.43	63.78	0.01	0.93
Mobile Combustion: Road & Other	N ₂ O	48.86	57.94	0.01	0.94
CO ₂ Emissions from Cement Production	CO ₂	33.28	37.08	0.01	0.95
HFC-23 Emissions from HCFC-22 Manufacture	HFCs	34.98	31.12	<0.01	0.95
CH ₄ Emissions from Manure Management	CH ₄	29.19	34.20	0.01	0.96
Emissions from Substitutes for Ozone Depleting Substances	Several	0.94	30.62	<0.01	0.96
Indirect CO ₂ Emissions from CH ₄ Oxidation	CO ₂	30.90	28.89	<0.01	0.97
CH ₄ Emissions from Wastewater Handling	CH ₄	24.25	27.04	<0.01	0.97
SF ₆ Emissions from Electrical Equipment	SF ₆	31.23	26.77	<0.01	0.97
N ₂ O Emissions from Nitric Acid Production	N ₂ O	17.85	20.71	<0.01	0.98
CO ₂ Emissions from Waste Incineration	CO ₂	14.09	19.57	<0.01	0.98
N ₂ O Emissions from Adipic Acid Production	N ₂ O	14.89	17.75	<0.01	0.98
N ₂ O Emissions from Manure Management	N ₂ O	16.03	16.79	<0.01	0.99
Non-CO ₂ Emissions from Stationary Combustion	N ₂ O	12.82	14.06	<0.01	0.99
CO ₂ Emissions from Lime Production	CO ₂	11.24	13.49	<0.01	0.99
CO ₂ Emissions from Natural Gas Flaring	CO ₂	5.51	8.23	<0.01	0.99
PFC Emissions from Aluminum Production	PFCs	18.11	12.47	<0.01	0.99
Non-CO ₂ Emissions from Stationary Combustion	CH ₄	7.90	8.41	<0.01	0.99
N ₂ O Emissions from Wastewater Handling	N ₂ O	7.04	7.79	<0.01	1.00
CH ₄ Emissions from Rice Production	CH ₄	7.12	6.97	<0.01	1.00
SF ₆ Emissions from Magnesium Production	SF ₆	5.50	5.47	<0.01	1.00
PFC, HFC, and SF ₆ Emissions from Semiconductor Manufacturing	SF ₆	2.86	5.44	<0.01	1.00
Mobile Combustion: Road & Other	CH ₄	4.67	4.44	<0.01	1.00
Mobile Combustion: Aviation	N ₂ O	1.71	1.76	<0.01	1.00
CH ₄ Emissions from Other Industrial Processes	CH ₄	1.19	1.60	<0.01	1.00
CH ₄ Emissions from Agricultural Residue Burning	CH ₄	0.68	0.75	<0.01	1.00
N ₂ O Emissions from Agricultural Residue Burning	N ₂ O	0.37	0.42	<0.01	1.00
Mobile Combustion: Marine	N ₂ O	0.36	0.42	<0.01	1.00
N ₂ O Emissions from Waste Incineration	N ₂ O	0.29	0.28	<0.01	1.00
Mobile Combustion: Aviation	CH ₄	0.16	0.15	<0.01	1.00
CO ₂ Emissions from Stationary Combustion - Geothermal Energy	CO ₂	0.22	0.13	<0.01	1.00
Mobile Combustion: Marine	CH ₄	0.07	0.08	<0.01	1.00
TOTAL		6,130.72	6,669.76	1.00	

Note: Sinks (e.g., LUCF, Landfill Carbon Storage) are not included in this analysis.

Table P-10: 1997 Key Source Tier 1 Analysis - Level Assessment

IPCC Source Categories	Direct Greenhouse Gas	Base Year	Current	Level Assessment	Cumulative Total
		Estimate (Tg CO ₂ Eq.)	Year Estimate (Tg CO ₂ Eq.)		
CO ₂ Emissions from Stationary Combustion - Coal	CO ₂	1,692.60	1,930.51	0.29	0.29
Mobile Combustion: Road & Other	CO ₂	1,235.49	1,399.71	0.21	0.49
CO ₂ Emissions from Stationary Combustion - Gas	CO ₂	952.76	1,125.23	0.17	0.66
CO ₂ Emissions from Stationary Combustion - Oil	CO ₂	662.46	654.86	0.10	0.76
Direct N ₂ O Emissions from Agricultural Soils	N ₂ O	193.49	217.50	0.03	0.79
CH ₄ Emissions from Solid Waste Disposal Sites	CH ₄	213.41	206.37	0.03	0.82
Mobile Combustion: Aviation	CO ₂	176.88	178.94	0.03	0.85
Fugitive Emissions from Oil & Gas Operations	CH ₄	147.64	146.76	0.02	0.87
CH ₄ Emissions from Enteric Fermentation in Domestic Livestock	CH ₄	127.88	126.77	0.02	0.89

CO ₂ Emissions from Other Industrial Processes	CO ₂	123.65	119.15	0.02	0.90
Indirect N ₂ O Emissions from Nitrogen Used in Agriculture	N ₂ O	73.60	79.99	0.01	0.92
Fugitive Emissions from Coal Mining and Handling	CH ₄	87.12	68.13	0.01	0.93
Mobile Combustion: Road & Other	N ₂ O	48.86	57.61	0.01	0.94
Mobile Combustion: Marine	CO ₂	59.43	50.17	0.01	0.94
CO ₂ Emissions from Cement Production	CO ₂	33.28	38.32	0.01	0.95
Emissions from Substitutes for Ozone Depleting Substances	Several	0.94	37.99	0.01	0.95
CH ₄ Emissions from Manure Management	CH ₄	29.19	35.85	0.01	0.96
HFC-23 Emissions from HCFC-22 Manufacture	HFCs	34.98	29.99	<0.01	0.96
Indirect CO ₂ Emissions from CH ₄ Oxidation	CO ₂	30.90	28.35	<0.01	0.97
CH ₄ Emissions from Wastewater Handling	CH ₄	24.25	27.54	<0.01	0.97
SF ₆ Emissions from Electrical Equipment	SF ₆	31.23	24.49	<0.01	0.98
CO ₂ Emissions from Waste Incineration	CO ₂	14.09	21.34	<0.01	0.98
N ₂ O Emissions from Nitric Acid Production	N ₂ O	17.85	21.22	<0.01	0.98
N ₂ O Emissions from Manure Management	N ₂ O	16.03	17.08	<0.01	0.98
Non-CO ₂ Emissions from Stationary Combustion	N ₂ O	12.82	14.24	<0.01	0.99
CO ₂ Emissions from Lime Production	CO ₂	11.24	13.69	<0.01	0.99
CO ₂ Emissions from Natural Gas Flaring	CO ₂	5.51	7.57	<0.01	0.99
N ₂ O Emissions from Adipic Acid Production	N ₂ O	14.89	11.50	<0.01	0.99
PFC Emissions from Aluminum Production	PFCs	18.11	10.99	<0.01	0.99
N ₂ O Emissions from Wastewater Handling	N ₂ O	7.04	7.94	<0.01	0.99
Non-CO ₂ Emissions from Stationary Combustion	CH ₄	7.90	7.48	<0.01	1.00
CH ₄ Emissions from Rice Production	CH ₄	7.12	7.48	<0.01	1.00
SF ₆ Emissions from Magnesium Production	SF ₆	5.50	6.91	<0.01	1.00
PFC, HFC, and SF ₆ Emissions from Semiconductor Manufacturing	SF ₆	2.86	6.55	<0.01	1.00
Mobile Combustion: Road & Other	CH ₄	4.67	4.35	<0.01	1.00
Mobile Combustion: Aviation	N ₂ O	1.71	1.74	<0.01	1.00
CH ₄ Emissions from Other Industrial Processes	CH ₄	1.19	1.63	<0.01	1.00
CH ₄ Emissions from Agricultural Residue Burning	CH ₄	0.68	0.76	<0.01	1.00
N ₂ O Emissions from Agricultural Residue Burning	N ₂ O	0.37	0.45	<0.01	1.00
Mobile Combustion: Marine	N ₂ O	0.36	0.30	<0.01	1.00
N ₂ O Emissions from Waste Incineration	N ₂ O	0.29	0.26	<0.01	1.00
Mobile Combustion: Aviation	CH ₄	0.16	0.15	<0.01	1.00
CO ₂ Emissions from Stationary Combustion - Geothermal Energy	CO ₂	0.22	0.14	<0.01	1.00
Mobile Combustion: Marine	CH ₄	0.07	0.06	<0.01	1.00
TOTAL		6,130.72	6,748.06	1.00	

Note: Sinks (e.g., LUCF, Landfill Carbon Storage) are not included in this analysis.

Table P-11: 1998 Key Source Tier 1 Analysis - Level Assessment

IPCC Source Categories	Direct Greenhouse Gas	Base Year		Level Assessment	Cumulative Total
		Estimate (Tg CO ₂ Eq.)	Current Year Estimate (Tg CO ₂ Eq.)		
CO ₂ Emissions from Stationary Combustion - Coal	CO ₂	1,692.60	1,949.67	0.29	0.29
Mobile Combustion: Road & Other	CO ₂	1,235.49	1,424.19	0.21	0.50
CO ₂ Emissions from Stationary Combustion - Gas	CO ₂	952.76	1,090.90	0.16	0.66
CO ₂ Emissions from Stationary Combustion - Oil	CO ₂	662.46	660.47	0.10	0.76
Direct N ₂ O Emissions from Agricultural Soils	N ₂ O	193.49	218.59	0.03	0.79
CH ₄ Emissions from Solid Waste Disposal Sites	CH ₄	213.41	200.98	0.03	0.82

Mobile Combustion: Aviation	CO ₂	176.88	182.97	0.03	0.85
Fugitive Emissions from Oil & Gas Operations	CH ₄	147.64	145.61	0.02	0.87
CH ₄ Emissions from Enteric Fermentation in Domestic Livestock	CH ₄	127.88	124.90	0.02	0.89
CO ₂ Emissions from Other Industrial Processes	CO ₂	123.65	111.11	0.02	0.90
Indirect N ₂ O Emissions from Nitrogen Used in Agriculture	N ₂ O	73.60	79.81	0.01	0.92
Fugitive Emissions from Coal Mining and Handling	CH ₄	87.12	67.94	0.01	0.93
Mobile Combustion: Road & Other	N ₂ O	48.86	57.08	0.01	0.93
Mobile Combustion: Marine	CO ₂	59.43	47.83	0.01	0.94
Emissions from Substitutes for Ozone Depleting Substances	Several	0.94	44.87	0.01	0.95
HFC-23 Emissions from HCFC-22 Manufacture	HFCs	34.98	40.15	0.01	0.95
CO ₂ Emissions from Cement Production	CO ₂	33.28	39.22	0.01	0.96
CH ₄ Emissions from Manure Management	CH ₄	29.19	38.03	0.01	0.97
Indirect CO ₂ Emissions from CH ₄ Oxidation	CO ₂	30.90	28.18	<0.01	0.97
CH ₄ Emissions from Wastewater Handling	CH ₄	24.25	27.85	<0.01	0.97
N ₂ O Emissions from Nitric Acid Production	N ₂ O	17.85	20.89	<0.01	0.98
CO ₂ Emissions from Waste Incineration	CO ₂	14.09	20.25	<0.01	0.98
SF ₆ Emissions from Electrical Equipment	SF ₆	31.23	20.15	<0.01	0.98
Non-CO ₂ Emissions from Stationary Combustion	N ₂ O	12.82	14.32	<0.01	0.99
N ₂ O Emissions from Manure Management	N ₂ O	16.03	17.12	<0.01	0.99
CO ₂ Emissions from Lime Production	CO ₂	11.24	13.91	<0.01	0.99
PFC Emissions from Aluminum Production	PFCs	18.11	9.04	<0.01	0.99
N ₂ O Emissions from Wastewater Handling	N ₂ O	7.04	8.08	<0.01	0.99
CH ₄ Emissions from Rice Production	CH ₄	7.12	7.90	<0.01	0.99
N ₂ O Emissions from Adipic Acid Production	N ₂ O	14.89	7.71	<0.01	0.99
Non-CO ₂ Emissions from Stationary Combustion	CH ₄	7.90	7.01	<0.01	1.00
PFC, HFC, and SF ₆ Emissions from Semiconductor Manufacturing	SF ₆	2.86	7.26	<0.01	1.00
SF ₆ Emissions from Magnesium Production	SF ₆	5.50	6.18	<0.01	1.00
Mobile Combustion: Road & Other	CH ₄	4.67	4.25	<0.01	1.00
CO ₂ Emissions from Natural Gas Flaring	CO ₂	5.51	6.25	<0.01	1.00
Mobile Combustion: Aviation	N ₂ O	1.71	1.78	<0.01	1.00
CH ₄ Emissions from Other Industrial Processes	CH ₄	1.19	1.66	<0.01	1.00
CH ₄ Emissions from Agricultural Residue Burning	CH ₄	0.68	0.78	<0.01	1.00
N ₂ O Emissions from Agricultural Residue Burning	N ₂ O	0.37	0.45	<0.01	1.00
Mobile Combustion: Marine	N ₂ O	0.36	0.26	<0.01	1.00
N ₂ O Emissions from Waste Incineration	N ₂ O	0.29	0.24	<0.01	1.00
Mobile Combustion: Aviation	CH ₄	0.16	0.15	<0.01	1.00
CO ₂ Emissions from Stationary Combustion - Geothermal Energy	CO ₂	0.22	0.13	<0.01	1.00
Mobile Combustion: Marine	CH ₄	0.07	0.05	<0.01	1.00
TOTAL		6,130.72	6,756.19	1.00	

Note: Sinks (e.g., LUCF, Landfill Carbon Storage) are not included in this analysis.

Table P-12: 1999 Key Source Tier 1 Analysis - Level Assessment

IPCC Source Categories	Direct Greenhouse Gas	Base Year		Level Assessment	Cumulative Total
		Estimate (Tg CO ₂ Eq.)	Current Year Estimate (Tg CO ₂ Eq.)		
CO ₂ Emissions from Stationary Combustion - Coal	CO ₂	1,692.60	1,956.91	0.29	0.29
Mobile Combustion: Road & Other	CO ₂	1,235.49	1,478.53	0.22	0.50

CO ₂ Emissions from Stationary Combustion - Gas	CO ₂	952.76	1,105.01	0.16	0.66
CO ₂ Emissions from Stationary Combustion - Oil	CO ₂	662.46	658.39	0.10	0.76
Direct N ₂ O Emissions from Agricultural Soils	N ₂ O	193.49	216.49	0.03	0.79
CH ₄ Emissions from Solid Waste Disposal Sites	CH ₄	213.41	203.09	0.03	0.82
Mobile Combustion: Aviation	CO ₂	176.88	186.69	0.03	0.85
Fugitive Emissions from Oil & Gas Operations	CH ₄	147.64	140.84	0.02	0.87
CH ₄ Emissions from Enteric Fermentation in Domestic Livestock	CH ₄	127.88	124.51	0.02	0.89
CO ₂ Emissions from Other Industrial Processes	CO ₂	123.65	107.90	0.02	0.90
Indirect N ₂ O Emissions from Nitrogen Used in Agriculture	N ₂ O	73.60	79.79	0.01	0.92
Fugitive Emissions from Coal Mining and Handling	CH ₄	87.12	63.69	0.01	0.93
Mobile Combustion: Marine	CO ₂	59.43	63.02	0.01	0.93
Mobile Combustion: Road & Other	N ₂ O	48.86	56.44	0.01	0.94
Emissions from Substitutes for Ozone Depleting Substances	Several	0.94	51.28	0.01	0.95
CO ₂ Emissions from Cement Production	CO ₂	33.28	39.99	0.01	0.96
CH ₄ Emissions from Manure Management	CH ₄	29.19	37.56	0.01	0.96
HFC-23 Emissions from HCFC-22 Manufacture	HFCs	34.98	30.41	<0.01	0.97
CH ₄ Emissions from Wastewater Handling	CH ₄	24.25	28.35	<0.01	0.97
Indirect CO ₂ Emissions from CH ₄ Oxidation	CO ₂	30.90	27.00	<0.01	0.97
CO ₂ Emissions from Waste Incineration	CO ₂	14.09	21.84	<0.01	0.98
N ₂ O Emissions from Nitric Acid Production	N ₂ O	17.85	20.12	<0.01	0.98
Non-CO ₂ Emissions from Stationary Combustion	N ₂ O	12.82	14.60	<0.01	0.98
N ₂ O Emissions from Manure Management	N ₂ O	16.03	17.15	<0.01	0.99
SF ₆ Emissions from Electrical Equipment	SF ₆	31.23	15.51	<0.01	0.99
CO ₂ Emissions from Lime Production	CO ₂	11.24	13.47	<0.01	0.99
PFC Emissions from Aluminum Production	PFCs	18.11	8.94	<0.01	0.99
N ₂ O Emissions from Wastewater Handling	N ₂ O	7.04	8.36	<0.01	0.99
CH ₄ Emissions from Rice Production	CH ₄	7.12	8.29	<0.01	0.99
PFC, HFC, and SF ₆ Emissions from Semiconductor Manufacturing	SF ₆	2.86	7.73	<0.01	0.99
Non-CO ₂ Emissions from Stationary Combustion	CH ₄	7.90	7.34	<0.01	1.00
N ₂ O Emissions from Adipic Acid Production	N ₂ O	14.89	7.68	<0.01	1.00
SF ₆ Emissions from Magnesium Production	SF ₆	5.50	6.11	<0.01	1.00
Mobile Combustion: Road & Other	CH ₄	4.67	4.15	<0.01	1.00
CO ₂ Emissions from Natural Gas Flaring	CO ₂	5.51	6.68	<0.01	1.00
Mobile Combustion: Aviation	N ₂ O	1.71	1.82	<0.01	1.00
CH ₄ Emissions from Other Industrial Processes	CH ₄	1.19	1.68	<0.01	1.00
CH ₄ Emissions from Agricultural Residue Burning	CH ₄	0.68	0.76	<0.01	1.00
N ₂ O Emissions from Agricultural Residue Burning	N ₂ O	0.37	0.44	<0.01	1.00
Mobile Combustion: Marine	N ₂ O	0.36	0.43	<0.01	1.00
N ₂ O Emissions from Waste Incineration	N ₂ O	0.29	0.23	<0.01	1.00
Mobile Combustion: Aviation	CH ₄	0.16	0.15	<0.01	1.00
Mobile Combustion: Marine	CH ₄	0.07	0.08	<0.01	1.00
CO ₂ Emissions from Stationary Combustion - Geothermal Energy	CO ₂	0.22	0.04	<0.01	1.00
TOTAL		6,130.72	6,829.49	1.00	

Note: Sinks (e.g., LUCF, Landfill Carbon Storage) are not included in this analysis.

Table P-13: 2000 Key Source Tier 1 Analysis - Level Assessment

IPCC Source Categories	Direct Greenhouse	Base Year Estimate	Current Year Estimate	Level Assessment	Cumulative Total
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	Gas	(Tg CO ₂	(Tg CO ₂ Eq.)		
		Eq.)			
CO ₂ Emissions from Stationary Combustion - Coal	CO ₂	1692.60	2,030.09	0.29	0.29
Mobile Combustion: Road & Other	CO ₂	1235.49	1,503.16	0.21	0.50
CO ₂ Emissions from Stationary Combustion - Gas	CO ₂	952.76	1,162.91	0.17	0.67
CO ₂ Emissions from Stationary Combustion - Oil	CO ₂	662.46	640.72	0.09	0.76
Direct N ₂ O Emissions from Agricultural Soils	N ₂ O	193.49	217.75	0.03	0.79
CH ₄ Emissions from Solid Waste Disposal Sites	CH ₄	213.41	203.48	0.03	0.82
Mobile Combustion: Aviation	CO ₂	176.88	196.45	0.03	0.85
Fugitive Emissions from Oil & Gas Operations	CH ₄	147.64	138.22	0.02	0.87
CH ₄ Emissions from Enteric Fermentation in Domestic Livestock	CH ₄	127.88	123.86	0.02	0.89
CO ₂ Emissions from Other Industrial Processes	CO ₂	123.65	107.56	0.02	0.90
Mobile Combustion: Marine	CO ₂	59.43	89.91	0.01	0.92
Indirect N ₂ O Emissions from Nitrogen Used in Agriculture	N ₂ O	73.60	79.81	0.01	0.93
Fugitive Emissions from Coal Mining and Handling	CH ₄	87.12	60.97	0.01	0.94
Emissions from Substitutes for Ozone Depleting Substances	Several	0.94	57.78	0.01	0.94
Mobile Combustion: Road & Other	N ₂ O	48.86	55.74	0.01	0.95
CO ₂ Emissions from Cement Production	CO ₂	33.28	41.07	0.01	0.96
CH ₄ Emissions from Manure Management	CH ₄	29.19	37.46	0.01	0.96
HFC-23 Emissions from HCFC-22 Manufacture	HFCs	34.98	29.79	<0.01	0.97
CH ₄ Emissions from Wastewater Handling	CH ₄	24.25	28.70	<0.01	0.97
Indirect CO ₂ Emissions from CH ₄ Oxidation	CO ₂	30.90	26.30	<0.01	0.98
CO ₂ Emissions from Waste Incineration	CO ₂	14.09	22.47	<0.01	0.98
N ₂ O Emissions from Nitric Acid Production	N ₂ O	17.85	19.79	<0.01	0.98
Non-CO ₂ Emissions from Stationary Combustion	N ₂ O	12.82	14.93	<0.01	0.98
N ₂ O Emissions from Manure Management	N ₂ O	16.03	17.52	<0.01	0.99
SF ₆ Emissions from Electrical Equipment	SF ₆	31.23	14.45	<0.01	0.99
CO ₂ Emissions from Lime Production	CO ₂	11.24	13.32	<0.01	0.99
N ₂ O Emissions from Wastewater Handling	N ₂ O	7.04	8.46	<0.01	0.99
PFC Emissions from Aluminum Production	PFCs	18.11	7.95	<0.01	0.99
Non-CO ₂ Emissions from Stationary Combustion	CH ₄	7.90	7.50	<0.01	0.99
CH ₄ Emissions from Rice Production	CH ₄	7.12	7.50	<0.01	0.99
N ₂ O Emissions from Adipic Acid Production	N ₂ O	14.89	8.11	<0.01	1.00
PFC, HFC, and SF ₆ Emissions from Semiconductor Manufacturing	SF ₆	2.86	7.37	<0.01	1.00
Mobile Combustion: Road & Other	CH ₄	4.67	4.09	<0.01	1.00
SF ₆ Emissions from Magnesium Production	SF ₆	5.50	4.00	<0.01	1.00
CO ₂ Emissions from Natural Gas Flaring	CO ₂	5.51	6.06	<0.01	1.00
Mobile Combustion: Aviation	N ₂ O	1.71	1.92	<0.01	1.00
CH ₄ Emissions from Other Industrial Processes	CH ₄	1.19	1.67	<0.01	1.00
CH ₄ Emissions from Agricultural Residue Burning	CH ₄	0.68	0.79	<0.01	1.00
Mobile Combustion: Marine	N ₂ O	0.36	0.63	<0.01	1.00
N ₂ O Emissions from Agricultural Residue Burning	N ₂ O	0.37	0.46	<0.01	1.00
N ₂ O Emissions from Waste Incineration	N ₂ O	0.29	0.23	<0.01	1.00
Mobile Combustion: Aviation	CH ₄	0.16	0.16	<0.01	1.00
Mobile Combustion: Marine	CH ₄	0.07	0.12	<0.01	1.00
CO ₂ Emissions from Stationary Combustion - Geothermal Energy	CO ₂	0.22	0.02	<0.01	1.00
TOTAL		6,130.72	7,001.22	1.00	

Note: Sinks (e.g., LUCF, Landfill Carbon Storage) are not included in this analysis.

Table P-14: 1990-2000 Key Source Tier 1 Analysis - Trend Assessment

IPCC Source Categories	Direct Greenhouse Gas	Base Year		Current Year Estimate (Tg CO ₂ Eq.)	Trend Assessment	Percent Contribution to Trend	Cumulative Total
		Estimate (Tg CO ₂ Eq.)					
CO ₂ Emissions from Stationary Combustion - Oil	CO ₂	662.46		640.72	0.01	16	16
CO ₂ Emissions from Stationary Combustion - Coal	CO ₂	1,692.60		2,030.09	0.01	13	29
Mobile Combustion: Road & Other	CO ₂	1,235.49		1,503.16	0.01	13	42
CO ₂ Emissions from Stationary Combustion - Gas	CO ₂	952.76		1,162.91	0.01	10	52
Emissions from Substitutes for Ozone Depleting Substances	Several	0.94		57.78	0.01	8	60
CH ₄ Emissions from Solid Waste Disposal Sites	CH ₄	213.41		203.48	0.01	6	66
Fugitive Emissions from Coal Mining and Handling	CH ₄	87.12		60.97	<0.01	5	71
CO ₂ Emissions from Other Industrial Processes	CO ₂	123.65		107.56	<0.01	5	76
Fugitive Emissions from Oil & Gas Operations	CH ₄	147.64		138.22	<0.01	4	80
CH ₄ Emissions from Enteric Fermentation in Domestic Livestock	CH ₄	127.88		123.86	<0.01	3	83
Mobile Combustion: Marine	CO ₂	59.43		89.91	<0.01	3	86
SF ₆ Emissions from Electrical Equipment	SF ₆	31.23		14.45	<0.01	3	89
PFC Emissions from Aluminum Production	PFCs	18.11		7.95	<0.01	2	90
HFC-23 Emissions from HCFC-22 Manufacture	HFCs	34.98		29.79	<0.01	1	92
Indirect CO ₂ Emissions from CH ₄ Oxidation	CO ₂	30.90		26.30	<0.01	1	93
N ₂ O Emissions from Adipic Acid Production	N ₂ O	14.89		8.11	<0.01	1	94
CO ₂ Emissions from Waste Incineration	CO ₂	14.09		22.47	<0.01	1	95
Mobile Combustion: Aviation	CO ₂	176.88		196.45	<0.01	1	96
Indirect N ₂ O Emissions from Nitrogen Used in Agriculture	N ₂ O	73.60		79.81	<0.01	1	97
CH ₄ Emissions from Manure Management	CH ₄	29.19		37.46	<0.01	1	97
PFC, HFC, and SF ₆ Emissions from Semiconductor Manufacturing	SF ₆	2.86		7.37	<0.01	1	98
Direct N ₂ O Emissions from Agricultural Soils	N ₂ O	193.49		217.75	<0.01	0	98
CO ₂ Emissions from Cement Production	CO ₂	33.28		41.07	<0.01	0	99
SF ₆ Emissions from Magnesium Production	SF ₆	5.50		4.00	<0.01	0	99
Non-CO ₂ Emissions from Stationary Combustion	CH ₄	7.90		7.50	<0.01	0	99
Mobile Combustion: Road & Other	CH ₄	4.67		4.09	<0.01	0	99
CH ₄ Emissions from Wastewater Handling	CH ₄	24.25		28.70	<0.01	0	99
N ₂ O Emissions from Manure Management	N ₂ O	16.03		17.52	<0.01	0	99
CH ₄ Emissions from Rice Production	CH ₄	7.12		7.50	<0.01	0	100
N ₂ O Emissions from Nitric Acid Production	N ₂ O	17.85		19.79	<0.01	0	100
CO ₂ Emissions from Lime Production	CO ₂	11.24		13.32	<0.01	0	100
N ₂ O Emissions from Wastewater Handling	N ₂ O	7.04		8.46	<0.01	0	100
CH ₄ Emissions from Other Industrial Processes	CH ₄	1.19		1.67	<0.01	0	100

Non-CO ₂ Emissions from Stationary							
Combustion	N ₂ O	12.82	14.93	<0.01	0	100	
CO ₂ Emissions from Natural Gas Flaring	CO ₂	5.51	6.06	<0.01	0	100	
CO ₂ Emissions from Stationary Combustion							
- Geothermal Energy	CO ₂	0.22	0.02	<0.01	0	100	
Mobile Combustion: Marine	N ₂ O	0.36	0.63	<0.01	0	100	
N ₂ O Emissions from Waste Incineration	N ₂ O	0.29	0.23	<0.01	0	100	
Mobile Combustion: Road & Other	N ₂ O	48.86	55.74	<0.01	0	100	
Mobile Combustion: Marine	CH ₄	0.07	0.12	<0.01	0	100	
Mobile Combustion: Aviation	N ₂ O	1.71	1.92	<0.01	0	100	
N ₂ O Emissions from Agricultural Residue							
Burning	N ₂ O	0.37	0.46	<0.01	0	100	
Mobile Combustion: Aviation	CH ₄	0.16	0.16	<0.01	0	100	
CH ₄ Emissions from Agricultural Residue							
Burning	CH ₄	0.68	0.79	<0.01	0	100	
TOTAL		6,130.72	7,001.22	0.09	100		

Note: Sinks (e.g., LUCF, Landfill Carbon Storage) are not included in this analysis.

ANNEX Q: Global Warming Potential Values

Global Warming Potentials (GWPs) are intended as a quantified measure of the globally averaged relative radiative forcing impacts of a particular greenhouse gas. It is defined as the cumulative radiative forcing—both direct and indirect effects—integrated over a period of time from the emission of a unit mass of gas relative to some reference gas (IPCC 1996). Carbon dioxide (CO₂) was chosen as this reference gas. Direct effects occur when the gas itself is a greenhouse gas. Indirect radiative forcing occurs when chemical transformations involving the original gas produce a gas or gases that are greenhouse gases, or when a gas influences other radiatively important processes such as the atmospheric lifetimes of other gases. The relationship between gigagrams (Gg) of a gas and Tg CO₂ Eq. can be expressed as follows:

$$\text{Tg CO}_2 \text{ Eq} = (\text{Gg of gas}) \times (\text{GWP}) \times \left(\frac{\text{Tg}}{1,000 \text{ Gg}} \right)$$

where,

Tg CO₂ Eq. = Teragrams of Carbon Dioxide Equivalents

Gg = Gigagrams (equivalent to a thousand metric tons)

GWP = Global Warming Potential

Tg = Teragrams

GWP values allow policy makers to compare the impacts of emissions and reductions of different gases. According to the IPCC, GWPs typically have an uncertainty of roughly ±35 percent, though some GWPs have larger uncertainty than others, especially those in which lifetimes have not yet been ascertained. In the following decision, the parties to the UNFCCC have agreed to use consistent GWPs from the IPCC Second Assessment Report (SAR), based upon a 100 year time horizon, although other time horizon values are available (see Table Q-1).

In addition to communicating emissions in units of mass, Parties may choose also to use global warming potentials (GWPs) to reflect their inventories and projections in carbon dioxide-equivalent terms, using information provided by the Intergovernmental Panel on Climate Change (IPCC) in its Second Assessment Report. Any use of GWPs should be based on the effects of the greenhouse gases over a 100-year time horizon. In addition, Parties may also use other time horizons.¹

Greenhouse gases with relatively long atmospheric lifetimes (e.g., CO₂, CH₄, N₂O, HFCs, PFCs, and SF₆) tend to be evenly distributed throughout the atmosphere, and consequently global average concentrations can be determined. The short-lived gases such as water vapor, carbon monoxide, tropospheric ozone, other ambient air pollutants (e.g., NO_x, and NMVOCs), and tropospheric aerosols (e.g., SO₂ products and black carbon), however, vary spatially, and consequently it is difficult to quantify their global radiative forcing impacts. GWP values are generally not attributed to these gases that are short-lived and spatially inhomogeneous in the atmosphere.

Table Q-1: Global Warming Potentials (GWP) and Atmospheric Lifetimes (Years) of Gases Used in this Report

Gas	Atmospheric Lifetime	100-year GWP ^a	20-year GWP	500-year GWP
Carbon dioxide (CO ₂)	50-200	1	1	1
Methane (CH ₄) ^b	12±3	21	56	6.5
Nitrous oxide (N ₂ O)	120	310	280	170

¹ Framework Convention on Climate Change; FCCC/CP/1996/15/Add.1; 29 October 1996; Report of the Conference of the Parties at its second session; held at Geneva from 8 to 19 July 1996; Addendum; Part Two: Action taken by the Conference of the Parties at its second session; Decision 9/CP.2; Communications from Parties included in Annex I to the Convention: guidelines, schedule and process for consideration; Annex: Revised Guidelines for the Preparation of National Communications by Parties Included in Annex I to the Convention; p. 18. FCCC (1996)

HFC-23	264	11,700	9,100	9,800
HFC-125	32.6	2,800	4,600	920
HFC-134a	14.6	1,300	3,400	420
HFC-143a	48.3	3,800	5,000	1,400
HFC-152a	1.5	140	460	42
HFC-227ea	36.5	2,900	4,300	950
HFC-236fa	209	6,300	5,100	4,700
HFC-4310mee	17.1	1,300	3,000	400
CF ₄	50,000	6,500	4,400	10,000
C ₂ F ₆	10,000	9,200	6,200	14,000
C ₄ F ₁₀	2,600	7,000	4,800	10,100
C ₆ F ₁₄	3,200	7,400	5,000	10,700
SF ₆	3,200	23,900	16,300	34,900

Source: IPCC (1996)

^a GWPs used in this report are calculated over 100 year time horizon

^b The methane GWP includes the direct effects and those indirect effects due to the production of tropospheric ozone and stratospheric water vapor. The indirect effect due to the production of CO₂ is not included.

Table Q-2 presents direct and net (i.e., direct and indirect) GWPs for ozone-depleting substances (ODSs). Ozone-depleting substances directly absorb infrared radiation and contribute to positive radiative forcing; however, their effect as ozone-depleters also leads to a negative radiative forcing because ozone itself is a potent greenhouse gas. There is considerable uncertainty regarding this indirect effect; therefore, a range of net GWPs is provided for ozone depleting substances.

Table Q-2: Net 100-year Global Warming Potentials for Select Ozone Depleting Substances*

Gas	Direct	Net _{min}	Net _{max}
CFC-11	4,600	(600)	3,600
CFC-12	10,600	7,300	9,900
CFC-113	6,000	2,200	5,200
HCFC-22	1,700	1,400	1,700
HCFC-123	120	20	100
HCFC-124	620	480	590
HCFC-141b	700	(5)	570
HCFC-142b	2,400	1,900	2,300
CHCl ₃	140	(560)	0
CCl ₄	1,800	(3,900)	660
CH ₃ Br	5	(2,600)	(500)
Halon-1211	1,300	(24,000)	(3,600)
Halon-1301	6,900	(76,000)	(9,300)

Source: IPCC (2001)

* Because these compounds have been shown to deplete stratospheric ozone, they are typically referred to as ozone depleting substances (ODSs). However, they are also potent greenhouse gases. Recognizing the harmful effects of these compounds on the ozone layer, in 1987 many governments signed the *Montreal Protocol on Substances that Deplete the Ozone Layer* to limit the production and importation of a number of CFCs and other halogenated compounds. The United States furthered its commitment to phase-out ODSs by signing and ratifying the Copenhagen Amendments to the *Montreal Protocol* in 1992. Under these amendments, the United States committed to ending the production and importation of halons by 1994, and CFCs by 1996. The IPCC Guidelines and the UNFCCC do not include reporting instructions for estimating emissions of ODSs because their use is being phased-out under the *Montreal Protocol*. The effects of these compounds on radiative forcing are not addressed in this report.

The IPCC recently published its Third Assessment Report (TAR), providing the most current and comprehensive scientific assessment of climate change (IPCC 2001). Within this report, the GWPs of several gases were revised relative to the IPCC's Second Assessment Report (SAR) (IPCC 1996), and new GWPs have been calculated for an expanded set of gases. Since the SAR, the IPCC has applied an improved calculation of CO₂ radiative forcing and an improved CO₂ response function (presented in WMO 1999). The GWPs are drawn from WMO (1999) and the SAR, with updates for those cases where new laboratory or radiative transfer results have been published.

Additionally, the atmospheric lifetimes of some gases have been recalculated. Because the revised radiative forcing of CO₂ is about 12 percent lower than that in the SAR, the GWPs of the other gases relative to CO₂ tend to be larger, taking into account revisions in lifetimes. However, there were some instances in which other variables, such as the radiative efficiency or the chemical lifetime, were altered that resulted in further increases or decreases in particular GWP values. In addition, the values for radiative forcing and lifetimes have been calculated for a variety of halocarbons, which were not presented in the SAR. The changes are described in the TAR as follows:

New categories of gases include fluorinated organic molecules, many of which are ethers that are proposed as halocarbon substitutes. Some of the GWPs have larger uncertainties than that of others, particularly for those gases where detailed laboratory data on lifetimes are not yet available. The direct GWPs have been calculated relative to CO₂ using an improved calculation of the CO₂ radiative forcing, the SAR response function for a CO₂ pulse, and new values for the radiative forcing and lifetimes for a number of halocarbons.

Table Q-3 compares the lifetimes and GWPs for the SAR and TAR. As can be seen in Table Q-3, GWPs changed anywhere from a decrease of 15 percent to an increase of 49 percent.

Table Q-3: Comparison of GWPs and lifetimes used in the SAR and the TAR

Gas	Lifetime (years)		GWP (100 year)			
	SAR	TAR	SAR	TAR	Difference	
Carbon dioxide (CO₂)	50-200	5-200 ^a	1	1	NC	NC
Methane (CH₄)^b	12±3	8.4/12 ^c	21	23	2	10%
Nitrous oxide (N₂O)	120	120/114 ^e	310	296	(14)	-5%
Hydrofluorocarbons						
HFC-23	264	260	11,700	12,000	300	3%
HFC-32	5.6	5.0	650	550	(100)	-15%
HFC-41	3.7	2.6	150	97	(53)	-35%
HFC-125	32.6	29	2,800	3,400	600	21%
HFC-134	10.6	9.6	1,000	1,100	100	10%
HFC-134a	14.6	13.8	1,300	1,300	NC	NC
HFC-143	3.8	3.4	300	330	30	10%
HFC-143a	48.3	52	3,800	4,300	500	13%
HFC-152	NA	0.5	NA	43	NA	NA
HFC-152a	1.5	1.4	140	120	(20)	-14%
HFC-161	NA	0.3	NA	12	NA	NA
HFC-227ea	36.5	33.0	2,900	3,500	600	21%
HFC-236cb	NA	13.2	NA	1,300	NA	NA
HFC-236ea	NA	10	NA	1,200	NA	NA
HFC-236fa	209	220	6,300	9,400	3,100	49%
HFC-245ca	6.6	5.9	560	640	80	14%
HFC-245fa	NA	7.2	NA	950	NA	NA
HFC-365mfc	NA	9.9	NA	890	NA	NA
HFC-4310mee	17.1	15	1,300	1,500	200	15%
Iodocarbons						
FIC-1311	<0.005	0.005	<1	1	NC	NC
Fully Fluorinated Species						
SF ₆	3,200	3,200	23,900	22,000	(1,900)	-8%
CF ₄	50,000	50,000	6,500	5,700	(800)	-12%
C ₂ F ₆	10,000	10,000	9,200	11,900	2,700	29%
C ₃ F ₈	2,600	2,600	7,000	8,600	1,600	23%
C ₄ F ₁₀	2,600	2,600	7,000	8,600	1,600	23%
c-C ₄ F ₈	3,200	3,200	8,700	10,000	1,300	15%
C ₅ F ₁₂	4,100	4,100	7,500	8,900	1,400	19%
C ₆ F ₁₄	3,200	3,200	7,400	9,000	1,600	22%
Ethers and Halogenated Ethers						

CH ₃ OCH ₃	NA	0.015	NA	1	NA	NA
(CF ₃) ₂ CFOCH ₃	NA	3.4	NA	330	NA	NA
(CF ₃)CH ₂ OH	NA	0.5	NA	57	NA	NA
CF ₃ CF ₂ CH ₂ OH	NA	0.4	NA	40	NA	NA
(CF ₃) ₂ CHOH	NA	1.8	NA	190	NA	NA
HFE-125	NA	150	NA	14,900	NA	NA
HFE-134	NA	26.2	NA	6,100	NA	NA
HFE-143a	NA	4.4	NA	750	NA	NA
HCFE-235da2	NA	2.6	NA	340	NA	NA
HFE-245cb2	NA	4.3	NA	580	NA	NA
HFE-245fa2	NA	4.4	NA	570	NA	NA
HFE-254cb2	NA	0.22	NA	30	NA	NA
HFE-347mcc3	NA	4.5	NA	480	NA	NA
HFE-356pcf3	NA	3.2	NA	430	NA	NA
HFE-374pcf2	NA	5.0	NA	540	NA	NA
HFE-7100	NA	5.0	NA	390	NA	NA
HFE-7200	NA	0.77	NA	55	NA	NA
H-Galden 1040x	NA	6.3	NA	1,800	NA	NA
HG-10	NA	12.1	NA	2,700	NA	NA
HG-01	NA	6.2	NA	1,500	NA	NA
Others^d						
NF ₃	NA	740	NA	10,800	NA	NA
SF ₅ CF ₃	NA	>1,000	NA	>17,500	NA	NA
c-C ₃ F ₆	NA	>1,000	NA	>16,800	NA	NA
HFE-227ea	NA	11	NA	1,500	NA	NA
HFE-236ea2	NA	5.8	NA	960	NA	NA
HFE-236fa	NA	3.7	NA	470	NA	NA
HFE-245fa1	NA	2.2	NA	280	NA	NA
HFE-263fb2	NA	0.1	NA	11	NA	NA
HFE-329mcc2	NA	6.8	NA	890	NA	NA
HFE-338mcf2	NA	4.3	NA	540	NA	NA
HFE-347-mcf2	NA	2.8	NA	360	NA	NA
HFE-356mec3	NA	0.94	NA	98	NA	NA
HFE-356pcc3	NA	0.93	NA	110	NA	NA
HFE-356pcf2	NA	2.0	NA	260	NA	NA
HFE-365mcf3	NA	0.11	NA	11	NA	NA
(CF ₃) ₂ CHOCHF ₂	NA	3.1	NA	370	NA	NA
(CF ₃) ₂ CHOCH ₃	NA	0.25	NA	26	NA	NA
-(CF ₂) ₄ CH(OH)-	NA	0.85	NA	70	NA	NA

^a No single lifetime can be determined for carbon dioxide. (See IPCC 2001)

^b The methane GWP includes the direct effects and those indirect effects due to the production of tropospheric ozone and stratospheric water vapor. The indirect effect due to the production of CO₂ is not included.

^c Methane and nitrous oxide have chemical feedback systems that can alter the length of the atmospheric response, in these cases, global mean atmospheric lifetime (LT) is given first, followed by perturbation time (PT).

^d Gases whose lifetime has been determined only via indirect means or for whom there is uncertainty over the loss process.

Source: IPCC (2001)

NC (No Change)

NA (Not Applicable)

When the GWPs from the TAR are applied to the emission estimates presented in this report, total emissions for the year 2000 are 7,044.3 Tg CO₂ Eq., as compared to 7,001.2 Tg CO₂ Eq. when the GWPs from the SAR are used (0.6% difference). Table Q-4 provides a detailed summary of U.S. greenhouse gas emissions and sinks for 1990 through 2000, using the GWPs from the TAR. The adjusted greenhouse gas emissions are shown for each gas in units of Tg CO₂ Eq. in Table Q-5. The correlating percent change in emissions of each gas is shown in Table Q-6. The percent change in emissions is equal to the percent change in the GWP, however, in cases where multiple gases

are emitted in varying amounts the percent change is variable over the years, such as with substitutes for ozone depleting substances. Table Q-7 summarizes the emissions and resulting change in emissions using GWPs from the SAR or the TAR for 1990 and 2000.

Table Q-4: Recent Trends in U.S. Greenhouse Gas Emissions and Sinks using the TAR GWPs (Tg CO₂ Eq.)

Gas/Source	1990	1995	1996	1997	1998	1999	2000
CO₂	4,998.5	5,305.9	5,483.7	5,568.0	5,575.1	5,665.5	5,840.0
Fossil Fuel Combustion	4,779.8	5,085.0	5,266.6	5,339.6	5,356.2	5,448.6	5,623.3
Natural Gas Flaring	5.5	8.7	8.2	7.6	6.3	6.7	6.1
Cement Manufacture	33.3	36.8	37.1	38.3	39.2	40.0	41.1
Lime Manufacture	11.2	12.8	13.5	13.7	13.9	13.5	13.3
Limestone and Dolomite Use	5.2	7.0	7.4	8.4	8.2	9.1	9.2
Soda Ash Manufacture and Consumption	4.1	4.3	4.2	4.4	4.3	4.2	4.2
Carbon Dioxide Consumption	0.8	1.0	1.1	1.3	1.4	1.6	1.4
Waste Combustion	14.1	18.6	19.6	21.3	20.3	21.8	22.5
Titanium Dioxide Production	1.3	1.7	1.7	1.8	1.8	1.9	2.0
Aluminum Production	6.3	5.3	5.6	5.6	5.8	5.9	5.4
Iron and Steel Production	85.4	74.4	68.3	76.1	67.4	64.4	65.7
Ferroalloys	2.0	1.9	2.0	2.0	2.0	2.0	1.7
Indirect CO ₂	30.9	29.5	28.9	28.4	28.2	27.0	26.3
Ammonia Manufacture	18.5	18.9	19.5	19.5	20.1	18.9	18.0
Land-Use Change and Forestry (Sink) ^a	(1,097.7)	(1,110.0)	(1,108.1)	(887.5)	(885.9)	(896.4)	(902.5)
International Bunker Fuels ^b	113.9	101.0	102.3	109.9	112.9	105.3	100.2
CH₄	713.3	720.2	705.0	693.7	686.8	679.6	673.0
Stationary Sources	8.6	9.0	9.2	8.2	7.7	8.0	8.2
Mobile Sources	5.4	5.2	5.1	5.0	4.9	4.8	4.8
Coal Mining	95.4	80.5	74.9	74.6	74.4	69.8	66.8
Natural Gas Systems	132.8	137.6	138.7	134.4	133.9	129.9	127.4
Petroleum Systems	28.9	26.5	26.3	26.3	25.6	24.4	23.9
Petrochemical Production	1.3	1.7	1.7	1.8	1.8	1.8	1.8
Silicon Carbide Production	+	+	+	+	+	+	+
Enteric Fermentation	140.1	145.9	141.9	138.8	136.8	136.4	135.7
Manure Management	32.0	38.1	37.5	39.3	41.6	41.1	41.0
Rice Cultivation	7.8	8.3	7.6	8.2	8.7	9.1	8.2
Agricultural Residue Burning	0.7	0.7	0.8	0.8	0.9	0.8	0.9
Landfills	233.7	237.3	231.6	226.0	220.1	222.4	222.9
Wastewater Treatment	26.6	29.3	29.6	30.2	30.5	31.0	31.4
International Bunker Fuels ^b	0.2	0.1	0.1	0.2	0.2	0.1	0.1
N₂O	369.8	400.8	411.0	410.4	407.1	404.4	406.1
Stationary Source	12.2	12.9	13.4	13.6	13.7	13.9	14.3
Mobile Sources	48.6	57.7	57.4	57.0	56.5	56.0	55.7
Adipic Acid	14.2	17.1	17.0	11.0	7.4	7.3	7.7
Nitric Acid	17.0	19.0	19.8	20.3	19.9	19.2	18.9
Manure Management	15.3	15.6	16.0	16.3	16.3	16.4	16.7
Agricultural Soil Management	255.0	270.6	279.4	284.1	284.9	282.9	284.1
Agricultural Residue Burning	0.4	0.4	0.4	0.4	0.4	0.4	0.4
Human Sewage	6.7	7.3	7.4	7.6	7.7	8.0	8.1
Waste Combustion	0.3	0.3	0.3	0.2	0.2	0.2	0.2
International Bunker Fuels ^b	0.9	0.8	0.9	0.9	0.9	0.9	0.9
HFCs, PFCs, and SF₆	91.1	97.8	111.9	117.6	129.8	122.9	125.1
Substitution of Ozone Depleting Substances	0.9	22.6	32.3	39.9	47.4	54.4	61.5
Aluminum Production	16.8	10.9	11.5	10.1	8.3	8.2	7.3
HCFC-22 Production ^c	35.9	27.7	31.9	30.8	41.2	31.2	30.6
Semiconductor Manufacture ^d	3.3	6.8	6.3	7.6	8.4	9.0	8.5
Electrical Transmission and Distribution ^e	29.0	24.6	24.9	22.7	18.7	14.4	13.4

Magnesium Production and Processing ^e	5.1	5.1	5.1	6.4	5.7	5.7	3.7
Total	6,172.7	6,524.8	6,711.7	6,789.6	6,798.8	6,872.3	7,044.3

+ Does not exceed 0.05 Tg CO₂ Eq.

^a Sinks are only included in net emissions total, and are based partially on projected activity data. Parentheses indicate negative values (or sequestration).

^b Emissions from International Bunker Fuels are not included in totals.

^c HFC-23 emitted

^d Emissions from HFC-23, CF₄, C₂F₆, C₃F₈, SF₆, and the addition of NF₃

^e SF₆ emitted

Note: Totals may not sum due to independent rounding.

Table Q-5: Change in U.S. Greenhouse Gas Emissions and Sinks Using TAR vs SAR GWPs (Tg CO₂ Eq.)

Gas	1990	1995	1996	1997	1998	1999	2000
CO ₂	NC	NC	NC	NC	NC	NC	NC
CH ₄	62.0	62.6	61.3	60.3	59.7	59.1	58.5
N ₂ O	(17.5)	(19.0)	(19.4)	(19.4)	(9.3)	(9.1)	(9.2)
HFCs, PFCs, and SF ₆ *	(2.6)	(0.7)	0.1	0.6	2.1	2.9	3.8
Total	42.0	42.9	41.9	41.6	42.6	42.8	43.1

NC (No change)

*Includes NF₃

Note: Totals may not sum due to independent rounding.

Table Q-6: Change in U.S. Greenhouse Gas Emissions Using TAR vs SAR GWPs (Percent)

Gas/Source	1990	1995	1996	1997	1998	1999	2000
CO ₂	NC	NC	NC	NC	NC	NC	NC
CH ₄	9.5	9.5	9.5	9.5	9.5	9.5	9.5
N ₂ O	(4.5)	(4.5)	(4.5)	(4.5)	(4.5)	(4.5)	(4.5)
HFCs, PFCs, and SF ₆	(2.7)	(0.7)	0.1	0.6	1.6	2.4	3.1
Substitution of Ozone Depleting Substances	(3.2)	3.7	5.4	5.1	5.6	6.1	6.5
Aluminum Production ^a	(7.0)	(7.7)	(7.8)	(7.9)	(7.9)	(7.9)	(8.0)
HCFC-22 Production ^b	2.6	2.6	2.6	2.6	2.6	2.6	2.6
Semiconductor Manufacture ^c	15.9	15.9	15.9	15.9	15.9	15.9	15.9
Electrical Transmission and Distribution ^d	(7.1)	(7.1)	(7.1)	(7.1)	(7.1)	(7.1)	(7.1)
Magnesium Production and Processing ^d	(7.1)	(7.1)	(7.1)	(7.1)	(7.1)	(7.1)	(7.1)
Total	0.7	0.7	0.6	0.6	0.6	0.6	0.6

NC (No change)

^a PFC emissions from CF₄ and C₂F₆

^b HFC-23 emitted

^c Emissions from HFC-23, CF₄, C₂F₆, C₃F₈, SF₆, and the addition of NF₃

^d SF₆ emitted

Note: Excludes Sinks.

Table Q-7: Effects on U.S. Greenhouse Gas Emissions Using TAR vs SAR GWPs (Tg CO₂ Eq.)

Gas	Trend from 1990 to 2000		Revisions to Annual Estimates	
	SAR	TAR	1990	2000
CO ₂	841.5	841.5	NC	NC
CH ₄	(36.8)	(40.3)	62.0	58.5
N ₂ O	38.0	36.3	(17.5)	(19.2)
HFCs, PFCs, and SF ₆ *	27.7	34.0	(2.6)	3.8
Total	870.5	871.6	42.0	43.1
Percent Change	14.2%	14.1%	0.7%	0.6%

NC (No Change)

*Includes NF₃

Note: Totals may not sum due to independent rounding. Excludes Sinks.

Overall, these revisions to GWP values do not have a significant effect on U.S. emission trends, as shown in Table Q-5 and Table Q-6. Table Q-8 below shows a comparison of total emissions estimates by sector using both the IPCC SAR and TAR GWP values. For most sectors, the change in emissions was minimal. The effect on emissions from waste was by far the greatest (9.1 percent), due the predominance of CH₄ emissions in this sector. Emissions from all other sectors were comprised of mainly CO₂ or a mix of gases, which moderated the effect of the changes.

Table Q-8: Comparison of Emissions by Sector using IPCC SAR and TAR GWP Values (Tg CO₂ Eq.)

Sector	1990	1995	1996	1997	1998	1999	2000
Energy							
SAR GWP (Used in Inventory)	5,141.9	5,452.4	5,629.9	5,697.9	5,709.5	5,793.9	5,962.6
TAR GWP	5,162.6	5,471.6	5,648.6	5,716.2	5,727.6	5,811.2	5,979.4
Difference (%)	0.4%	0.4%	0.3%	0.3%	0.3%	0.3%	0.3%
Industrial Processes							
SAR GWP (Used in Inventory)	295.7	301.9	312.3	322.4	322.1	310.8	312.8
TAR GWP	291.8	299.6	310.8	321.7	323.1	312.6	315.5
Difference (%)	-1.3%	-0.8%	-0.5%	-0.2%	0.3%	0.6%	0.8%
Agriculture							
SAR GWP (Used in Inventory)	448.4	476.4	481.3	485.9	487.6	485.0	485.1
TAR GWP	451.3	479.6	483.6	487.9	489.7	487.1	487.1
Difference (%)	0.6%	0.7%	0.5%	0.4%	0.4%	0.4%	0.4%
Land-Use Change and Forestry							
SAR GWP (Used in Inventory)	(1,097.7)	(1,110.0)	(1,108.1)	(887.5)	(885.9)	(896.4)	(902.5)
TAR GWP	(1,097.7)	(1,110.0)	(1,108.1)	(887.5)	(885.9)	(896.4)	(902.5)
Difference (%)	NC	NC	NC	NC	NC	NC	NC
Waste							
SAR GWP (Used in Inventory)	244.7	251.1	246.3	241.9	236.9	239.8	240.6
TAR GWP	267.0	273.9	268.7	263.8	258.3	261.5	262.4
Difference (%)	9.1%	9.1%	9.1%	9.1%	9.0%	9.0%	9.0%
Net Emissions (Sources and Sinks)							
SAR GWP (Used in Inventory)	5,033.0	5,371.8	5,561.7	5,860.5	5,870.3	5,933.1	6,098.7
TAR GWP	5,074.9	5,414.7	5,603.6	5,902.1	5,912.9	5,975.9	6,141.8
Difference (%)	0.8%	0.8%	0.7%	0.7%	0.7%	0.7%	0.7%

NC (No change)

Note: Totals may not sum due to independent rounding.

ANNEX R: Ozone Depleting Substance Emissions

Ozone is present in both the stratosphere,¹ where it shields the earth from harmful levels of ultraviolet radiation, and at lower concentrations in the troposphere,² where it is the main component of anthropogenic photochemical “smog.” Chlorofluorocarbons (CFCs), halons, carbon tetrachloride, methyl chloroform, and hydrochlorofluorocarbons (HCFCs), along with certain other chlorine and bromine containing compounds, have been found to deplete the ozone levels in the stratosphere. These compounds are commonly referred to as ozone depleting substances (ODSs). If left unchecked, stratospheric ozone depletion could result in a dangerous increase of ultraviolet radiation reaching the earth’s surface. In 1987, nations around the world signed the *Montreal Protocol on Substances that Deplete the Ozone Layer*. This landmark agreement created an international framework for limiting, and ultimately eliminating, the production of most ozone depleting substances. ODSs have historically been used in a variety of industrial applications, including refrigeration and air conditioning, foam blowing, fire extinguishing, as an aerosol propellant, sterilization, and solvent cleaning.

In the United States, the Clean Air Act Amendments of 1990 provide the legal instrument for implementation of the *Montreal Protocol* controls. The Clean Air Act classifies ozone depleting substances as either Class I or Class II, depending upon the ozone depletion potential (ODP) of the compound.³ The production of CFCs, halons, carbon tetrachloride, and methyl chloroform—all Class I substances—has already ended in the United States. However, large amounts of these chemicals remain in existing equipment,⁴ and stockpiles of the ODSs are used for maintaining the equipment. In addition, U.S. regulations require the recovery of ODSs in order to minimize “venting” to the atmosphere. As a result, emissions of Class I compounds will continue, albeit in ever decreasing amounts, into the early part of the next century. Class II designated substances, all of which are hydrochlorofluorocarbons (HCFCs), are being phased out at later dates because they have lower ozone depletion potentials. These compounds serve as interim replacements for Class I compounds in many industrial applications. The use and emissions of HCFCs in the United States is anticipated to increase over the next several years as equipment that use Class I substances are retired from use. Under current controls, however, the production for domestic use of all HCFCs in the United States will end by the year 2030.

In addition to contributing to ozone depletion, CFCs, halons, carbon tetrachloride, methyl chloroform, and HCFCs are also potent greenhouse gases. However, the depletion of the ozone layer has a cooling effect on the climate that counteracts the direct warming from tropospheric emissions of ODSs. Stratospheric ozone influences the earth’s radiative balance by absorption and emission of longwave radiation from the troposphere as well as absorption of shortwave radiation from the sun.

The IPCC has prepared both direct GWPs and net (combined direct warming and indirect cooling) GWP ranges for some of the most common ozone depleting substances (IPCC 1996). See Annex Q for a listing of the net GWP values for ODS.

¹ The stratosphere is the layer from the top of the troposphere up to about 50 kilometers. Approximately 90 percent of atmospheric ozone is within the stratosphere. The greatest concentration of ozone occurs in the middle of the stratosphere, in a region commonly called the ozone layer.

² The troposphere is the layer from the ground up to about 11 kilometers near the poles and 16 kilometers in equatorial regions (i.e., the lowest layer of the atmosphere, where humans live). It contains roughly 80 percent of the mass of all gases in the atmosphere and is the site for weather processes including most of the water vapor and clouds.

³ Substances with an ozone depletion potential of 0.2 or greater are designated as Class I. All other substances that may deplete stratospheric ozone but which have an ODP of less than 0.2 are Class II.

⁴ Older refrigeration and air-conditioning equipment, fire extinguishing systems, meter-dose inhalers, and foam products blown with CFCs/HCFCs may still contain ODS.

Although the IPCC emission inventory guidelines do not require the reporting of emissions of ozone depleting substances, the United States believes that no inventory is complete without the inclusion of these compounds. Emission estimates for several ozone depleting substances are provided in Table R-1.

Table R-1: Emissions of Ozone Depleting Substances (Gg)

Compound	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000
Class I											
CFC-11	53.5	48.3	45.1	45.4	36.6	36.2	26.6	25.1	24.9	24.0	22.8
CFC-12	112.6	103.5	80.5	79.3	57.6	51.8	35.5	23.1	21.0	14.0	17.2
CFC-113	52.7	41.1	34.2	34.2	17.1	17.1	+	+	+	+	+
CFC-114	4.7	3.6	3.0	3.0	1.6	1.6	0.3	0.1	0.1	+	+
CFC-115	4.2	4.0	3.8	3.6	3.3	3.0	3.2	2.9	2.7	2.6	2.3
Carbon Tetrachloride	32.3	+	21.7	18.6	15.5	4.7	+	+	+	+	+
Methyl Chloroform	316.6	309.4	216.6	185.7	154.7	92.8	+	+	+	+	+
Halon-1211	1.0	1.1	1.0	1.1	1.0	1.1	1.1	1.1	1.1	1.1	1.1
Halon-1301	1.8	1.8	1.7	1.7	1.4	1.4	1.4	1.3	1.3	1.3	1.3
Class II											
HCFC-22	34.0	35.4	35.2	35.3	37.7	39.3	41.0	42.4	43.8	74.1	79.1
HCFC-123	+	+	0.1	0.3	0.5	0.6	0.7	0.8	0.9	1.0	1.1
HCFC-124	+	+	0.2	2.7	5.3	5.6	5.9	6.2	6.4	6.5	6.5
HCFC-141b	1.3	1.7	1.7	2.9	6.2	9.9	9.9	8.8	9.7	10.9	10.9
HCFC-142b	0.8	1.4	1.9	2.6	3.3	3.6	4.0	4.3	4.7	5.0	5.4
HCFC-225ca/cb	+	+	+	+	+	+	+	+	+	+	+

Source: EPA, Office of Atmospheric Programs

+ Does not exceed 0.05 Gg

Methodology and Data Sources

Emissions of ozone depleting substances were estimated using two simulation models: the Atmospheric and Health Effects Framework (AHEF) and the EPA's Vintaging Model.

AHEF contains estimates of U.S. domestic use of each of the ozone depleting substances. These estimates were based upon data that industry reports to the EPA and other published material. The annual consumption of each compound was divided into various end-uses based upon historical trends and research into specific industrial applications. These end-uses include refrigerants, foam blowing agents, solvents, aerosol propellants, sterilants, and fire extinguishing agents.

With the exception of aerosols, solvents, and certain foam blowing agents, emissions of ozone depleting substances are not instantaneous, but instead occur gradually over time (i.e., emissions in a given year are the result of both ODS use in that year and use in previous years). Each end-use has a certain release profile, which gives the percentage of the compound that is released to the atmosphere each year until all releases have occurred. In refrigeration equipment, for example, the initial charge is released or leaked slowly over the lifetime of the equipment, which could be 20 or more years. In addition, not all of the refrigerant is ultimately emitted—some will be recovered when the equipment is retired from operation.

The AHEF model was used to estimate emissions of ODSs that were in use prior to the controls implemented under the *Montreal Protocol*. This included CFCs, halons, carbon tetrachloride, methyl chloroform, and HCFC-22. Certain HCFCs, such as HCFC-123, HCFC-124, HCFC-141b, HCFC-142b, HCFC-225ca and HCFC-225cb, have also entered the market as interim substitutes for ODSs. Emissions estimates for these compounds were taken from the EPA's Vintaging Model.

The Vintaging Model was used to estimate the use and emissions of various ODS substitutes, including HCFCs. The model, named for its method of tracking the emissions of annual "vintages" of new equipment that enter into service, is a "bottom-up" model. It models the consumption of chemicals based on estimates of the quantity of equipment or

products sold, serviced, and retired each year, and the amount of the chemical required to manufacture and/or maintain the equipment. The Vintaging model makes use of this market information to build an inventory of the in-use stocks of the equipment in each of the end-uses. Emissions are estimated by applying annual leak rates, service emission rates, and disposal emission rates to each population of equipment, as in the AHEF. By aggregating the emission and consumption output from the different end-uses, the model produces estimates of total annual use and emissions of each chemical.

Please see Annex J of this Inventory for a more detailed discussion of the Vintaging Model.

Uncertainties

Uncertainties exist with regard to the levels of chemical production, equipment sales, equipment characteristics, and end-use emissions profiles that are used by these models.

ANNEX S: Sulfur Dioxide Emissions

Sulfur dioxide (SO₂) emitted into the atmosphere through natural and anthropogenic processes affects the Earth's radiative budget through photochemical transformation into sulfate aerosols that can (1) scatter sunlight back to space, thereby reducing the radiation reaching the Earth's surface; (2) affect cloud formation; and (3) affect atmospheric chemical composition (e.g., stratospheric ozone, by providing surfaces for heterogeneous chemical reactions). The overall effect of SO₂ derived aerosols on radiative forcing is believed to be negative (IPCC 1996). However, because SO₂ is short-lived and unevenly distributed through the atmosphere, its radiative forcing impacts are highly uncertain. Sulfur dioxide emissions have been provided below in Table S-1.

The major source of SO₂ emissions in the United States was the burning of sulfur containing fuels, mainly coal. Metal smelting and other industrial processes also released significant quantities of SO₂. As a result, the largest contributors to U.S. emissions of SO₂ were electric utilities, accounting for 63 percent in 2000 (see Table S-2). Coal combustion accounted for approximately 94 percent of SO₂ emissions from electric utilities in the same year. The second largest source was industrial fuel combustion, which produced 4 percent of 2000 SO₂ emissions. Overall, SO₂ emissions in the United States decreased by 23 percent from 1990 to 2000. The majority of this decline came from reductions from electric utilities, primarily due to increased consumption of low sulfur coal from surface mines in western states.

Sulfur dioxide is important for reasons other than its effect on radiative forcing. It is a major contributor to the formation of urban smog and acid rain. As a contributor to urban smog, high concentrations of SO₂ can cause significant increases in acute and chronic respiratory diseases. In addition, once SO₂ is emitted, it is chemically transformed in the atmosphere and returns to earth as the primary contributor to acid deposition, or acid rain. Acid rain has been found to accelerate the decay of building materials and paints, and to cause the acidification of lakes and streams and damage trees. As a result of these harmful effects, the United States has regulated the emissions of SO₂ under the Clean Air Act. The EPA has also developed a strategy to control these emissions via four programs: (1) the National Ambient Air Quality Standards program,¹ (2) New Source Performance Standards,² (3) the New Source Review/Prevention of Significant Deterioration Program,³ and (4) the sulfur dioxide allowance program.⁴

References

EPA (2001) *National Air Pollutant Emissions Trends Report, 1900-2000*, U.S. Environmental Protection Agency, Office of Air Quality Planning and Standards, Research Triangle Park, NC.

Table S-1: SO₂ Emissions (Gg)

Sector/Source	1990	1995	1996	1997	1998	1999	2000
Energy	20,136	16,247	16,641	17,052	17,157	16,517	15,435
Stationary Combustion	18,407	14,724	14,726	15,104	15,192	14,540	13,496
Mobile Combustion	1,339	1,189	1,612	1,636	1,655	1,668	1,626
Oil and Gas Activities	390	334	304	312	310	309	314
Industrial Processes	1,306	1,117	958	993	996	992	1,031
Chemical Manufacturing	269	260	231	235	237	238	243
Metals Processing	658	481	354	369	367	363	373
Storage and Transport	6	2	5	5	5	5	5
Other Industrial Processes	362	365	354	371	376	376	392

¹ [42 U.S.C § 7409, CAA § 109]

² [42 U.S.C § 7411, CAA § 111]

³ [42 U.S.C § 7473, CAA § 163]

⁴ [42 U.S.C § 7651, CAA § 401]

Miscellaneous*	11	9	15	14	11	11	19
Solvent Use	0	1	1	1	1	1	1
Degreasing	0	0	0	0	0	0	0
Graphic Arts	0	0	0	0	0	0	0
Dry Cleaning	NA	0	0	0	0	0	0
Surface Coating	0	0	0	0	0	0	0
Other Industrial	0	0	1	1	1	1	1
Non-industrial	NA	NA	NA	NA	NA	NA	NA
Agriculture	NA	NA	NA	NA	NA	NA	NA
Agricultural Burning	NA	NA	NA	NA	NA	NA	NA
Waste	39	43	29	30	31	31	32
Waste Combustion	39	42	28	29	30	30	31
Landfills	0	0	1	1	1	1	1
Wastewater Treatment	0	1	0	0	0	0	0
Miscellaneous Waste	0	0	0	0	0	0	0
Total	21,481	17,408	17,629	18,076	18,185	17,541	16,499

Source: (EPA 2000)

* Miscellaneous includes other combustion and fugitive dust categories.

NA (Not Available)

Note: Totals may not sum due to independent rounding.

Table S-2: SO₂ Emissions from Electric Utilities (Gg)

Fuel Type	1990	1995	1996	1997	1998	1999	2000
Coal	13,807	10,526	11,073	11,444	11,313	10,729	9,728
Petroleum	580	375	417	466	691	580	464
Natural Gas	1	8	6	5	5	6	8
Misc. Internal Combustion	45	50	48	51	52	53	54
Other	NA	NA	4	4	110	112	80
Total	14,432	10,959	11,549	11,971	12,171	11,479	10,333

Source: (EPA 2000)

Note: Totals may not sum due to independent rounding.

ANNEX T: Complete List of Source Categories

Chapter/Source	Gas(es)
Energy	
Carbon Dioxide Emissions from Fossil Fuel Combustion	CO ₂
Carbon Stored in Products from Non-Energy Uses of Fossil Fuels	CO ₂
Stationary Combustion (excluding CO ₂)	CH ₄ , N ₂ O, CO, NO _x , NMVOC
Mobile Combustion (excluding CO ₂)	CH ₄ , N ₂ O, CO, NO _x , NMVOC
Coal Mining	CH ₄
Natural Gas Systems	CH ₄
Petroleum Systems	CH ₄
Municipal Solid Waste Combustion	CO ₂ , N ₂ O
Natural Gas Flaring and Ambient Air Pollutant Emissions from Oil and Gas Activities	CO ₂ , CO, NO _x , NMVOC
Indirect CO ₂ from CH ₄ Oxidation	CO ₂
International Bunker Fuels	CO ₂ , CH ₄ , N ₂ O, CO, NO _x , NMVOC
Wood Biomass and Ethanol Consumption	CO ₂
Industrial Processes	
Iron and Steel Production	CO ₂
Cement Manufacture	CO ₂
Ammonia Manufacture	CO ₂
Lime Manufacture	CO ₂
Limestone and Dolomite Use	CO ₂
Soda Ash Manufacture and Consumption	CO ₂
Ferroalloy Production	CO ₂
Titanium Dioxide Production	CO ₂
Carbon Dioxide Consumption	CO ₂
Petrochemical Production	CH ₄
Silicon Carbide Production	CH ₄
Adipic Acid Production	N ₂ O
Nitric Acid Production	N ₂ O
Substitution of Ozone Depleting Substances	HFCs, PFCs ^a
Aluminum Production	CO ₂ , CF ₄ , C ₂ F ₆
HCFC-22 Production	HFC-23
Semiconductor Manufacture	HFCs, PFCs, SF ₆ ^b
Electrical Transmission and Distribution	SF ₆
Magnesium Production and Processing	SF ₆
Industrial Sources of Ambient Air Pollutants	CO, NO _x , NMVOC
Solvent Use	CO, NO _x , NMVOC
Agriculture	
Enteric Fermentation	CH ₄
Manure Management	CH ₄ , N ₂ O
Rice Cultivation	CH ₄
Agricultural Soil Management	N ₂ O
Agricultural Residue Burning	CH ₄ , N ₂ O, CO, NO _x
Land-Use Change and Forestry	
Changes in Forest Carbon Stocks	CO ₂ (sink)
Changes in Carbon Stocks in Urban Trees	CO ₂ (sink)
Changes in Agricultural Soil Carbon Stocks	CO ₂ (sink)
Changes in Yard Trimming Carbon Stocks in Landfills	CO ₂ (sink)
Waste	
Landfills	CH ₄
Wastewater Treatment	CH ₄
Human Sewage	N ₂ O
Waste Sources of Ambient Air Pollutants	CO, NO _x , NMVOC

^a In 1999, included HFC-23, HFC-125, HFC-134a, HFC-143a, HFC-152a, HFC-227ea, HFC-236fa, HFC-4310mee, C₄F₁₀, C₆F₁₄, PFC/PFPEs

^b Included such gases as HFC-23, CF₄, C₂F₆, SF₆

ANNEX U: IPCC Reference Approach for Estimating CO₂ Emissions from Fossil Fuel Combustion

It is possible to estimate carbon dioxide (CO₂) emissions from fossil fuel consumption using alternative methodologies and different data sources than those described in Annex A. For example, the UNFCCC reporting guidelines request that countries, in addition to their “bottom-up” sectoral methodology, to complete a “top-down” Reference Approach for estimating CO₂ emissions from fossil fuel combustion. Section 1.3 of the *Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories: Reporting Instructions* states, “If a detailed, Sectoral Approach for energy has been used for the estimation of CO₂ from fuel combustion you are still asked to complete...the Reference Approach...for verification purposes” (IPCC/UNEP/OECD/IEA 1997). This reference method estimates fossil fuel consumption by adjusting national aggregate fuel production data for imports, exports, and stock changes rather than relying on end-user consumption surveys. The basic principle is that once carbon-based fuels are brought into a national economy, they are either saved in some way (e.g., stored in products, kept in fuel stocks, or left unoxidized in ash) or combusted, and therefore the carbon in them is oxidized and released into the atmosphere. Accounting for actual consumption of fuels at the sectoral or sub-national level is not required. The following discussion provides the detailed calculations for estimating CO₂ emissions from fossil fuel combustion from the United States using the IPCC-recommended Reference Approach.

Step 1: Collect and Assemble Data in Proper Format

To ensure the comparability of national inventories, the IPCC has recommended that countries report energy data using the International Energy Agency (IEA) reporting convention. National energy statistics were collected in physical units from several DOE/EIA documents in order to obtain the necessary data on production, imports, exports, and stock changes (EIA 2001a).

It was necessary to make a number of modifications to these data to generate more accurate apparent consumption estimates of these fuels. The first modification adjusts for consumption of fossil fuel feedstocks accounted for in the Industrial Processes chapter, which include unspecified coal for coal coke used in iron and steel production, natural gas used for ammonia production, and petroleum coke used in the production of aluminum, ferroalloys, and titanium dioxide. The second modification adjusts for consumption of bunker fuels, which refer to quantities of fuels used for international transportation estimated separately from U.S. totals. The third modification consists of the addition of U.S. territories data that are typically excluded from the national aggregate energy statistics. The territories include Puerto Rico, U.S. Virgin Islands, Guam, American Samoa, Wake Island, and U.S. Pacific Islands. These data, as well as the production, import, export, and stock change statistics, are presented in Table U-1.

The carbon content of fuel varies with the fuel's heat content. Therefore, for an accurate estimation of CO₂ emissions, fuel statistics were provided on an energy content basis (e.g., BTUs or joules). Because detailed fuel production statistics are typically provided in physical units (as in Table U-1), they were converted to units of energy before CO₂ emissions were calculated. Fuel statistics were converted to their energy equivalents by using conversion factors provided by DOE/EIA. These factors and their data sources are displayed in Table U-2. The resulting fuel type-specific energy data are provided in Table U-3.

Step 2: Estimate Apparent Fuel Consumption

The next step of the IPCC Reference Approach is to estimate “apparent consumption” of fuels within the country. This requires a balance of primary fuels produced, plus imports, minus exports, and adjusting for stock changes. In this way, carbon enters an economy through energy production and imports (and decreases in fuel stocks) and is transferred out of the country through exports (and increases in fuel stocks). Thus, apparent consumption of primary fuels (including crude oil, natural gas liquids, anthracite, bituminous, subbituminous and lignite coal, and natural gas) can be calculated as follows:

$$\text{Apparent Consumption} = \text{Production} + \text{Imports} - \text{Exports} - \text{Stock Change}$$

Flows of secondary fuels (e.g., gasoline, residual fuel, coke) should be added to primary apparent consumption. The production of secondary fuels, however, should be ignored in the calculations of apparent consumption since the carbon contained in these fuels is already accounted for in the supply of primary fuels from which they were derived (e.g., the estimate for apparent consumption of crude oil already contains the carbon from which gasoline would be refined). Flows of secondary fuels should therefore be calculated as follows:

$$\text{Secondary Consumption} = \text{Imports} - \text{Exports} - \text{Stock Change}$$

Note that this calculation can result in negative numbers for apparent consumption of secondary fuels. This result is perfectly acceptable since it merely indicates a net export or stock increase in the country of that fuel when domestic production is not considered.

Next, the apparent consumption and secondary consumption need to be adjusted for feedstock uses of fuels accounted for in the Industrial Processes chapter, international bunker fuels, and U.S. territory fuel consumption. Bunker fuels and feedstocks accounted for in the Industrial Processes chapter are subtracted from these estimates, while fuel consumption in U.S. territories is added.

The IPCC Reference Approach calls for estimating apparent fuel consumption before converting to a common energy unit. However, certain primary fuels in the United States (e.g., natural gas and steam coal) have separate conversion factors for production, imports, exports, and stock changes. In these cases, it is not appropriate to multiply apparent consumption by a single conversion factor since each of its components have different heat contents. Therefore, United States fuel statistics were converted to their heat equivalents before estimating apparent consumption. Results are provided in Table U-2.

Step 3: Estimate Carbon Emissions

Once apparent consumption is estimated, the remaining calculations are virtually identical to those for the “bottom-up” Sectoral Approach (see Annex A). That is:

- Potential CO₂ emissions were estimated using fuel-specific carbon coefficients (see Table U-3).¹
- The carbon in products from non-energy uses of fossil fuels (e.g., plastics or asphalt) was then estimated and subtracted (see Table U-4).
- Finally, to obtain actual CO₂ emissions, net emissions were adjusted for any carbon that remained unoxidized as a result of incomplete combustion (e.g., carbon contained in ash or soot).²

Step 4: Convert to CO₂ Emissions

Because the IPCC reporting guidelines recommend that countries report greenhouse gas emissions on a full molecular weight basis, the final step in estimating CO₂ emissions from fossil fuel consumption was converting from units of carbon to units of CO₂. Actual carbon emissions were multiplied by the molecular-to-atomic weight ratio of CO₂ to carbon (44/12) to obtain total carbon dioxide emitted from fossil fuel combustion in teragrams (Tg). The results are contained in Table U-5.

¹ Carbon coefficients from EIA were used wherever possible. Because EIA did not provide coefficients for coal, the IPCC-recommended emission factors were used in the top-down calculations for these fuels. See notes in Table U-4 for more specific source information.

² For the portion of carbon that is unoxidized during coal combustion, the IPCC suggests a global average value of 2 percent. However, because combustion technologies in the United States are more efficient, the United States inventory uses 1 percent in its calculations for petroleum and coal and 0.5 percent for natural gas.

Comparison Between Sectoral and Reference Approaches

These two alternative approaches can both produce reliable estimates that are comparable within a few percent. The major difference between methodologies employed by each approach lies in the energy data used to derive carbon emissions (i.e., the actual surveyed consumption for the Sectoral Approach versus apparent consumption derived for the Reference Approach). In theory, both approaches should yield identical results. In practice, however, slight discrepancies occur. For the United States, these differences are discussed below.

Differences in Total Amount of Energy Consumed

Table U-7³ summarizes the differences between the Reference and Sectoral approaches in estimating total energy consumption in the United States. Although theoretically the two methods should arrive at the same estimate for U.S. energy consumption, the Reference Approach provides an energy total that is 2.9 percent lower than the Sectoral Approach for 2000. The greatest difference lies in the higher estimate of petroleum consumption with the Sectoral Approach (3.5 percent).

There are several potential sources for the discrepancies in consumption estimates:

- *Product Definitions.* The fuel categories in the Reference Approach are different from those used in the Sectoral Approach, particularly for petroleum. For example, the Reference Approach estimates apparent consumption for crude oil. Crude oil is not typically consumed directly, but refined into other products. As a result, the United States does not focus on estimating the energy content of the various grades of crude oil, but rather estimating the energy content of the various products resulting from crude oil refining. The United States does not believe that estimating apparent consumption for crude oil, and the resulting energy content of the crude oil, is the most reliable method for the United States to estimate its energy consumption. Other differences in product definitions include using sector-specific coal statistics in the Sectoral Approach (i.e., residential, commercial, industrial coking, industrial other, and transportation coal), while the Reference Approach characterizes coal by rank (i.e. anthracite, bituminous, etc.). Also, the liquefied petroleum gas (LPG) statistics used in the bottom-up calculations are actually a composite category composed of natural gas liquids (NGL) and LPG.
- *Heat Equivalents.* It can be difficult to obtain heat equivalents for certain fuel types, particularly for categories such as "crude oil" where the key statistics are derived from thousands of producers in the United States and abroad. For heat equivalents by coal rank, it was necessary to refer back to EIA's *State Energy Data Report 1992* (1994) because this information is no longer published.
- *Possible inconsistencies in U.S. Energy Data.* The United States has not focused its energy data collection efforts on obtaining the type of aggregated information used in the Reference Approach. Rather, the United States believes that its emphasis on collection of detailed energy consumption data is a more accurate methodology for the United States to obtain reliable energy data. Therefore, top-down statistics used in the Reference Approach may not be as accurately collected as bottom-up statistics applied to the Sectoral Approach.
- *Balancing Item.* The Reference Approach uses *apparent* consumption estimates while the Sectoral Approach uses *reported* consumption estimates. While these numbers should be equal, there always seems to be a slight difference that is often accounted for in energy statistics as a "balancing item."

Differences in Estimated CO₂ Emissions

Given these differences in energy consumption data, the next step for each methodology involved estimating emissions of CO₂. Table U-8 summarizes the differences between the two methods in estimated carbon emissions.

As mentioned above, for 2000, the Reference Approach resulted in a 2.9 percent lower estimate of energy consumption in the United States than the Sectoral Approach. However, the resulting emissions estimate for the

³ Although complete energy consumption data and calculations are not presented, comparison tables are also presented for 1996.

Reference Approach was 0.6 percent lower. Both methods' estimates of natural gas emissions are almost exactly the same, but coal emission estimates from the Reference Approach are lower than the Sectoral Approach, while higher for petroleum. Potential reasons for these differences may include:

- *Product Definitions.* Coal data is aggregated differently in each methodology, as noted above. The format used for the Sectoral Approach likely results in more accurate estimates than in the Reference Approach. Also, the Reference Approach relies on a "crude oil" category for determining petroleum-related emissions. Given the many sources of crude oil in the United States, it is not an easy matter to track potential differences in carbon content between many different sources of crude, particularly since information on the carbon content of crude oil is not regularly collected.
- *Carbon Coefficients.* The Reference Approach relies on several default carbon coefficients by rank provided by IPCC (IPCC/UNEP/OECD/IEA 1997), while the Sectoral Approach uses annually updated category-specific coefficients by sector that are likely to be more accurate. Also, as noted above, the carbon coefficient for crude oil is more uncertain than that for specific secondary petroleum products, given the many sources and grades of crude oil consumed in the United States.

Although the two approaches produce similar results, the United States believes that the "bottom-up" Sectoral Approach provides a more accurate assessment of CO₂ emissions at the fuel level. This improvement in accuracy is largely a result of the data collection techniques used in the United States, where there has been more emphasis on obtaining the detailed products-based information used in the Sectoral Approach than obtaining the aggregated energy flow data used in the Reference Approach. The United States believes that it is valuable to understand both methods.

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Table U-1: 2000 U.S. Energy Statistics (Physical Units)

Fuel Category (Units)	Fuel Type	Stock			U.S. Territories			
		Production	Imports	Exports		Change	Adjustment	Bunkers
Solid Fuels (Thousand Short Tons)	Anthracite Coal	4,508	a	a	a			
	Bituminous Coal	548,467	a	a	a			
	Sub-bituminous Coal	433,784	a	a	a			
	Lignite	88,740	a	a	a			
Gas Fuels (Million Cubic Feet)	Coke	3,781	1,146	227				
	Unspecified Coal	12,513	58,489	(40,245)		27,325		441
	Natural Gas	18,377,829	3,726,290	237,035	(845,082)	330,653		11,400
	Crude Oil	2,130,707	3,319,816	18,352	(25,538)			
	Nat Gas Liquids and LRGs	699,415	93,755	28,470	(6,999)			2,203
Liquid Fuels (Thousand Barrels)	Other Liquids	52,196	211,023	17,935	2,699			
	Motor Gasoline	85,938	156,230	52,539	(1,089)			35,408
	Aviation Gasoline		336	0	(309)			
	Kerosene		822	779	(764)			1,150
	Jet Fuel		59,125	11,628	4,017		143,769	
	Distillate Fuel		107,919	63,198	(7,436)		19,704	21,590
	Residual Fuel		128,912	50,858	370		70,634	16,071
	Naptha for petrochemical feedstocks		43,357	0	488			
	Petroleum Coke		394	116,589	1,360		13,966	
	Other Oil for petrochemical feedstocks		52,338	0	125			
	Special Napthas		3,873	7,425	(239)			
	Lubricants		4,950	9,472	279			
	Waxes		860	1,293	70			223
Asphalt/Road Oil		10,300	2,104	8,183				
Still Gas		0	0	0				
Misc. Products		76	67	(601)			38,359	

[a] Included in Unspecified Coal

Data Sources: Solid and Gas Fuels – EIA (2001a) Annual Energy Review 2000; Liquid Fuels – Petroleum Supply Annual 2000, Table 2.

Table U-2: Conversion Factors to Energy Units (Heat Equivalents)

Fuel Category (Units)	Fuel Type	Stock			U.S. Territories			
		Production	Imports	Exports		Change	Adjustment	Bunkers
Solid Fuels (Million Btu/Short Ton)	Anthracite Coal		22.57					
	Bituminous Coal		23.89					

Other Liquids	304.0	1,229.2	104.5	15.7			1,413.1
Motor Gasoline	447.7	814.0	273.7	(5.7)		184.5	1,178.1
Aviation Gasoline		1.7	0.0	(1.6)		6.5	3.3
Kerosene		4.7	4.4	(4.3)			11.1
Jet Fuel		335.2	65.9	22.8			(568.6)
Distillate Fuel		628.6	368.1	(43.3)		815.2	314.8
Residual Oil		810.5	319.7	2.3		125.8	145.4
Naptha for petrochemical feedstocks		227.5	0.0	2.6		101.0	225.0
Petroleum Coke		2.4	702.3	8.2		84.1	(792.3)
Other Oil for petrochemical feedstocks		304.9	0.0	0.7			304.1
Special Napthas		20.3	39.0	(1.3)			(17.4)
Lubricants		30.0	57.4	1.7		1.4	(27.8)
Waxes		4.8	7.2	0.4			(2.8)
Asphalt/Road Oil		68.4	14.0	54.3			0.1
Still Gas		0.0	0.0	0.0			0.0
Misc. Products		0.4	0.4	(3.5)		222.3	225.9
Total	56,376.3	28,826.8	3,971.2	(1,835.5)	1,173.1	660.8	81,181.0

Note: Totals may not sum due to independent rounding.

Table U-4: 2000 Potential Carbon Dioxide Emissions

Fuel Category	Fuel Type	Apparent Consumption (QBtu)	Carbon Coefficients (Tg Carbon/QBtu)	Potential Emissions (Tg CO ₂ Eq.)
Solid Fuels	Anthracite Coal	0.102	26.86	10.0
	Bituminous Coal	13.103	25.86	1,242.4
	Sub-bituminous Coal	7.435	26.26	715.9
	Lignite	1.142	27.66	115.8
	Coke	0.060	25.56	5.6
	Unspecified	(1.112)	25.34	(103.3)
Gas Fuels	Natural Gas	22,972	14.47	1,218.8
	Crude Oil	32,179	20.23	2,387.2
Liquid Fuels	Nat Gas Liquids and LRGs	2,889	16.99	180.0
	Other Liquids	1,413	20.23	104.8
	Motor Gasoline	1,178	19.34	83.5
	Aviation Gasoline	0,003	18.87	0.2
	Kerosene	0,011	19.72	0.8
	Jet Fuel	(0,569)	19.33	(40.3)
	Distillate Fuel	0,315	19.95	23.0
	Residual Oil	0,145	21.49	11.5
	Naptha for petrochemical feedstocks	0,225	18.14	15.0

Petroleum Coke	(0.792)	27.85	(80.9)
Other Oil for petrochemical feedstocks	0.304	19.95	22.2
Special Naphthas	(0.017)	19.86	(1.3)
Lubricants	(0.028)	20.24	(2.1)
Waxes	(0.003)	19.81	(0.2)
Asphalt/Road Oil	0.000	20.62	0.0
Still Gas	0.000	17.51	0.0
Misc. Products	0.226	20.23	16.8
Total			5,925.5

Data Sources: Coal and Lignite – IPCC (1997) Revised 1996 IPCC Guidelines Reference Manual, Table 1-1; Unspecified Solid Fuels - EIA (2001b) Monthly Energy Review, November 2001 Table C1 (U.S. Average); Natural Gas and Liquid Fuels - EIA (2000) *Emissions of Greenhouse Gases in the United States 1999*.

Note: Totals may not sum due to independent rounding.

Table U-5: 2000 Non-Energy Carbon Stored in Products

Fuel Type	Consumption for Non-Energy Use (TBtu)	Carbon Coefficients (Tg Carbon/QBtu)	Carbon Content (Tg Carbon)	Fraction Sequestered	Carbon Stored (Tg CO₂ Eq.)
Coal	26.4	25.56	0.7	0.75	1.9
Natural Gas	342.4	14.47	5.0	0.63	11.5
Asphalt & Road Oil	1,275.7	20.62	26.3	1.00	96.5
LPG	1,707.3	16.88	28.8	0.63	66.8
Lubricants	370.6	20.24	7.5	0.09	2.5
Pentanes Plus	286.8	18.24	5.2	0.63	12.1
Petrochemical Feedstocks	a	a	a	a	54.9
Petroleum Coke	141.4	27.85	3.9	0.50	7.2
Special Naptha	97.4	19.86	1.9	0.00	0.0
Waxes/Misc.	a	a	a	a	13.5
Misc. U.S. Territories Petroleum	a	a	a	a	16.5
Total					283.4

[a] Values for Misc. U.S. Territories Petroleum, Petrochemical Feedstocks and Waxes/Misc. are not shown because these categories are aggregates of numerous smaller components.

Note: Totals may not sum due to independent rounding.

Table U-6: 2000 Reference Approach CO₂ Emissions from Fossil Fuel Consumption (Tg CO₂ Eq. unless otherwise noted)

Fuel Category	Potential Emissions	Carbon Sequestered	Net Emissions	Fraction Oxidized	Total Emissions
Coal	1,986.4	1.9	1,984.5	99.0%	1,964.7
Petroleum	2,720.3	270.0	2,450.3	99.0%	2,425.8
Natural Gas	1,218.8	11.5	1,207.3	99.5%	1,201.3
Total	5,925.5	283.4	5,642.2	-	5,591.8

Note: Totals may not sum due to independent rounding.

1 Table U-7: Energy Consumption in the United States by Estimating Approach (TBtu)*

Approach	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000
Sectoral	70,996.2	70,588.5	71,959.5	73,512.7	75,089.7	76,034.7	78,557.0	79,713.8	79,744.1	81,242.7	83,606.3
Coal	18,175.6	18,072.9	18,278.9	18,929.1	19,113.6	19,205.8	20,126.9	20,670.7	20,872.0	20,949.9	21,732.3
Natural Gas	18,912.2	19,233.6	19,755.5	20,437.1	20,921.5	21,813.1	22,212.5	22,307.8	21,532.4	21,907.9	23,038.0
Petroleum	33,908.4	33,282.0	33,925.2	34,146.6	35,054.6	35,015.8	36,217.6	36,735.2	37,339.8	38,384.9	38,836.0
Reference (Apparent)	69,497.8	67,934.4	69,395.8	71,364.6	72,997.9	73,815.8	76,242.4	77,589.9	77,561.4	79,015.7	80,181.0
Coal	18,408.1	17,498.9	17,817.6	18,339.5	18,791.5	18,601.1	19,567.8	20,230.2	20,104.1	20,188.7	20,728.7
Natural Gas	19,264.8	19,244.0	19,752.1	20,466.7	20,910.5	21,794.5	22,179.3	22,155.9	21,520.0	21,918.3	22,972.2
Petroleum	31,824.9	31,191.5	31,826.1	33,558.5	33,295.9	33,420.2	34,495.3	35,203.9	35,937.3	36,908.7	37,480.1
Difference	-2.1%	-3.8%	-3.6%	-2.9%	-2.8%	-2.9%	-2.9%	-2.7%	-2.7%	-2.7%	-2.9%
Coal	1.3%	-3.2%	-2.5%	-3.1%	-1.7%	-3.1%	-2.8%	-2.1%	-3.7%	-3.6%	-4.6%
Natural Gas	1.9%	-0.1%	+	0.1%	-0.1%	-0.1%	-0.1%	-0.7%	-0.1%	+	-0.3%
Petroleum	-6.1%	-6.3%	-6.2%	-4.7%	-5.0%	-4.6%	-4.8%	-4.2%	-3.8%	-3.8%	-3.5%

2 * Includes U.S. territories

3 + Does not exceed 0.05%

4 Note: Totals may not sum due to independent rounding.

5

6 Table U-8: CO₂ Emissions from Fossil Fuel Combustion by Estimating Approach (Tg CO₂ Eq.)

Approach	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000
Sectoral	4,779.6	4,732.8	4,835.8	4,950.7	5,047.0	5,084.8	5,266.3	5,339.2	5,355.8	5,448.4	5,623.1
Coal	1,692.6	1,684.0	1,702.2	1,764.1	1,782.6	1,792.7	1,878.4	1,930.5	1,949.7	1,956.9	2,030.1
Natural Gas	988.8	1,006.1	1,035.0	1,070.3	1,093.6	1,141.3	1,162.4	1,166.7	1,125.8	1,145.2	1,204.8
Petroleum	2,098.2	2,042.6	2,098.6	2,116.3	2,170.8	2,150.9	2,25.6	2,242.0	2,280.3	2,346.3	2,388.2
Reference (Apparent)	4,832.3	4,693.2	4,800.5	4,931.7	5,035.8	5,067.2	5,249.7	5,349.3	5,357.7	5,437.1	5,591.8
Coal	1,746.9	1,657.8	1,686.0	1,735.1	1,778.8	1,761.3	1,852.3	1,914.5	1,905.6	1,915.9	1,964.7
Natural Gas	1,007.4	1,006.7	1,034.8	1,071.8	1,093.0	1,140.3	1,160.6	1,158.7	1,125.2	1,145.8	1,201.3
Petroleum	2,078.0	2,028.7	2,079.6	2,124.8	2,163.9	2,165.6	2,236.8	2,276.1	2,326.9	2,375.4	2,425.8
Difference	1.1%	-0.8%	-0.7%	-0.4%	-0.2%	-0.3%	-0.3%	0.2%	+	-0.2%	-0.6%
Coal	3.2%	-1.6%	-0.9%	-1.6%	-0.2%	-1.7%	-1.4%	-0.8%	-2.3%	-2.1%	-3.2%
Natural Gas	1.9%	0.1%	-0.0%	-0.1%	-0.1%	-0.1%	-0.2%	-0.7%	-0.1%	+	-0.3%
Petroleum	-1.0%	-0.7%	-0.9%	-0.4%	-0.3%	0.7%	0.5%	1.5%	2.0%	1.2%	1.6%

7 + Does not exceed 0.05%

8 Note: Totals may not sum due to independent rounding. Includes U.S. territories.

9

ANNEX V: Sources of Greenhouse Gas Emissions Excluded

Although this report is intended to be a comprehensive assessment of anthropogenic¹ sources and sinks of greenhouse gas emissions for the United States, certain sources have been identified yet excluded from the estimates presented for various reasons. Before discussing these sources, however, it is important to note that processes or activities that are not *anthropogenic in origin* or do not result in a *net source or sink* of greenhouse gas emissions are intentionally excluded from a national inventory of anthropogenic greenhouse gas emissions. In general, processes or activities that are not anthropogenic are considered natural (i.e., not directly influenced by human activity) in origin and, as an example, would include the following:

- Volcanic eruptions
- Carbon dioxide (CO₂) exchange (i.e., uptake or release) by oceans
- Natural forest fires²
- Methane (CH₄) emissions from wetlands not affected by human induced land-use changes

Some processes or activities may be anthropogenic in origin but do not result in net emissions of greenhouse gases, such as the respiration of CO₂ by people or domesticated animals.³ Given a source category that is both anthropogenic and results in net greenhouse gas emissions, reasons for excluding a source related to an anthropogenic activity include one or more of the following:

- There is insufficient scientific understanding to develop a reliable method for estimating emissions at a national level.
- Although an estimating method has been developed, data were not adequately available to calculate emissions.
- Emissions were implicitly accounted for within another source category (e.g., CO₂ from Fossil Fuel Combustion).

It is also important to note that the United States believes the exclusion of the sources discussed below introduces only a minor bias in its overall estimate of U.S. greenhouse gas emissions.

Separate Cruise and LTO Emissions from the Combustion of Jet Fuel

The combustion of jet fuel by aircraft results in emissions of CH₄, N₂O, CO, NO_x, and NMVOCs. The emissions per mass of fuel combusted during landing/take-off (LTO) operations differ from those during aircraft cruising. Accurate estimation of these emissions requires a detailed accounting of LTO cycles and fuel consumption during cruising by aircraft model (e.g., Boeing 747-400) as well as appropriate emission factors. Sufficient data for separately calculating near ground-level emissions during landing and take-off and cruise altitude emissions by aircraft model were not available for this report. (See *Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories: Reference Manual*, pp. 1.93 - 1.96)

CO₂ from Burning in Coal Deposits and Waste Piles

¹ The term “anthropogenic”, in this context, refers to greenhouse gas emissions and removals that are a direct result of human activities or are the result of natural processes that have been affected by human activities (IPCC/UNEP/OECD/IEA 1997).

² In some cases forest fires that are started either intentionally or unintentionally are viewed as mimicking natural burning processes that have been suppressed by other human forest management activities. The United States does not consider forest fires within its national boundaries to be a net source of greenhouse emissions.

³ Respiration of CO₂ by biological organisms is simply part of the broader global carbon cycle that also includes uptake of CO₂ by photosynthetic organisms.

Coal is periodically burned in deposits and waste piles. It has been estimated that the burning of coal in deposits and waste piles would represent less than 1.3 percent of total U.S. coal consumption, averaged over ten-years. Because there is currently no known source of data on the quantity of coal burned in waste piles and there is uncertainty as to the fraction of coal oxidized during such burnings, these CO₂ emissions are not currently estimated. Further research would be required to develop accurate emission factors and activity data for these emissions to be estimated (see *Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories: Reference Manual*, p. 1.112 - 1.113).

Fossil CO₂ from Petroleum and Natural Gas Wells, CO₂ Separated from Natural Gas, and CO₂ from Enhanced Oil Recovery (EOR)

Petroleum and natural gas well drilling, petroleum and natural gas production, and natural gas processing—including removal of CO₂—may result in emissions of CO₂ that was at one time stored in underground formations.

Carbon dioxide and other gases are naturally present in raw natural gas, in proportions that vary depending on the geochemical circumstances that caused the formation of the gas. After the heavier gases are removed during processing, small amounts of carbon dioxide may be allowed to remain in the natural gas. If the amount of CO₂ sufficiently lowers the heating value of the natural gas, it is typically extracted by amine scrubbing and, in most cases, released into the atmosphere. These emissions can be estimated by calculating the difference between the average carbon dioxide content of raw natural gas and the carbon dioxide content of pipeline gas. The Energy Information Administration (EIA) estimates that annual CO₂ emissions from scrubbing are about 15 Tg CO₂ Eq. Because of imprecision in the reporting of U.S. natural gas production and processing, emissions estimates from energy production sources may be double-counted or under-reported, and thus are uncertain.

Carbon dioxide is also injected into underground deposits to increase crude oil reservoir pressure in a field technique known as enhanced oil recovery (EOR). It is thought that much of the injected CO₂ may be effectively and permanently sequestered, but the fraction of injected CO₂ that is re-released remains uncertain. The fraction re-released varies from well to well depending upon the field geology and the gas capture/re-injection technology employed at the wellhead. Over time, carbon dioxide may also seep into the producing well and mix with the oil and natural gas present there. If the gas portion of this mixture has a sufficiently high energy content, it may be collected and sent to a natural gas plant; if not, it may be vented or flared. The EIA estimates that the amount of CO₂ used for EOR is on the order of 44 Tg CO₂ Eq., of which emissions would be some fraction yet to be defined. This figure is based on the difference between U.S. Department of Commerce sales figures for industrial CO₂ (62 Tg CO₂ Eq.) minus the 18 Tg CO₂ Eq. reported by the Freedomia Group that is used for purposes other than EOR. Further research into EOR is required before the resulting CO₂ emissions can be adequately quantified. (See Carbon Dioxide Consumption in the Industrial Processes chapter).

Carbon Sequestration in Underground Injection Wells

Data for sequestration of carbon in underground injection wells is obtained from the EPA Toxic Release Inventory (EIA 2000). The carbon content of wastes reported in the EPA TRI as being injected into underground injection wells is estimated from the TRI data, and the carbon is assumed to be sequestered. The sequestration of underground injection carbon is one of the many elements in calculating the storage factor for petrochemical feedstock (see Annex B). The "base year" for this storage factor calculation is 1998 and only EPA TRA data for calendar year 1998 is used in the storage factor calculation. Further research is required if the entire time series for this potential sink is to be fully quantified.

CH₄ from Abandoned Coal Mines

Abandoned coal mines are a source of CH₄ emissions. In general, many of the same factors that affect emissions from operating coal mines will affect emissions from abandoned mines such as the permeability and gassiness of the coal, the mine's depth, geologic characteristics, and whether it has been flooded. A few gas developers have recovered methane from abandoned mine workings; therefore, emissions from this source may be significant. Further research and methodological development is needed if these emissions are to be estimated. (See Coal Mining in the Energy chapter.)

CO₂ from “Unaccounted for” Natural Gas

There is a discrepancy between the amount of natural gas sold by producers and that reported as purchased by consumers. This discrepancy, known as “unaccounted for” or unmetered natural gas, was assumed to be the sum of leakage, measurement errors, data collection problems, undetected non-reporting, undetected over reporting, and undetected under reporting. Historically, the amount of gas sold by producers has always exceeded that reportedly purchased by consumers; therefore, some portion of unaccounted for natural gas was assumed to be a source of CO₂ emissions. In other words, it was assumed that consumers were underreporting their usage of natural gas. In DOE/EIA’s energy statistics for 1996, however, reported consumption of natural gas exceeded the amount sold by producers. Therefore, the historical explanation given for this discrepancy has lost credibility and unaccounted for natural gas is no longer used to calculate CO₂ emissions.

CO₂ from Shale Oil Production

Oil shale is shale saturated with kerogen.⁴ It can be thought of as the geological predecessor to crude oil. Carbon dioxide is released as a by-product of the process of producing petroleum products from shale oil. As of now, it is not cost-effective to mine and process shale oil into usable petroleum products. The only identified large-scale oil shale processing facility in the United States was operated by Unocal during the years 1985 to 1990. There have been no known emissions from shale oil processing in the United States since 1990 when the Unocal facility closed.

CH₄ from the Production of Carbides other than Silicon Carbide

Methane (CH₄) may be emitted from the production of carbides because the petroleum coke used in the process contains volatile organic compounds, which form CH₄ during thermal decomposition. Methane emissions from the production of silicon carbide were estimated and accounted for, but emissions from the production of calcium carbide and other carbides were not. Further research is needed to estimate CH₄ emissions from the production of calcium carbide and other carbides other than silicon carbide. (See *Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories: Reference Manual*, pp. 2.20 - 2.21)

CO₂ from Calcium Carbide and Silicon Carbide Production

Carbon dioxide is formed by the oxidation of petroleum coke in the production of both calcium carbide and silicon carbide. These CO₂ emissions are implicitly accounted for with emissions from the combustion of petroleum coke in the Energy chapter. There is currently not sufficient data on coke consumption to estimate emissions from these sources explicitly. (See *Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories: Reference Manual*, pp. 2.20 - 2.21)

CO₂ from Graphite Consumption in Ferroalloy and Steel Production

Emissions from "graphite" "wood" or "biomass" in calculating CO₂ emissions from ferroalloy production, iron and steel production or other "Industrial Processes" included in Chapter 3 of the inventory are not explicitly calculated. It is assumed that 100% of the carbon used in ferroalloy production is derived from petroleum coke and that all of the carbon used in iron and steel production is derived from coal coke or petroleum coke. It is possible that some non-coke carbon is used in the production of ferroalloys and iron and steel, but no data are available to conduct inventory calculations for sources of carbon other than petroleum coke and coal coke used in these processes.

⁴ Kerogen is fossilized insoluble organic material found in sedimentary rocks, usually shales, which can be converted to petroleum products by distillation.

Non-fuel uses of coal coke and petroleum coke are accounted for in the Industrial Process chapter, either directly for iron and steel, aluminum, ferroalloy, and titanium dioxide production, or indirectly by applying a storage factor to "uncharacterized" non-fuel uses of petroleum coke and coal coke.

Non-fuel uses of wood and biomass are not accounted for in the Energy or Industrial Process chapters, as all uses of wood and biomass are accounted for in the Land Use and Forestry chapter. It is assumed for the purposes of the CO₂ emission calculation that no wood or other biogenic carbon is used in any of these industrial processes. Some biogenic carbon may be used in these industrial processes but sufficient data to estimate emissions are not available.

Consumption of either natural or synthetic graphite is not explicitly accounted for in the Industrial Process chapter. It is assumed that all of the carbon used in manufacturing carbon anodes for production of aluminum, ferroalloys, and electric arc furnace (EAF) steel are derived directly from petroleum coke and coal tar pitch (a coal coke byproduct), not from natural graphite or synthetic graphite sources. Some amount of carbon used in these industrial processes may be derived from natural or synthetic graphite sources, but sufficient data to estimate emissions are not currently available.

N₂O from Caprolactam Production

Caprolactam is a widely used chemical intermediate, primarily to produce nylon-6. All processes for producing caprolactam involve the catalytic oxidation of ammonia, with N₂O being produced as a by-product. Caprolactam production could be a significant source of N₂O—it has been identified as such in the Netherlands. More research is required to determine this source's significance because there is currently insufficient information available on caprolactam production to estimate emissions in the United States. (See *Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories: Reference Manual*, pp. 2.22 - 2.23)

N₂O from Cracking of Certain Oil Fractions

In order to improve the gasoline yield in crude oil refining, certain oil fractions are processed in a catcracker. Because crude oil contains some nitrogen, N₂O emissions may result from this cracking process. There is currently insufficient data to develop a methodology for estimating these emissions. (See *Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories: Reference Manual*, p. 2.23)

CH₄ from Coke Production

Coke production may result in CH₄ emissions. Detailed coke production statistics were not available for the purposes of estimating CH₄ emissions from this minor source. (See *Petrochemical Production in the Industrial Processes* chapter and the *Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories: Reference Manual*, p. 2.23)

CO₂ from Metal Production

Coke is used as a reducing agent in the production of some metals from their ores, including magnesium, chromium, lead, nickel, silicon, tin, and zinc. Carbon dioxide may be emitted during the metal's production from the oxidation of this coke and, in some cases, from the carbonate ores themselves (e.g., some magnesium ores contain carbonate). The CO₂ emissions from the carbonate ores are not presently accounted for, but their quantities are thought to be minor. (See *Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories: Reference Manual*, p. 2.37 - 2.38)

N₂O from Acrylonitrile Production

Nitrous oxide may be emitted during acrylonitrile production. No methodology was available for estimating these emissions, and therefore further research is needed if these emissions are to be included. (See *Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories: Reference Manual*, p. 2.22)

SF₆ from Aluminum Fluxing and Degassing

Occasionally, sulfur hexafluoride (SF₆) is used by the aluminum industry as a fluxing and degassing agent in experimental and specialized casting operations. In these cases it is normally mixed with argon, nitrogen, and/or chlorine and blown through molten aluminum; however, this practice is not used by primary aluminum production firms in the United States and is not believed to be extensively used by secondary casting firms. Where it does occur, the concentration of SF₆ in the mixture is small and a portion of the SF₆ is decomposed in the process (Waite and Bernard 1990, Corns 1990). It has been estimated that 230 Mg of SF₆ were used by the aluminum industry in the United States and Canada (Maiss and Brenninkmeijer 1998); however, this estimate is highly uncertain.

SF₆ from Production/Leakage/Breakage of Soundproofed Double-glazed Windows

Sulfur hexafluoride (SF₆) may be emitted from the production, breakage, or leakage of soundproof double-glazed windows. No methodology was available for estimating these emissions, and therefore further research is needed if these emissions are to be included.

SF₆ from Production/Leakage/Dismantling of Radar, Tracer and Night Vision Equipment

Sulfur hexafluoride (SF₆) may be emitted from the production, leakage, and dismantling of radar, tracer, and night vision equipment. Emissions from this source are believed to be minor, and no data were available for estimating the emissions.

SF₆ from Applications in Sports Shoes, Tires, and Tennis Balls

Sulfur hexafluoride (SF₆) may be emitted from application involving the production of sport shoes, tires, and tennis balls. These emissions are believed to be minor, and no data were available for estimating emissions.

SF₆ from Applications to Trace Leakage of Pressure Vessels and Used as a Tracer Gas in Open Air

Sulfur hexafluoride (SF₆) may be emitted from application involving tracer gasses to detect leakage from pressure vessels and as a tracer gas in the open air. Although emissions from this source are believed to be minor, emissions estimation data and methodologies were not available.

Miscellaneous SF₆ Uses

Sulfur hexafluoride may be used in foam insulation, for dry etching, in laser systems, for indoor air quality testing, for laboratory hood testing, for chromatography, in tandem accelerators, in loudspeakers, in shock absorbers, and for certain biomedical applications. Data need to be gathered and methodologies developed if these emissions are to be estimated. A preliminary global assessment of aggregate emissions from these applications can be found in Maiss, M. Brenninkmeijer, and C.A.M. Brenninkmeijer (1998).

CO₂ from Solvent Incineration

Carbon dioxide may be released during the incineration of solvents. Although emissions from this source are believed to be minor, data need to be gathered and methodologies developed if these emissions are to be estimated. Solvents are hazardous wastes, and emissions from solvent incineration were taken into account to estimate the carbon storage factor for hazardous waste incineration. However, sufficient data is not available to obtain a complete time series estimate for this source category. Further research is required for these potential emissions to be fully quantified.

N₂O from Domestic House Animal Waste Deposited on Soils

A substantial amount of liquid and solid waste is produced by domestic animals that are kept as pets. A preliminary methodology was developed to estimate nitrous oxide (N₂O) emissions from the deposition of domestic house animal

(i.e., dogs and cats) waste on lawns, fields and parks. Estimates calculated with this methodology suggest that, in 1990, approximately 330 Gg of nitrogen originating as domestic house animal waste were deposited on soils resulting in approximately 2.9 Tg CO₂ Eq. of N₂O emissions from soils. To estimate the amount of nitrogen deposited by domestic house animals, only those excretions that remained on land surfaces—as opposed to wastes that were collected by owners and are managed as municipal solid waste—were included.

Annual dog and cat population numbers were obtained from the Pet Food Institute.⁵ Annual nitrogen excretion rates were estimated from protein intake. The recommended protein intake for an average size adult of each animal type⁶ was multiplied by the average amount of nitrogen per unit of protein (0.16 kg N/kg protein, from the *Revised 1996 IPCC Guidelines*) to estimate nitrogen consumption. It was then assumed that 95 percent of this nitrogen was excreted, either in solid or liquid form (i.e., it was assumed that 5 percent was retained for fur and milk production). Of the total nitrogen excretion, 90 percent was assumed to occur through liquid waste, with the balance from solid waste⁷. Both cat and dog populations were divided into urban and rural fractions, using the metropolitan and non-metropolitan human population categories, respectively, of the U.S. Census Bureau⁸. Both liquid and solid wastes from the urban cat population, and solid waste from the urban dog population were assumed to be collected (i.e., not deposited on soils). Nitrous oxide emission estimates from domestic house animal excretion were calculated in the same manner as performed for estimating emissions from livestock excretion. Producing these estimates involved making a number of simplifying assumptions regarding average animal size and protein consumption, as well as the proportions of animal populations residing in urban and rural areas and the proportions of wastes that are deposited on land. Further methodological development and data collection is required in order to reduce the uncertainty involved in the domestic house animal excretion estimates.

CO₂ from Food Scraps Disposed in Landfills

A certain amount of food scraps generated from food processing or as leftovers join the waste stream and are landfilled. Nationally, an estimated 0.4 Tg CO₂ Eq. per year are stored in the form of organic carbon contained in food scraps in landfills, acting as a carbon sink. A portion of the landfilled food scraps becomes a source of methane emissions, which offset the sink estimates to an extent. Further data collection on the amount and composition of food scraps generated and landfilled is required in order to reduce the uncertainty associated with this estimate.

CO₂ from Industrial Waste Combustion

Waste combustion is incorporated in two sections of the energy chapter of the inventory: in the section on CO₂ emissions from waste combustion, and in the calculation of emissions and storage from non-energy uses of fossil fuels. The former section addresses fossil-derived materials (such as plastics) that are discarded as part of the municipal wastestream and combusted (generally for energy recovery). The latter addresses two types of combustion: hazardous waste incineration of organic materials (assumed to be fossil-derived), in which regulated wastes are burned without energy recovery, and burning of fossil-derived materials for energy recovery. There is one potentially significant category of waste combustion that is not included in our calculus: industrial non-hazardous waste, burned for disposal (rather than energy recovery). Data are not readily available for this source; further research is needed to estimate the magnitude of CO₂ emissions.

CH₄ from Land-Use Changes Including Wetlands Creation or Destruction

⁵ Pet Food Institute (1999) *Pet Incidence Trend Report*. Pet Food Institute, Washington DC.

⁶ Bright, S. (1999) Personal communication between Marco Alcaraz of ICF Consulting and Susan Bright of the Dupont Animal Clinic, Washington, DC, August 1999.

⁷ Swenson, M.J. and W.G. Reece, eds. (1993) *Duke's Physiology of Domestic Animals*. Cornell University Press. 11th Edition.

⁸ U.S. Census Bureau (1999) <<http://www.census.gov/population/estimates/metro-city/ma96-08.txt>>

Wetlands are a known source of methane (CH₄) emissions. When wetlands are destroyed, CH₄ emissions may be reduced. Conversely, when wetlands are created (e.g., during the construction of hydroelectric plants), CH₄ emissions may increase. Grasslands and forestlands may also be weak sinks for CH₄ due to the presence of methanotrophic bacteria that use CH₄ as an energy source (i.e., they oxidize CH₄ to CO₂). Currently, an adequate scientific basis for estimating these emissions and sinks does not exist, and therefore further research and methodological development is required.

N₂O from Wastewater Treatment and Biological Processes

As a result of nitrification and denitrification processes, nitrous oxide (N₂O) may be produced and emitted from large-scale composting, small scale composting (e.g. households), post-composting of anaerobic digested wastes, and both domestic and industrial wastewater treatment plants. Nitrogen-containing compounds are found in composted wastes and wastewater due to the presence of both human excrement and other nitrogen-containing constituents (e.g. garbage, industrial wastes, animal carcasses, etc.) The portion of emitted N₂O that originates from human excrement is currently estimated under the Human Sewage source category- based upon average dietary assumptions. The portion of emitted N₂O that originates from other nitrogen-containing constituents is not currently estimated. Further research and methodological development is needed if these emissions are to be accurately estimated.

CH₄ from Large and Small Scale Composting

Methane (CH₄) may be released through large and small scale (e.g. household) composting. Detailed composting data is necessary in order to estimate emissions but were not available.

CH₄ from Treatment of Dredging Sludge, Remediation of Groundwater, Intermediate Storage of Slaughter Waste, Production of Process Water from Groundwater, and Post Composting of Anaerobic Digested Wastes.

Methane (CH₄) may be released through the treatment of dredging sludge, remediation of groundwater, intermediate storage of slaughter waste, production of process water from groundwater, and post composting of anaerobic digested wastes. No methodology was available for estimating these emissions, and therefore further research is needed if these emissions are to be included.

N₂O from Applications of Anesthetics in Healthcare, Consumer Packaging of Whipped Cream, Fireworks, Applications as Party-drug/Horns/Balloons, Laboratories, Engine Booster Fuel, and Explosives.

Nitrous oxide (N₂O) may be released from anesthetic in healthcare (i.e. dentists, doctors, veterinarians, and elderly care), consumer packaging of whipped cream, fireworks, applications as party-drug/horns/balloons, laboratories, engine booster fuel, and explosives. Although emissions from these sources are believed to be minor, emissions estimation data and methodologies were not available.

References

EPA (2000b). Toxics Release Inventory, 1998. U.S. Environmental Protection Agency, Office of Environmental Information, Office of Information Analysis and Access, Washington, DC. Available online at <<http://www.epa.gov/triexplorer/chemical.htm>>

ANNEX W: Constants, Units, and Conversions

Metric Prefixes

Although most activity data for the United States is gathered in customary U.S. units, these units are converted into metric units per international reporting guidelines. The following table provides a guide for determining the magnitude of metric units.

Table W-1: Guide to Metric Unit Prefixes

Prefix/Symbol	Factor
atto (a)	10^{-18}
femto (f)	10^{-15}
pico (p)	10^{-12}
nano (n)	10^{-9}
micro (μ)	10^{-6}
milli (m)	10^{-3}
centi (c)	10^{-2}
deci (d)	10^{-1}
deca (da)	10
hecto (h)	10^2
kilo (k)	10^3
mega (M)	10^6
giga (G)	10^9
tera (T)	10^{12}
peta (P)	10^{15}
exa (E)	10^{18}

Unit Conversions

1 kilogram = 2.205 pounds
1 pound = 0.454 kilograms
1 short ton = 2,000 pounds = 0.9072 metric tons
1 metric ton = 1,000 kilograms = 1.1023 short tons

1 cubic meter = 35.315 cubic feet
1 cubic foot = 0.02832 cubic meters
1 U.S. gallon = 3.785412 liters
1 barrel (bbl) = 0.159 cubic meters
1 barrel (bbl) = 42 U.S. gallons
1 liter = 0.1 cubic meters

1 foot = 0.3048 meters
1 meter = 3.28 feet
1 mile = 1.609 kilometers
1 kilometer = 0.622 miles

1 acre = 43,560 square feet = 0.4047 hectares = 4,047 square meters
1 square mile = 2.589988 square kilometers

To convert degrees Fahrenheit to degrees Celsius, subtract 32 and multiply by 5/9
To convert degrees Celsius to Kelvin, add 273.15 to the number of Celsius degrees

Density Conversions¹

Methane 1 cubic meter = 0.67606 kilograms
Carbon dioxide 1 cubic meter = 1.85387 kilograms

Natural gas liquids	1 metric ton	=	11.6 barrels	=	1,844.2 liters
Unfinished oils	1 metric ton	=	7.46 barrels	=	1,186.04 liters
Alcohol	1 metric ton	=	7.94 barrels	=	1,262.36 liters
Liquefied petroleum gas	1 metric ton	=	11.6 barrels	=	1,844.2 liters
Aviation gasoline	1 metric ton	=	8.9 barrels	=	1,415.0 liters
Naphtha jet fuel	1 metric ton	=	8.27 barrels	=	1,314.82 liters
Kerosene jet fuel	1 metric ton	=	7.93 barrels	=	1,260.72 liters
Motor gasoline	1 metric ton	=	8.53 barrels	=	1,356.16 liters
Kerosene	1 metric ton	=	7.73 barrels	=	1,228.97 liters
Naphtha	1 metric ton	=	8.22 barrels	=	1,306.87 liters
Distillate	1 metric ton	=	7.46 barrels	=	1,186.04 liters
Residual oil	1 metric ton	=	6.66 barrels	=	1,058.85 liters
Lubricants	1 metric ton	=	7.06 barrels	=	1,122.45 liters
Bitumen	1 metric ton	=	6.06 barrels	=	963.46 liters
Waxes	1 metric ton	=	7.87 barrels	=	1,251.23 liters
Petroleum coke	1 metric ton	=	5.51 barrels	=	876.02 liters
Petrochemical feedstocks	1 metric ton	=	7.46 barrels	=	1,186.04 liters
Special naphtha	1 metric ton	=	8.53 barrels	=	1,356.16 liters
Miscellaneous products	1 metric ton	=	8.00 barrels	=	1,271.90 liters

Energy Conversions

Converting Various Energy Units to Joules

The common energy unit used in international reports of greenhouse gas emissions is the joule. A joule is the energy required to push with a force of one Newton for one meter. A terajoule (TJ) is one trillion (10^{12}) joules. A British thermal unit (Btu, the customary U.S. energy unit) is the quantity of heat required to raise the temperature of one pound of water one degree Fahrenheit at or near 39.2 Fahrenheit.

1 TJ = 2.388×10¹¹ calories
23.88 metric tons of crude oil equivalent
947.8 million Btus
277,800 kilowatt-hours

Converting Various Physical Units to Energy Units

Data on the production and consumption of fuels are first gathered in physical units. These units must be converted to their energy equivalents. The values in the following table of conversion factors can be used as default factors, if

¹ Reference: EIA (1998a)

local data are not available. See Appendix A of EIA's *Annual Energy Review 1997* (EIA 1998) for more detailed information on the energy content of various fuels.

Table W-2: Conversion Factors to Energy Units (Heat Equivalents)

Fuel Type (Units)	Factor
Solid Fuels (Million Btu/Short ton)	
Anthracite coal	22.573
Bituminous coal	23.89
Sub-bituminous coal	17.14
Lignite	12.866
Coke	24.8
Natural Gas (Btu/Cubic foot)	1,027
Liquid Fuels (Million Btu/Barrel)	
Crude oil	5.800
Natural gas liquids and LRGs	3.777
Other liquids	5.825
Motor gasoline	5.253
Aviation gasoline	5.048
Kerosene	5.670
Jet fuel, kerosene-type	5.670
Distillate fuel	5.825
Residual oil	6.287
Naphtha for petrochemicals	5.248
Petroleum coke	6.024
Other oil for petrochemicals	5.825
Special naphthas	5.248
Lubricants	6.065
Waxes	5.537
Asphalt	6.636
Still gas	6.000
Misc. products	5.796

Note: For petroleum and natural gas, *Annual Energy Review 1997* (EIA 1998b). For coal ranks, *State Energy Data Report 1992* (EIA 1993). All values are given in higher heating values (gross calorific values).

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ANNEX X: Abbreviations

AAPFCO	American Association of Plant Food Control Officials
AFEAS	Alternative Fluorocarbon Environmental Acceptability Study
AGA	American Gas Association
APC	American Plastics Council
ASAE	American Society of Agricultural Engineers
BEA	Bureau of Economic Analysis, U.S. Department of Commerce
BOD ₅	Biochemical oxygen demand over a 5-day period
BTS	Bureau of Transportation Statistics, U.S. Department of Transportation
Btu	British thermal unit
CAAA	Clean Air Act Amendments of 1990
CAPP	Canadian Association of Petroleum Producers
C&EN	Chemical and Engineering News
CFC	Chlorofluorocarbon
CFR	Code of Federal Regulations
CMA	Chemical Manufacturer's Association
CMOP	Coalbed Methane Outreach Program
CVD	Chemical vapor deposition
DESC	Defense Energy Support Center-DoD's defense logistics agency
DIC	Dissolved inorganic carbon
DOC	U.S. Department of Commerce
DoD	U.S. Department of Defense
DOE	U.S. Department of Energy
DOI	U.S. Department of the Interior
DOT	U.S. Department of Transportation
EIA	Energy Information Administration, U.S. Department of Energy
EIIP	Emissions Inventory Improvement Program
EOR	Enhanced oil recovery
EPA	U.S. Environmental Protection Agency
FAA	Federal Aviation Administration
FAO	Food and Agricultural Organization
FCCC	Framework Convention on Climate Change
FEB	Fiber Economics Bureau
FHWA	Federal Highway Administration
GAA	Governmental Advisory Associates
GCV	Gross calorific value
GDP	Gross domestic product
GHG	Greenhouse gas
GRI	Gas Research Institute
GSAM	Gas Systems Analysis Model
GWP	Global warming potential
HBFC	Hydrobromofluorocarbon
HCFC	Hydrochlorofluorocarbon
HDGV	Heavy duty gas vehicle
HDDV	Heavy duty diesel vehicle
HDPE	High density polyethylene
HFC	Hydrofluorocarbon
HFE	Hydrofluoroethers
ICAO	International Civil Aviation Organization
IEA	International Energy Association
IISRP	International Institute of Synthetic Rubber Products
ILENR	Illinois Department of Energy and Natural Resources
IMO	International Maritime Organization

IPAA	Independent Petroleum Association of America
IPCC	Intergovernmental Panel on Climate Change
LDDT	Light duty diesel truck
LDDV	Light duty diesel vehicle
LDGV	Light duty gas vehicle
LDGT	Light duty gas truck
LDPE	Low density polyethylene
LEV	Low emission vehicles
LFG	Landfill gas
LFGTE	Landfill gas-to-energy
LLPDE	Linear low density polyethylene
LMOP	EPA's Landfill Methane Outreach Program
LPG	Liquefied petroleum gas(es)
LTO	Landing and take-off
LULUCF	Land use, land-use change, and forestry
MC	Motorcycle
MCF	Methane conversion factor
MMS	Minerals Management Service
MMTCE	Million metric tons carbon equivalent
MSHA	Mine Safety and Health Administration
MSW	Municipal solid waste
NAPAP	National Acid Precipitation and Assessment Program
NASS	USDA's National Agriculture Statistics Service
NCV	Net calorific value
NIAR	Norwegian Institute for Air Research
NMVOC	Non-methane volatile organic compound
NO _x	Nitrogen Oxides
NRCS	Natural Resources Conservation Service
NSCR	Non-selective catalytic reduction
NVFEL	National Vehicle Fuel Emissions Laboratory
OAQPS	EPA Office of Air Quality Planning and Standards
ODS	Ozone depleting substances
OECD	Organization of Economic Co-operation and Development
OMS	EPA Office of Mobile Sources
ORNL	Oak Ridge National Laboratory
OSHA	Occupational Safety and Health Administration
OTA	Office of Technology Assessment
PPC	Precipitated calcium carbonate
PFC	Perfluorocarbon
PFPE	Perfluoropolyether
POTW	Publicly Owned Treatment Works
ppmv	Parts per million(10 ⁶) by volume
ppbv	Parts per billion (10 ⁹) by volume
pptv	Parts per trillion (10 ¹²) by volume
PVC	Polyvinyl chloride
RCRA	Resource Conservation and Recovery Act
SAE	Society of Automotive Engineers
SBSTA	Subsidiary Body for Scientific and Technical Advice
SCR	Selective catalytic reduction
SNG	Synthetic natural gas
SWANA	Solid Waste Association of North America
TBtu	Trillion Btu
TgCO ₂ Eq	Teragrams carbon dioxide equivalent
TJ	Terajoule
TSDF	Hazardous waste treatment, storage, and disposal facility

TVA	Tennessee Valley Authority
UEP	United Egg Producers
U.S.	United States
USAF	United States Air Force
USDA	United States Department of Agriculture
USFS	United States Forest Service
USGS	United States Geological Survey
UNEP	United Nations Environmental Programme
UNFCCC	United Nations Framework Convention on Climate Change
VAIP	EPA's Voluntary Aluminum Industrial Partnership
VMT	Vehicle miles traveled
WMO	World Meteorological Organization

ANNEX Y: Chemical Formulas

Table Y-1: Guide to Chemical Formulas

Symbol	Name
Al	Aluminum
Al ₂ O ₃	Aluminum Oxide
Br	Bromine
C	Carbon
CH ₄	Methane
C ₂ H ₆	Ethane
C ₃ H ₈	Propane
CF ₄	Perfluoromethane
C ₂ F ₆	Perfluoroethane, hexafluoroethane
c-C ₃ F ₆	Perfluorocyclopropane
C ₃ F ₈	Perfluoropropane
c-C ₄ F ₈	Perfluorocyclobutane
C ₄ F ₁₀	Perfluorobutane
C ₅ F ₁₂	Perfluoropentane
C ₆ F ₁₄	Perfluorohexane
CF ₃ I	Trifluoroiodomethane
CFCl ₃	Trichlorofluoromethane (CFC-11)
CF ₂ Cl ₂	Dichlorodifluoromethane (CFC-12)
CF ₃ Cl	Chlorotrifluoromethane (CFC-13)
C ₂ F ₃ Cl ₃	Trichlorotrifluoroethane (CFC-113)*
CCl ₃ CF ₃	CFC-113a*
C ₂ F ₄ Cl ₂	Dichlorotetrafluoroethane (CFC-114)
C ₂ F ₅ Cl	Chloropentafluoroethane (CFC-115)
CHCl ₂ F	HCFC-21
CHF ₂ Cl	Chlorodifluoromethane (HCFC-22)
C ₂ F ₃ HCl ₂	HCFC-123
C ₂ F ₄ HCl	HCFC-124
C ₂ FH ₃ Cl ₂	HCFC-141b
C ₂ H ₃ F ₂ Cl	HCFC-142b
CF ₃ CF ₂ CHCl ₂	HCFC-225ca
CClF ₂ CF ₂ CHClF	HCFC-225cb
CCl ₄	Carbon tetrachloride
CHClCCl ₂	Trichloroethylene
CCl ₂ CCl ₂	Perchloroethylene, tetrachloroethene
CH ₃ Cl	Methylchloride
CH ₃ CCl ₃	Methylchloroform
CH ₂ Cl ₂	Methylenechloride
CHCl ₃	Chloroform, trichloromethane
CHF ₃	HFC-23
CH ₂ F ₂	HFC-32
CH ₃ F	HFC-41
C ₂ HF ₅	HFC-125
C ₂ H ₂ F ₄	HFC-134
CH ₂ FCF ₃	HFC-134a
C ₂ H ₃ F ₃	HFC-143*
C ₂ H ₃ F ₃	HFC-143a*
CH ₂ FCH ₂ F	HFC-152*
C ₂ H ₄ F ₂	HFC-152a*

CH ₃ CH ₂ F	HFC-161
C ₃ HF ₇	HFC-227ea
CF ₃ CF ₂ CH ₂ F	HFC-236cb
CF ₃ CHFCHF ₂	HFC-236ea
C ₃ H ₂ F ₆	HFC-236fa
C ₃ H ₃ F ₅	HFC-245ca
CHF ₂ CH ₂ CF ₃	HFC-245fa
CF ₃ CH ₂ CF ₂ CH ₃	HFC-365mfc
C ₅ H ₂ F ₁₀	HFC-43-10mee
CF ₃ OCHF ₂	HFE-125
CF ₂ HOCHF ₂ H	HFE-134
CH ₃ OCF ₃	HFE-143a
CF ₃ CHFOCF ₃	HFE-227ea
CF ₃ CHClOCHF ₂	HCFE-235da2
CF ₃ CHFOCHF ₂	HFE-236ea2
CF ₃ CH ₂ OCF ₃	HFE-236fa
CF ₃ CF ₂ OCH ₃	HFE-245cb2
CHF ₂ CH ₂ OCF ₃	HFE-245fa1
CF ₃ CH ₂ OCHF ₂	HFE-245fa2
CHF ₂ CF ₂ OCH ₃	HFE-254cb2
CF ₃ CH ₂ OCH ₃	HFE-263fb2
CF ₃ CF ₂ OCF ₂ CHF ₂	HFE-329mcc2
CF ₃ CF ₂ OCH ₂ CF ₃	HFE-338mcf2
CF ₃ CF ₂ CF ₂ OCH ₃	HFE-347mcc3
CF ₃ CF ₂ OCH ₂ CHF ₂	HFE-347mcf2
CF ₃ CHF ₂ CF ₂ OCH ₃	HFE-356mec3
CHF ₂ CF ₂ CF ₂ OCH ₃	HFE-356pcc3
CHF ₂ CF ₂ OCH ₂ CHF ₂	HFE-356pcf2
CHF ₂ CF ₂ CH ₂ OCHF ₂	HFE-356pcf3
CF ₃ CF ₂ CH ₂ OCH ₃	HFE-365mcf3
CHF ₂ CF ₂ OCH ₂ CH ₃	HFE-374pcf2
C ₄ F ₉ OCH ₃	HFE-7100
C ₄ F ₉ OC ₂ H ₅	HFE-7200
CHF ₂ OCF ₂ OC ₂ F ₄ OCHF ₂	H-Galden 1040x
CHF ₂ OCF ₂ OCHF ₂	HG-10
CHF ₂ OCF ₂ CF ₂ OCHF ₂	HG-01
CH ₃ OCH ₃	Dimethyl ether
CH ₂ Br ₂	Dibromomethane
CH ₂ BrCl	Dibromochloromethane
CHBr ₃	Tribromomethane
CHBrF ₂	Bromodifluoromethane
CH ₃ Br	Methylbromide
CF ₂ BrCl	Bromodichloromethane (Halon 1211)
CF ₃ Br(CBrF ₃)	Bromotrifluoromethane (Halon 1301)
CF ₃ I	FIC-131I
CO	Carbon monoxide
CO ₂	Carbon dioxide
CaCO ₃	Calcium carbonate, Limestone
CaMg(CO ₃) ₂	Dolomite
CaO	Calcium oxide, Lime
Cl	atomic Chlorine
F	Fluorine
Fe	Iron
Fe ₂ O ₃	Ferric oxide
FeSi	Ferrosilicon

H, H ₂	atomic Hydrogen, molecular Hydrogen
H ₂ O	Water
H ₂ O ₂	Hydrogen peroxide
OH	Hydroxyl
N, N ₂	atomic Nitrogen, molecular Nitrogen
NH ₃	Ammonia
NH ₄ ⁺	Ammonium ion
HNO ₃	Nitric Acid
NF ₃	Nitrogen trifluoride
N ₂ O	Nitrous oxide
NO	Nitric oxide
NO ₂	Nitrogen dioxide
NO ₃	Nitrate radical
Na	Sodium
Na ₂ CO ₃	Sodium carbonate, soda ash
Na ₃ AlF ₆	Synthetic cryolite
O, O ₂	atomic Oxygen, molecular Oxygen
O ₃	Ozone
S	atomic Sulfur
H ₂ SO ₄	Sulfuric acid
SF ₆	Sulfur hexafluoride
SF ₅ CF ₃	Trifluoromethylsulphur pentafluoride
SO ₂	Sulfur dioxide
Si	Silicon
SiC	Silicon carbide
SiO ₂	Quartz

* Distinct isomers.

ANNEX Z: Glossary

Abiotic.⁷ Nonliving. Compare *biotic*.

Absorption of radiation.¹ The uptake of radiation by a solid body, liquid or gas. The absorbed energy may be transferred or re-emitted.

Acid deposition.⁶ A complex chemical and atmospheric process whereby recombined emissions of sulfur and nitrogen compounds are redeposited on earth in wet or dry form. See *acid rain*.

Acid rain.⁶ Rainwater that has an acidity content greater than the postulated natural pH of about 5.6. It is formed when sulfur dioxides and nitrogen oxides, as gases or fine particles in the atmosphere, combine with water vapor and precipitate as sulfuric acid or nitric acid in rain, snow, or fog. The dry forms are acidic gases or particulates. See *acid deposition*.

Acid solution.⁷ Any water solution that has more hydrogen ions (H⁺) than hydroxide ions (OH⁻); any water solution with a pH less than 7. See *basic solution, neutral solution*.

Acidic.⁷ See acid solution.

Adiabatic process.⁹ A thermodynamic change of state of a system such that no heat or mass is transferred across the boundaries of the system. In an adiabatic process, expansion always results in cooling, and compression in warming.

Aerosol.^{1&9} Particulate matter, solid or liquid, larger than a molecule but small enough to remain suspended in the atmosphere. Natural sources include salt particles from sea spray, dust and clay particles as a result of weathering of rocks, both of which are carried upward by the wind. Aerosols can also originate as a result of human activities and are often considered pollutants. Aerosols are important in the atmosphere as nuclei for the condensation of water droplets and ice crystals, as participants in various chemical cycles, and as absorbers and scatters of solar radiation, thereby influencing the radiation budget of the Earth's climate system. See *climate, particulate matter*.

Afforestation.² Planting of new forests on lands that have not been recently forested.

Air carrier⁸ An operator (e.g., airline) in the commercial system of air transportation consisting of aircraft that hold certificates of, Public Convenience and Necessity, issued by the Department of Transportation, to conduct scheduled or non-scheduled flights within the country or abroad.

Air pollutant. See *air pollution*.

Air pollution.⁷ One or more chemicals or substances in high enough concentrations in the air to harm humans, other animals, vegetation, or materials. Such chemicals or physical conditions (such as excess heat or noise) are called air pollutants.

Albedo.⁹ The fraction of the total solar radiation incident on a body that is reflected by it.

Alkalinity.⁶ Having the properties of a base with a pH of more than 7. A common alkaline is baking soda.

Alternative energy.⁶ Energy derived from nontraditional sources (e.g., compressed natural gas, solar, hydroelectric, wind).

Anaerobic.⁶ A life or process that occurs in, or is not destroyed by, the absence of oxygen.

Anaerobic decomposition.² The breakdown of molecules into simpler molecules or atoms by microorganisms that can survive in the partial or complete absence of oxygen.

Anaerobic lagoon.² A liquid-based manure management system, characterized by waste residing in water to a depth of at least six feet for a period ranging between 30 and 200 days. Bacteria produce methane in the absence of oxygen while breaking down waste.

Anaerobic organism.⁷ Organism that does not need oxygen to stay alive. See *aerobic organism*.

Antarctic "Ozone Hole."⁶ Refers to the seasonal depletion of stratospheric ozone in a large area over Antarctica. See *ozone layer*.

Anthracite.² A hard, black, lustrous coal containing a high percentage of fixed carbon and a low percentage of volatile matter. Often referred to as hard coal.

Anthropogenic.² Human made. In the context of greenhouse gases, emissions that are produced as the result of human activities.

Arable land.⁷ Land that can be cultivated to grow crops.

Aromatic.⁶ Applied to a group of hydrocarbons and their derivatives characterized by the presence of the benzene ring.

Ash.⁶ The mineral content of a product remaining after complete combustion.

Asphalt.² A dark-brown-to-black cement-like material containing bitumen as the predominant constituent. It is obtained by petroleum processing. The definition includes crude asphalt as well as the following finished products: cements, fluxes, the asphalt content of emulsions (exclusive of water), and petroleum distillates blended with asphalt to make cutback asphalt.

Atmosphere.¹ The mixture of gases surrounding the Earth. The Earth's atmosphere consists of about 79.1 percent nitrogen (by volume), 20.9 percent oxygen, 0.036 percent carbon dioxide and trace amounts of other gases. The atmosphere can be divided into a number of layers according to its mixing or chemical characteristics, generally determined by its thermal properties (temperature). The layer nearest the Earth is the *troposphere*, which reaches up to an altitude of about 8 kilometers (about 5 miles) in the polar regions and up to 17 kilometers (nearly 11 miles) above the equator. The *stratosphere*, which reaches to an altitude of about 50 kilometers (31 miles) lies atop the troposphere. The *mesosphere*, which extends from 80 to 90 kilometers atop the stratosphere, and finally, the *thermosphere*, or *ionosphere*, gradually diminishes and forms a fuzzy border with outer space. There is relatively little mixing of gases between layers.

Atmospheric lifetime. See *lifetime*.

Atomic weight.⁶ The average weight (or mass) of all the isotopes of an element, as determined from the proportions in which they are present in a given element, compared with the mass of the 12 isotope of carbon (taken as precisely 12.000), that is the official international standard; measured in daltons.

Atoms.⁷ Minute particles that are the basic building blocks of all chemical elements and thus all matter.

Aviation Gasoline.⁸ All special grades of gasoline for use in aviation reciprocating engines, as given in the American Society for Testing and Materials (ASTM) specification D 910. Includes all refinery products within the gasoline range that are to be marketed straight or in blends as aviation gasoline without further processing (any refinery operation except mechanical blending). Also included are finished components in the gasoline range, which will be used for blending or compounding into aviation gasoline.

Bacteria.⁷ One-celled organisms. Many act as decomposers that break down dead organic matter into substances that dissolve in water and are used as nutrients by plants.

Barrel (bbl).⁶ A liquid-volume measure equal to 42 United States gallons at 60 degrees Fahrenheit; used in expressing quantities of petroleum-based products.

Basic solution.⁷ Water solution with more hydroxide ions (OH⁻) than hydrogen ions (H⁺); water solutions with pH greater than 7. See *acid solution*, *alkalinity*, *acid*.

Biodegradable.⁷ Material that can be broken down into simpler substances (elements and compounds) by bacteria or other decomposers. Paper and most organic wastes such as animal manure are biodegradable. See *nonbiodegradable*.

Biofuel.^{3&7} Gas or liquid fuel made from plant material (biomass). Includes wood, wood waste, wood liquors, peat, railroad ties, wood sludge, spent sulfite liquors, agricultural waste, straw, tires, fish oils, tall oil, sludge waste, waste alcohol, municipal solid waste, landfill gases, other waste, and ethanol blended into motor gasoline.

Biogeochemical cycle.⁷ Natural processes that recycle nutrients in various chemical forms from the environment, to organisms, and then back to the environment. Examples are the carbon, oxygen, nitrogen, phosphorus, and hydrologic cycles.

Biological oxygen demand (BOD).⁷ Amount of dissolved oxygen needed by aerobic decomposers to break down the organic materials in a given volume of water at a certain temperature over a specified time period. See *BOD5*.

Biomass.⁷ Total dry weight of all living organisms that can be supported at each trophic level in a food chain. Also, materials that are biological in origin, including organic material (both living and dead) from above and below ground, for example, trees, crops, grasses, tree litter, roots, and animals and animal waste.

Biomass energy.¹ Energy produced by combusting biomass materials such as wood. The carbon dioxide emitted from burning biomass will not increase total atmospheric carbon dioxide if this consumption is done on a sustainable basis (i.e., if in a given period of time, regrowth of biomass takes up as much carbon dioxide as is released from biomass combustion). Biomass energy is often suggested as a replacement for fossil fuel combustion.

Biosphere.^{2&7} The living and dead organisms found near the earth's surface in parts of the lithosphere, atmosphere, and hydrosphere. The part of the global carbon cycle that includes living organisms and biogenic organic matter.

Biotic.⁷ Living. Living organisms make up the biotic parts of ecosystems. See *abiotic*.

Bitumen.⁷ Goopy, black, high-sulfur, heavy oil extracted from tar sand and then upgraded to synthetic fuel oil. See *tar sand*.

Bituminous coal.² A dense, black, soft coal, often with well-defined bands of bright and dull material. The most common coal, with moisture content usually less than 20 percent. Used for generating electricity, making coke, and space heating.

BOD5.² The biochemical oxygen demand of wastewater during decomposition occurring over a 5-day period. A measure of the organic content of wastewater. See *biological oxygen demand*.

British thermal unit (Btu).³ The quantity of heat required to raise the temperature of one pound of water one degree of Fahrenheit at or near 39.2 degrees Fahrenheit.

Bunker fuel.² Fuel supplied to ships and aircraft for international transportation, irrespective of the flag of the carrier, consisting primarily of residual and distillate fuel oil for ships and jet fuel for aircraft.

Bus.^{6&8} A rubber-tired, self-propelled, manually steered vehicle that is generally designed to transport 30 individuals or more. Bus types include intercity, school and transit.

Capacity Factor.³ The ratio of the electrical energy produced by a generating unit for a given period of time to the electrical energy that could have been produced at continuous full- power operation during the same period.

Carbon black.² An amorphous form of carbon, produced commercially by thermal or oxidative decomposition of hydrocarbons and used principally in rubber goods, pigments, and printer's ink.

Carbon cycle.² All carbon reservoirs and exchanges of carbon from reservoir to reservoir by various chemical, physical, geological, and biological processes. Usually thought of as a series of the four main reservoirs of carbon interconnected by pathways of exchange. The four reservoirs, regions of the Earth in which carbon behaves in a systematic manner, are the atmosphere, terrestrial biosphere (usually includes freshwater systems), oceans, and sediments (includes fossil fuels). Each of these global reservoirs may be subdivided into smaller pools, ranging in size from individual communities or ecosystems to the total of all living organisms (biota).

Carbon dioxide.² A colorless, odorless, non-poisonous gas that is a normal part of the ambient air. Carbon dioxide is a product of fossil fuel combustion. Although carbon dioxide does not directly impair human health, it is a greenhouse gas that traps terrestrial (i.e., infrared) radiation and contributes to the potential for global warming. See *global warming*.

Carbon equivalent (CE).¹ A metric measure used to compare the emissions of the different greenhouse gases based upon their global warming potential (GWP). Greenhouse gas emissions in the United States are most commonly expressed as "million metric tons of carbon equivalents" (MMTCE). Global warming potentials are used to convert greenhouse gases to carbon dioxide equivalents. See *global warming potential, greenhouse gas*.

Carbon flux.⁹ The rate of exchange of carbon between pools (i.e., reservoirs).

Carbon intensity. The relative amount of carbon emitted per unit of energy or fuels consumed.

Carbon pool.⁹ The reservoir containing carbon as a principal element in the geochemical cycle.

Carbon sequestration.¹ The uptake and storage of carbon. Trees and plants, for example, absorb carbon dioxide, release the oxygen and store the carbon. Fossil fuels were at one time biomass and continue to store the carbon until burned. See *carbon sinks*.

Carbon sinks.¹ Carbon reservoirs and conditions that take-in and store more carbon (i.e., carbon sequestration) than they release. Carbon sinks can serve to partially offset greenhouse gas emissions. Forests and oceans are large carbon sinks. See *carbon sequestration*.

Carbon tetrachloride (CCl₄).¹¹ A compound consisting of one carbon atom and four chlorine atoms. It is an ozone depleting substance. Carbon tetrachloride was widely used as a raw material in many industrial applications, including the production of chlorofluorocarbons, and as a solvent. Solvent use was ended in the United States when it was discovered to be carcinogenic. See *ozone depleting substance*.

Chemical reaction.⁷ Interaction between chemicals in which there is a change in the chemical composition of the elements or compounds involved.

Chlorofluorocarbons (CFCs).⁷ Organic compounds made up of atoms of carbon, chlorine, and fluorine. An example is CFC-12 (CCl₂F₂), used as a refrigerant in refrigerators and air conditioners and as a foam blowing agent. Gaseous CFCs can deplete the ozone layer when they slowly rise into the stratosphere, are broken down by strong ultraviolet radiation, release chlorine atoms, and then react with ozone molecules. See *Ozone Depleting Substance*.

Climate.^{1&9} The average weather, usually taken over a 30 year time period, for a particular region and time period. Climate is not the same as weather, but rather, it is the average pattern of weather for a particular region. Weather describes the short-term state of the atmosphere. Climatic elements include precipitation, temperature, humidity, sunshine, wind velocity, phenomena such as fog, frost, and hail-storms, and other measures of the weather. See *weather*.

Climate change.¹ The term “climate change” is sometimes used to refer to all forms of climatic inconsistency, but because the Earth’s climate is never static, the term is more properly used to imply a significant change from one climatic condition to another. In some cases, “climate change” has been used synonymously with the term, “global warming”; scientists however, tend to use the term in the wider sense to also include natural changes in climate. See *global warming, greenhouse effect, enhanced greenhouse effect, radiative forcing*.

Climate feedback.¹ An atmospheric, oceanic, terrestrial, or other process that is activated by direct climate change induced by changes in radiative forcing. Climate feedbacks may increase (positive feedback) or diminish (negative feedback) the magnitude of the direct climate change.

Climate lag.¹ The delay that occurs in climate change as a result of some factor that changes very slowly. For example, the effects of releasing more carbon dioxide into the atmosphere may not be known for some time because a large fraction is dissolved in the ocean and only released to the atmosphere many years later.

Climate sensitivity.¹ The equilibrium response of the climate to a change in radiative forcing; for example, a doubling of the carbon dioxide concentration. See *radiative forcing*.

Climate system (or Earth system).¹ The atmosphere, the oceans, the biosphere, the cryosphere, and the geosphere, together make up the climate system.

Coal.² A black or brownish black solid, combustible substance formed by the partial decomposition of vegetable matter without access to air. The rank of coal, which includes anthracite, bituminous coal, subbituminous coal, and lignite, is based on fixed carbon, volatile matter, and heating value. Coal rank indicates the progressive alteration, or coalification, from lignite to anthracite. See *anthracite, bituminous coal, subbituminous coal, lignite*.

Coal coke.² A hard, porous product made from baking bituminous coal in ovens at temperatures as high as 2,000 degrees Fahrenheit. It is used both as a fuel and as a reducing agent in smelting iron ore in a blast furnace.

Coal gasification.⁷ Conversion of solid coal to synthetic natural gas (SNG) or a gaseous mixture that can be burned as a fuel.

Coal liquefaction.⁷ Conversion of solid coal to a liquid fuel such as synthetic crude oil or methanol.

Coalbed methane.² Methane that is produced from coalbeds in the same manner as natural gas produced from other strata. Methane is the principal component of natural gas.

Co-control benefit.¹⁰ It is the additional benefit derived from an environmental policy that is designed to control one type of pollution, while reducing the emissions of other pollutants as well. For example, a policy to reduce carbon dioxide emissions might reduce the combustion of coal, but when coal combustion is reduced, so too are the emissions of particulates and sulfur dioxide. The benefits associated with reductions in emissions of particulates and sulfur dioxide are the co-control benefits of reductions in carbon dioxide.

Cogeneration.⁷ Production of two useful forms of energy such as high-temperature heat and electricity from the same process.

Combustion.² Chemical oxidation accompanied by the generation of light and heat.

Commercial End-Use Sector: Defined economically, consists of business establishments that are not engaged in transportation or in manufacturing or other types of industrial activities (e.g., agriculture, mining, or construction). Commercial establishments include hotels, motels, restaurants, wholesale businesses, retail stores, laundries, and other service enterprises; religious and nonprofit organizations; health, social, and educational institutions; and Federal, State, and local governments. Street lights, pumps, bridges, and public services are also included if the establishment operating them is considered commercial.

Compost.⁷ Partially decomposed organic plant and animal matter that can be used as a soil conditioner or fertilizer.

Composting.⁷ Partial breakdown of organic plant and animal matter by aerobic bacteria to produce a material that can be used as a soil conditioner or fertilizer. See *compost*.

Compound.⁷ Combination of two or more different chemical elements held together by chemical bonds. See *element*. See *inorganic compound, organic compound*.

Concentration.⁷ Amount of a chemical in a particular volume or weight of air, water, soil, or other medium. See *parts per billion, parts per million*.

Conference Of Parties (COP).¹⁰ The supreme body of the United Nations Framework Convention on Climate Change (UNFCCC). It comprises more than 170 nations that have ratified the Convention. Its first session was held in Berlin, Germany, in 1995 and is expected to continue meeting on a yearly basis. The COP’s role is to promote and review the implementation of the Convention. It will periodically review existing commitments in light of the Convention’s objective, new scientific findings, and the effectiveness of national climate change programs. See *United Nations Framework Convention on Climate Change*.

Conifer.⁷ See *coniferous trees*.

Coniferous trees.⁷ Cone-bearing trees, mostly evergreens, that have needle-shaped or scale-like leaves. They produce wood known commercially as softwood. See *deciduous trees*.

Cooling Degree Days: The number of degrees per day that the average daily temperature is above 65° Fahrenheit. The daily average temperature is the mean of the maximum and minimum temperatures for a 24 hour period. (See Degree Days)

Criteria pollutant.² A pollutant determined to be hazardous to human health and regulated under EPA's National Ambient Air Quality Standards. The 1970 amendments to the Clean Air Act require EPA to describe the health and welfare impacts of a pollutant as the "criteria" for inclusion in the regulatory regime. In this report, emissions of the criteria pollutants CO, NO_x, NMVOCs, and SO₂ are reported because they are thought to be precursors to greenhouse gas formation.

Crop residue.² Organic residue remaining after the harvesting and processing of a crop.

Crop rotation.⁷ Planting the same field or areas of fields with different crops from year to year to reduce depletion of soil nutrients. A plant such as corn, tobacco, or cotton, which remove large amounts of nitrogen from the soil, is planted one year. The next year a legume such as soybeans, which add nitrogen to the soil, is planted.

Crude oil.² A mixture of hydrocarbons that exist in liquid phase in underground reservoirs and remain liquid at atmospheric pressure after passing through surface separating facilities. See *petroleum*.

Deciduous trees.⁷ Trees such as oaks and maples that lose their leaves during part of the year. See *coniferous trees*.

Decomposition.⁹ The breakdown of matter by bacteria and fungi. It changes the chemical composition and physical appearance of the materials.

Deforestation.¹ Those practices or processes that result in the conversion of forested lands for non-forest uses. This is often cited as one of the major causes of the enhanced greenhouse effect for two reasons: 1) the burning or decomposition of the wood releases carbon dioxide; and 2) trees that once removed carbon dioxide from the atmosphere in the process of photosynthesis are no longer present.

Degradable.⁷ See *biodegradable*.

Degree Days (Population Weighted): Heating or cooling degree days weighted by the population of the area in which the degree days are recorded. To compute State population-weighted degree days, each State is divided into from one to nine climatically homogeneous divisions, which are assigned weights based on the ratio of the population of the division to the total population of the State. Degree day readings for each division are multiplied by the corresponding population weight for each division and those products are then summed to arrive at the State population-weighted degree day value. To compute national population-weighted degree days, the Nation is divided into nine Census regions, each comprising from three to eight States, which are assigned weights based on the ratio of the population of the Nation. Degree day readings for each region are multiplied by the corresponding population weight for each region and those products are then summed to arrive at the national population-weighted degree day value. (See Heating Degree Days, Cooling Degree Days, and Degree Day Normals)

Degree Day Normals: Simple arithmetic averages of monthly or annual degree days over a long period of time (usually the 30 year period of 1961 through 1990). The averages may be simple degree day normals or population-weighted degree day normals.

Desertification.¹ The progressive destruction or degradation of existing vegetative cover to form a desert. This can occur due to overgrazing, deforestation, drought, and the burning of extensive areas. Once formed, deserts can only support a sparse range of vegetation. Climatic effects associated with this phenomenon include increased reflectivity of solar radiation, reduced atmospheric humidity, and greater atmospheric dust (aerosol) loading.

Distillate fuel oil.² A general classification for the petroleum fractions produced in conventional distillation operations. Included are products known as No. 1, No. 2, and No. 4 fuel oils and No. 1, No. 2, and No. 4 diesel fuels. Used primarily for space heating, on and off-highway diesel engine fuel (including railroad engine fuel and fuel for agricultural machinery), and electric power generation.

Economy.⁷ System of production, distribution, and consumption of economic goods.

Ecosystem.¹⁰ The complex system of plant, animal, fungal, and microorganism communities and their associated non-living environment interacting as an ecological unit. Ecosystems have no fixed boundaries; instead their parameters are set to the scientific, management, or policy question being examined. Depending upon the purpose of analysis, a single lake, a watershed, or an entire region could be considered an ecosystem.

Electric Utility Sector: Privately and publicly owned establishments that generate, transmit, distribute, or sell electricity primarily for use by the public and meet the definition of an electric utility. Electric utilities include investor-owned, publicly owned, cooperative, and Federal utilities. Historically, they have generally been vertically

integrated companies that provide for generation, transmission, distribution, and/or energy services for all customers in a designated service territory. Nonutility power producers are not included in the electric utility sector.

Electrons.⁷ Tiny particle moving around outside the nucleus of an atom. Each electron has one unit of negative charge (-) and almost no mass.

Element.⁷ Chemicals such as hydrogen (H), iron (Fe), sodium (Na), carbon (C), nitrogen (N), or oxygen (O), whose distinctly different atoms serve as the basic building blocks of all matter. There are 92 naturally occurring elements. Another 15 have been made in laboratories. Two or more elements combine to form compounds that make up most of the world's matter. See *compound*.

Emission inventory. A list of air pollutants emitted into a community's, state's, nation's, or the Earth's atmosphere in amounts per some unit time (e.g. day or year) by type of source. An emission inventory has both political and scientific applications.

Emissions coefficient/factor.² A unique value for scaling emissions to activity data in terms of a standard rate of emissions per unit of activity (e.g., grams of carbon dioxide emitted per barrel of fossil fuel consumed).

Emissions.² Releases of gases to the atmosphere (e.g., the release of carbon dioxide during fuel combustion). Emissions can be either intended or unintended releases. See *fugitive emissions*.

Energy conservation.⁷ Reduction or elimination of unnecessary energy use and waste. See *energy-efficiency*.

Energy intensity.⁵ Ratio between the consumption of energy to a given quantity of output; usually refers to the amount of primary or final energy consumed per unit of gross domestic product.

Energy quality.⁷ Ability of a form of energy to do useful work. High-temperature heat and the chemical energy in fossil fuels and nuclear fuels are concentrated high quality energy. Low-quality energy such as low-temperature heat is dispersed or diluted and cannot do much useful work.

Energy.³ The capacity for doing work as measured by the capability of doing work (potential energy) or the conversion of this capability to motion (kinetic energy). Energy has several forms, some of which are easily convertible and can be changed to another form useful for work. Most of the world's convertible energy comes from fossil fuels that are burned to produce heat that is then used as a transfer medium to mechanical or other means in order to accomplish tasks. In the United States, electrical energy is often measured in kilowatt-hours (kWh), while heat energy is often measured in British thermal units (Btu).

Energy-efficiency.^{6&8} The ratio of the useful output of services from an article of industrial equipment to the energy use by such an article; for example, vehicle miles traveled per gallon of fuel (mpg).

Enhanced greenhouse effect.¹ The concept that the natural greenhouse effect has been enhanced by anthropogenic emissions of greenhouse gases. Increased concentrations of carbon dioxide, methane, and nitrous oxide, CFCs, HFCs, PFCs, SF₆, NF₃, and other photochemically important gases caused by human activities such as fossil fuel consumption, trap more infra-red radiation, thereby exerting a warming influence on the climate. See *greenhouse gas, anthropogenic, greenhouse effect, climate*.

Enhanced oil recovery.⁷ Removal of some of the heavy oil left in an oil well after primary and secondary recovery. See *primary oil recovery, secondary oil recovery*.

Enteric fermentation.² A digestive process by which carbohydrates are broken down by microorganisms into simple molecules for absorption into the bloodstream of an animal.

Environment.⁷ All external conditions that affect an organism or other specified system during its lifetime.

Ethanol (C₂H₅OH).⁸ Otherwise known as ethyl alcohol, alcohol, or grain spirit. A clear, colorless, flammable oxygenated hydrocarbon with a boiling point of 78.5 degrees Celsius in the anhydrous state. In transportation, ethanol is used as a vehicle fuel by itself (E100), blended with gasoline (E85), or as a gasoline octane enhancer and oxygenate (10 percent concentration).

Evapotranspiration.¹⁰ The loss of water from the soil by evaporation and by transpiration from the plants growing in the soil, which rises with air temperature.

Exponential growth.⁷ Growth in which some quantity, such as population size, increases by a constant percentage of the whole during each year or other time period; when the increase in quantity over time is plotted, this type of growth yields a curve shaped like the letter J.

Feedlot.⁷ Confined outdoor or indoor space used to raise hundreds to thousands of domesticated livestock. See *rangeland*.

Fertilization, carbon dioxide.¹ An expression (sometimes reduced to 'fertilization') used to denote increased plant growth due to a higher carbon dioxide concentration.

Fertilizer.⁷ Substance that adds inorganic or organic plant nutrients to soil and improves its ability to grow crops, trees, or other vegetation. See *organic fertilizer*.

Flaring.⁹ The burning of waste gases through a flare stack or other device before releasing them to the air.

Fluidized bed combustion (FBC).⁷ Process for burning coal more efficiently, cleanly, and cheaply. A stream of hot air is used to suspend a mixture of powdered coal and limestone during combustion. About 90 to 98 percent of the sulfur dioxide produced during combustion is removed by reaction with limestone to produce solid calcium sulfate.

Fluorocarbons.¹ Carbon-fluorine compounds that often contain other elements such as hydrogen, chlorine, or bromine. Common fluorocarbons include chlorofluorocarbons (CFCs), hydrochlorofluorocarbons (HCFCs), hydrofluorocarbons (HFCs), and perfluorocarbons (PFCs). See *chlorofluorocarbons, hydrochlorofluorocarbons, hydrofluorocarbons, perfluorocarbons*.

Forcing mechanism.¹ A process that alters the energy balance of the climate system (i.e., changes the relative balance between incoming solar radiation and outgoing infrared radiation from Earth). Such mechanisms include changes in solar irradiance, volcanic eruptions, and enhancement of the natural greenhouse effect by emission of carbon dioxide.

Forest.⁷ Terrestrial ecosystem (biome) with enough average annual precipitation (at least 76 centimeters or 30 inches) to support growth of various species of trees and smaller forms of vegetation.

Fossil fuel. A general term for buried combustible geologic deposits of organic materials, formed from decayed plants and animals that have been converted to crude oil, coal, natural gas, or heavy oils by exposure to heat and pressure in the earth's crust over hundreds of millions of years. See *coal, petroleum, crude oil, natural gas*.

Fossil fuel combustion.¹ Burning of coal, oil (including gasoline), or natural gas. The burning needed to generate energy release carbon dioxide by-products that can include unburned hydrocarbons, methane, and carbon monoxide. Carbon monoxide, methane, and many of the unburned hydrocarbons slowly oxidize into carbon dioxide in the atmosphere. Common sources of fossil fuel combustion include cars and electric utilities.

Freon. See chlorofluorocarbon.

Fugitive emissions.² Unintended gas leaks from the processing, transmission, and/or transportation of fossil fuels, CFCs from refrigeration leaks, SF₆ from electrical power distributor, etc.

Gasohol.⁷ Vehicle fuel consisting of a mixture of gasoline and ethyl or methyl alcohol; typically 10 to 23 percent ethanol by volume.

General Aviation.⁸ That portion of civil aviation, which encompasses all facets of aviation except air carriers. It includes any air taxis, commuter air carriers, and air travel clubs, which do not hold Certificates of Public Convenience and Necessity. See *air carriers*.

General circulation model (GCM).¹ A global, three-dimensional computer model of the climate system which can be used to simulate human-induced climate change. GCMs are highly complex and they represent the effects of such factors as reflective and absorptive properties of atmospheric water vapor, greenhouse gas concentrations, clouds, annual and daily solar heating, ocean temperatures and ice boundaries. The most recent GCMs include global representations of the atmosphere, oceans, and land surface.

Geosphere.¹ The soils, sediments, and rock layers of the Earth's crust, both continental and beneath the ocean floors.

Geothermal energy.⁷ Heat transferred from the earth's molten core to under-ground deposits of dry steam (steam with no water droplets), wet steam (a mixture of steam and water droplets), hot water, or rocks lying fairly close to the earth's surface.

Global Warming Potential (GWP).¹ The index used to translate the level of emissions of various gases into a common measure in order to compare the relative radiative forcing of different gases without directly calculating the changes in atmospheric concentrations. GWPs are calculated as the ratio of the radiative forcing that would result from the emissions of one kilogram of a greenhouse gas to that from the emission of one kilogram of carbon dioxide over a period of time (usually 100 years). Gases involved in complex atmospheric chemical processes have not been assigned GWPs. See *lifetime*.

Global warming.¹⁰ The progressive gradual rise of the earth's surface temperature thought to be caused by the greenhouse effect and responsible for changes in global climate patterns. See *enhanced greenhouse effect, greenhouse effect, climate change*.

Grassland.⁷ Terrestrial ecosystem (biome) found in regions where moderate annual average precipitation (25 to 76 centimeters or 10 to 30 inches) is enough to support the growth of grass and small plants but not enough to support large stands of trees.

Greenhouse effect.⁷ Trapping and build-up of heat in the atmosphere (troposphere) near the earth's surface. Some of the heat flowing back toward space from the earth's surface is absorbed by water vapor, carbon dioxide, ozone,

and several other gases in the atmosphere and then reradiated back toward the earth's surface. If the atmospheric concentrations of these greenhouse gases rise, the average temperature of the lower atmosphere will gradually increase. See *enhanced greenhouse effect, climate change, global warming*.

Greenhouse gas (GHG).¹ Any gas that absorbs infrared radiation in the atmosphere. Greenhouse gases include, but are not limited to, water vapor, carbon dioxide (CO₂), methane (CH₄), nitrous oxide (N₂O), hydrochlorofluorocarbons (HCFCs), ozone (O₃), hydrofluorocarbons (HFCs), perfluorocarbons (PFCs), and sulfur hexafluoride (SF₆). See *carbon dioxide, methane, nitrous oxide, hydrochlorofluorocarbon, ozone, hydrofluorocarbon, perfluorocarbon, sulfur hexafluoride*.

Halocarbons.¹ Chemicals consisting of carbon, sometimes hydrogen, and either chlorine, fluorine, bromine or iodine.

Halons.¹ Compounds, also known as bromofluorocarbons, that contain bromine, fluorine, and carbon. They are generally used as fire extinguishing agents and cause ozone depletion. Bromine is many times more effective at destroying stratospheric ozone than chlorine. See *ozone depleting substance*.

Heat.⁷ Form of kinetic energy that flows from one body to another when there is a temperature difference between the two bodies. Heat always flows spontaneously from a hot sample of matter to a colder sample of matter. This is one way to state the second law of thermodynamics. See *temperature*.

Heat content.⁵ The amount of heat per unit mass released upon complete combustion.

Heating Degree Days: The number of degrees per day that the average daily temperature is below 65° Fahrenheit. The daily average temperature is the mean of the maximum and minimum temperatures for a 24 hour period. (See Degree Days)

Higher heating value.⁵ Quantity of heat liberated by the complete combustion of a unit volume or weight of a fuel assuming that the produced water vapor is completely condensed and the heat is recovered; also known as gross calorific value. See *lower heating value*.

Histosol.⁹ Wet organic soils, such as peats and mucks.

Hydrocarbons.¹ Substances containing only hydrogen and carbon. Fossil fuels are made up of hydrocarbons. Some hydrocarbon compounds are major air pollutants.

Hydrochlorofluorocarbons (HCFCs).¹ Compounds containing hydrogen, fluorine, chlorine, and carbon atoms. Although ozone depleting substances, they are less potent at destroying stratospheric ozone than chlorofluorocarbons (CFCs). They have been introduced as temporary replacements for CFCs and are also greenhouse gases. See *ozone depleting substance*.

Hydroelectric power plant.⁷ Structure in which the energy of fading or flowing water spins a turbine generator to produce electricity.

Hydrofluorocarbons (HFCs).¹ Compounds containing only hydrogen, fluorine, and carbon atoms. They were introduced as alternatives to ozone depleting substances in serving many industrial, commercial, and personal needs. HFCs are emitted as by-products of industrial processes and are also used in manufacturing. They do not significantly deplete the stratospheric ozone layer, but they are powerful greenhouse gases with global warming potentials ranging from 140 (HFC-152a) to 11,700 (HFC-23).

Hydrologic cycle. The process of evaporation, vertical and horizontal transport of vapor, condensation, precipitation, and the flow of water from continents to oceans. It is a major factor in determining climate through its influence on surface vegetation, the clouds, snow and ice, and soil moisture. The hydrologic cycle is responsible for 25 to 30 percent of the mid-latitudes' heat transport from the equatorial to polar regions.

Hydropower.⁷ Electrical energy produced by falling or flowing water. See *hydroelectric power plant*.

Hydrosphere.⁷ All the earth's liquid water (oceans, smaller bodies of fresh water, and underground aquifers), frozen water (polar ice caps, floating ice, and frozen upper layer of soil known as permafrost), and small amounts of water vapor in the atmosphere.

Industrial End-Use Sector: Comprises manufacturing industries, which make up the largest part of the sector, along with mining, construction, agriculture, fisheries, and forestry. Establishments in this sector range from steel mills to small farms to companies assembling electronic components. Nonutility power producers are also included in the industrial end-use sector.

Infrared radiation.¹ The heat energy that is emitted from all solids, liquids, and gases. In the context of the greenhouse issue, the term refers to the heat energy emitted by the Earth's surface and its atmosphere. Greenhouse gases strongly absorb this radiation in the Earth's atmosphere, and re-radiate some of it back towards the surface, creating the greenhouse effect.

Inorganic compound.⁷ Combination of two or more elements other than those used to form organic compounds. See *organic compound*.

Inorganic fertilizer.⁷ See *synthetic fertilizer*.

Intergovernmental Panel on Climate Change (IPCC).¹ The IPCC was established jointly by the United Nations Environment Programme and the World Meteorological Organization in 1988. The purpose of the IPCC is to assess information in the scientific and technical literature related to all significant components of the issue of climate change. The IPCC draws upon hundreds of the world's expert scientists as authors and thousands as expert reviewers. Leading experts on climate change and environmental, social, and economic sciences from some 60 nations have helped the IPCC to prepare periodic assessments of the scientific underpinnings for understanding global climate change and its consequences. With its capacity for reporting on climate change, its consequences, and the viability of adaptation and mitigation measures, the IPCC is also looked to as the official advisory body to the world's governments on the state of the science of the climate change issue. For example, the IPCC organized the development of internationally accepted methods for conducting national greenhouse gas emission inventories.

Irreversibilities.¹⁰ Changes that, once set in motion, cannot be reversed, at least on human time scales.

Jet fuel⁸ Includes both naphtha-type and kerosene-type fuels meeting standards for use in aircraft turbine engines. Although most jet fuel is used in aircraft, some is used for other purposes such as generating electricity.

Joule.¹ The energy required to push with a force of one Newton for one meter.

Kerogen.⁷ Solid, waxy mixture of hydrocarbons found in oil shale, with a fine grained sedimentary rock. When the rock is heated to high temperatures, the kerogen is vaporized. The vapor is condensed and then sent to a refinery to produce gasoline, heating oil, and other products. See *oil shale, shale oil*.

Kerosene.² A petroleum distillate that has a maximum distillation temperature of 401 degrees Fahrenheit at the 10 percent recovery point, a final boiling point of 572 degrees Fahrenheit, and a minimum flash point of 100 degrees Fahrenheit. Used in space heaters, cookstoves, and water heaters, and suitable for use as an illuminant when burned in wick lamps.

Kyoto Protocol.¹⁰ This is an international agreement struck by 159 nations attending the Third Conference of Parties (COP) to the United Nations Framework Convention on Climate Change (held in December of 1997 in Kyoto Japan) to reduce worldwide emissions of greenhouse gases. If ratified and put into force, individual countries have committed to reduce their greenhouse gas emissions by a specified amount. See *Framework Convention on Climate Change, Conference of Parties*.

Landfill.⁷ Land waste disposal site in which waste is generally spread in thin layers, compacted, and covered with a fresh layer of soil each day.

Lifetime (atmospheric).¹ The lifetime of a greenhouse gas refers to the approximate amount of time it would take for the anthropogenic increment to an atmospheric pollutant concentration to return to its natural level (assuming emissions cease) as a result of either being converted to another chemical compound or being taken out of the atmosphere via a sink. This time depends on the pollutant's sources and sinks as well as its reactivity. The lifetime of a pollutant is often considered in conjunction with the mixing of pollutants in the atmosphere; a long lifetime will allow the pollutant to mix throughout the atmosphere. Average lifetimes can vary from about a week (e.g., sulfate aerosols) to more than a century (e.g., CFCs, carbon dioxide). See *residence time*.

Light-duty vehicles.⁸ Automobiles and light trucks combined.

Lignite.² A brownish-black coal of low rank with high inherent moisture and volatile matter content, used almost exclusively for electric power generation. Also referred to as brown coal.

Liquefied natural gas (LNG).⁷ Natural gas converted to liquid form by cooling to a very low temperature.

Liquefied petroleum gas (LPG).² Ethane, ethylene, propane, propylene, normal butane, butylene, and isobutane produced at refineries or natural gas processing plants, including plants that fractionate new natural gas plant liquids.

Litter.⁹ Undecomposed plant residues on the soil surface. See *decomposition*.

Longwave radiation.⁹ The radiation emitted in the spectral wavelength greater than 4 micrometers corresponding to the radiation emitted from the Earth and atmosphere. It is sometimes referred to as terrestrial radiation or infrared radiation, although somewhat imprecisely. See *infrared radiation*.

Low Emission Vehicle (LEV).⁸ A vehicle meeting the low-emission vehicle standards.

Lower heating value.⁵ Quantity of heat liberated by the complete combustion of a unit volume or weight of a fuel assuming that the produced water remains as a vapor and the heat of the vapor is not recovered; also known as net calorific value. See *higher heating value*.

Lubricant.² A substance used to reduce friction between bearing surfaces or as a process material, either incorporated into other materials used as aids in manufacturing processes or as carriers of other materials. Petroleum

lubricants may be produced either from distillates or residues. Other substances may be added to impart or improve useful properties. Does not include by-products of lubricating oil from solvent extraction or tars derived from de-asphalting. Lubricants include all grades of lubricating oils from spindle oil to cylinder oil and those used in greases. Lubricant categories are paraffinic and naphthenic.

Manure.⁷ Dung and urine of animals that can be used as a form of organic fertilizer.

Mass balance.⁹ The application of the principle of the conservation of matter.

Mauna Loa.⁹ An intermittently active volcano 13,680 feet (4,170 meters) high in Hawaii.

Methane (CH₄).¹ A hydrocarbon that is a greenhouse gas with a global warming potential most recently estimated at 21. Methane is produced through anaerobic (without oxygen) decomposition of waste in landfills, animal digestion, decomposition of animal wastes, production and distribution of natural gas and petroleum, coal production, and incomplete fossil fuel combustion. The atmospheric concentration of methane has been shown to be increasing at a rate of about 0.6 percent per year and the concentration of about 1.7 per million by volume (ppmv) is more than twice its pre-industrial value. However, the rate of increase of methane in the atmosphere may be stabilizing.

Methanol (CH₃OH).⁸ A colorless poisonous liquid with essentially no odor and little taste. It is the simplest alcohol with a boiling point of 64.7 degrees Celsius. In transportation, methanol is used as a vehicle fuel by itself (M100), or blended with gasoline (M85).

Methanotrophic.⁷ Having the biological capacity to oxidize methane to CO₂ and water by metabolism under aerobic conditions. See *aerobic*.

Methyl bromide (CH₃Br).¹¹ An effective pesticide; used to fumigate soil and many agricultural products. Because it contains bromine, it depletes stratospheric ozone when released to the atmosphere. See *ozone depleting substance*.

Metric ton.¹ Common international measurement for the quantity of greenhouse gas emissions. A metric ton is equal to 1000 kilograms, 2204.6 pounds, or 1.1023 short tons.

Mineral.⁷ Any naturally occurring inorganic substance found in the earth's crust as a crystalline solid.

Model year.⁸ Refers to the "sales" model year; for example, vehicles sold during the period from October 1 to the next September 31 is considered one model year.

Molecule.⁷ Chemical combination of two or more atoms of the same chemical element (such as O₂) or different chemical elements (such as H₂O).

Montreal Protocol on Substances that Deplete the Ozone Layer.¹¹ The Montreal Protocol and its amendments control the phaseout of ozone depleting substances production and use. Under the Protocol, several international organizations report on the science of ozone depletion, implement projects to help move away from ozone depleting substances, and provide a forum for policy discussions. In the United States, the Protocol is implemented under the rubric of the Clean Air Act Amendments of 1990. See *ozone depleting substance*, *ozone layer*.

Motor gasoline.² A complex mixture of relatively volatile hydrocarbons, with or without small quantities of additives, obtained by blending appropriate refinery streams to form a fuel suitable for use in spark-ignition engines. Motor gasoline includes both leaded and unleaded grades of finished gasoline, blending components, and gasohol.

Municipal solid waste (MSW).² Residential solid waste and some non-hazardous commercial, institutional, and industrial wastes. This material is generally sent to municipal landfills for disposal. See *landfill*.

Naphtha.² A generic term applied to a petroleum fraction with an approximate boiling range between 122 and 400 degrees Fahrenheit.

Natural gas.⁷ Underground deposits of gases consisting of 50 to 90 percent methane (CH₄) and small amounts of heavier gaseous hydrocarbon compounds such as propane (C₃H₈) and butane (C₄H₁₀).

Natural gas liquids (NGLs).² Those hydrocarbons in natural gas that are separated as liquids from the gas. Includes natural gas plant liquids and lease condensate.

Nitrogen cycle.⁷ Cyclic movement of nitrogen in different chemical forms from the environment, to organisms, and then back to the environment.

Nitrogen fixation.⁷ Conversion of atmospheric nitrogen gas into forms useful to plants and other organisms by lightning, bacteria, and blue-green algae; it is part of the nitrogen cycle.

Nitrogen oxides (NO_x).¹ Gases consisting of one molecule of nitrogen and varying numbers of oxygen molecules. Nitrogen oxides are produced, for example, by the combustion of fossil fuels in vehicles and electric power plants. In the atmosphere, nitrogen oxides can contribute to formation of photochemical ozone (smog), impair visibility, and have health consequences; they are considered pollutants.

Nitrous oxide (N₂O).¹ A powerful greenhouse gas with a global warming potential most recently evaluated at 310. Major sources of nitrous oxide include soil cultivation practices, especially the use of commercial and organic fertilizers, fossil fuel combustion, nitric acid production, and biomass burning.

Nonbiodegradable.⁷ Substance that cannot be broken down in the environment by natural processes. See *biodegradable*.

Nonlinearities.¹⁰ Occur when changes in one variable cause a more than proportionate impact on another variable.

Non-methane volatile organic compounds (NMVOCs).² Organic compounds, other than methane, that participate in atmospheric photochemical reactions.

Non-point source.⁷ Large land area such as crop fields and urban areas that discharge pollutant into surface and underground water over a large area. See *point source*.

Nonutility Power Producer: A corporation, person, agency, authority, or other legal entity of instrumentality that owns electric generating capacity and is not an electric utility. Nonutility producers include qualifying cogenerators, qualifying small power producers, and other nonutility generators (including independent power producers) without a designated, franchised, service area that do not file forms listed in the Code of Federal Regulations, Title 18, Part 141.

Nuclear electric power.³ Electricity generated by an electric power plant whose turbines are driven by steam generated in a reactor by heat from the fissioning of nuclear fuel.

Nuclear energy.⁷ Energy released when atomic nuclei undergo a nuclear reaction such as the spontaneous emission of radioactivity, nuclear fission, or nuclear fusion.

Oil shale.⁷ Underground formation of a fine-grained sedimentary rock containing varying amounts of kerogen, a solid, waxy mixture of hydrocarbon compounds. Heating the rock to high temperatures converts the kerogen to a vapor, which can be condensed to form a slow flowing heavy oil called shale oil. See *kerogen, shale oil*.

Oil. See crude oil, petroleum.

Ore.⁷ Mineral deposit containing a high enough concentration of at least one metallic element to permit the metal to be extracted and sold at a profit.

Organic compound.⁷ Molecule that contains atoms of the element carbon, usually combined with itself and with atoms of one or more other element such as hydrogen, oxygen, nitrogen, sulfur, phosphorus, chlorine, or fluorine. See *inorganic compound*.

Organic fertilizer.⁷ Organic material such as manure or compost, applied to cropland as a source of plant nutrients.

Oxidize.² To chemically transform a substance by combining it with oxygen.

Oxygen cycle.⁷ Cyclic movement of oxygen in different chemical forms from the environment, to organisms, and then back to the environment.

Ozone.⁶ A colorless gas with a pungent odor, having the molecular form of O₃, found in two layers of the atmosphere, the stratosphere and the troposphere. Ozone is a form of oxygen found naturally in the stratosphere that provides a protective layer shielding the Earth from ultraviolet radiation's harmful health effects on humans and the environment. In the troposphere, ozone is a chemical oxidant and major component of photochemical smog. Ozone can seriously affect the human respiratory system.

Ozone Depleting Substance (ODS).¹¹ A family of man-made compounds that includes, but are not limited to, chlorofluorocarbons (CFCs), bromofluorocarbons (halons), methyl chloroform, carbon tetrachloride, methyl bromide, and hydrochlorofluorocarbons (HCFCs). These compounds have been shown to deplete stratospheric ozone, and therefore are typically referred to as ODSs.

Ozone layer.⁷ Layer of gaseous ozone (O₃) in the stratosphere that protects life on earth by filtering out harmful ultraviolet radiation from the sun. See *stratosphere, ultraviolet radiation*.

Ozone precursors.² Chemical compounds, such as carbon monoxide, methane, non-methane hydrocarbons, and nitrogen oxides, which in the presence of solar radiation react with other chemical compounds to form ozone, mainly in the troposphere. See *troposphere*

Particulate matter (PM).⁷ Solid particles or liquid droplets suspended or carried in the air.

Particulates. See *particulate matter*.

Parts per billion (ppb).⁷ Number of parts of a chemical found in one billion parts of a particular gas, liquid, or solid mixture. See *concentration*.

Parts per million (ppm).⁷ Number of parts of a chemical found in one million parts of a particular gas, liquid, or solid. See *concentration*.

Pentanes plus.² A mixture of hydrocarbons, mostly pentanes and heavier fractions, extracted from natural gas.

Perfluorocarbons (PFCs).¹ A group of human-made chemicals composed of carbon and fluorine only. These chemicals (predominantly CF₄ and C₂F₆) were introduced as alternatives, along with hydrofluorocarbons, to the ozone depleting substances. In addition, PFCs are emitted as by-products of industrial processes and are also used in manufacturing. PFCs do not harm the stratospheric ozone layer, but they are powerful greenhouse gases: CF₄ has a global warming potential (GWP) of 6,500 and C₂F₆ has a GWP of 9,200.

Petrochemical feedstock.² Feedstock derived from petroleum, used principally for the manufacture of chemicals, synthetic rubber, and a variety of plastics. The categories reported are naphtha (endpoint less than 401 degrees Fahrenheit) and other oils (endpoint equal to or greater than 401 degrees Fahrenheit).

Petrochemicals.⁷ Chemicals obtained by refining (i.e., distilling) crude oil. They are used as raw materials in the manufacture of most industrial chemicals, fertilizers, pesticides, plastics, synthetic fibers, paints, medicines, and many other products. See *crude oil*.

Petroleum coke.² A residue that is the final product of the condensation process in cracking.

Petroleum.² A generic term applied to oil and oil products in all forms, such as crude oil, lease condensate, unfinished oils, petroleum products, natural gas plant liquids, and non-hydrocarbon compounds blended into finished petroleum products. See *crude oil*.

Photosynthesis.⁷ Complex process that takes place in living green plant cells. Radiant energy from the sun is used to combine carbon dioxide (CO₂) and water (H₂O) to produce oxygen (O₂) and simple nutrient molecules, such as glucose (C₆H₁₂O₆).

Photovoltaic and solar thermal energy.² Energy radiated by the sun as electromagnetic waves (electromagnetic radiation) that is converted into electricity by means of solar (i.e., photovoltaic) cells or useable heat by concentrating (i.e., focusing) collectors.

Point source.⁷ A single identifiable source that discharges pollutants into the environment. Examples are smokestack, sewer, ditch, or pipe. See *non-point source*.

Pollution.⁷ A change in the physical, chemical, or biological characteristics of the air, water, or soil that can affect the health, survival, or activities of humans in an unwanted way. Some expand the term to include harmful effects on all forms of life.

Polyvinyl chloride (PVC).² A polymer of vinyl chloride. It is tasteless, odorless and insoluble in most organic solvents. A member of the family vinyl resin, used in soft flexible films for food packaging and in molded rigid products, such as pipes, fibers, upholstery, and bristles.

Population.⁷ Group of individual organisms of the same species living within a particular area.

Prescribed burning.⁷ Deliberate setting and careful control of surface fires in forests to help prevent more destructive fires and to kill off unwanted plants that compete with commercial species for plant nutrients; may also be used on grasslands.

Primary oil recovery.⁷ Pumping out the crude oil that flows by gravity into the bottom of an oil well. See *enhanced oil recovery*, *secondary oil recovery*.

Quad.⁸ Quad stands for quadrillion, which is, 10¹⁵.

Radiation.¹ Energy emitted in the form of electromagnetic waves. Radiation has differing characteristics depending upon the wavelength. Because the radiation from the Sun is relatively energetic, it has a short wavelength (e.g., ultraviolet, visible, and near infrared) while energy re-radiated from the Earth's surface and the atmosphere has a longer wavelength (e.g., infrared radiation) because the Earth is cooler than the Sun. See *ultraviolet radiation*, *infrared radiation*, *solar radiation*, *longwave radiation*, *terrestrial radiation*.

Radiative forcing.¹ A change in the balance between incoming solar radiation and outgoing infrared (i.e., thermal) radiation. Without any radiative forcing, solar radiation coming to the Earth would continue to be approximately equal to the infrared radiation emitted from the Earth. The addition of greenhouse gases to the atmosphere traps an increased fraction of the infrared radiation, reradiating it back toward the surface of the Earth and thereby creates a warming influence.

Rail.⁸ Includes "heavy" and "light" transit rail. Heavy transit rail is characterized by exclusive rights-of-way, multi-car trains, high speed rapid acceleration, sophisticated signaling, and high platform loading. Also known as subway, elevated railway, or metropolitan railway (metro). Light transit rail may be on exclusive or shared rights of way, high or low platform, multi-car trains or single cars, automated or manually operated. In generic usage, light rail includes streetcars, trolley cars, and tramways.

Rangeland.⁷ Land, mostly grasslands, whose plants can provide food (i.e., forage) for grazing or browsing animals. See *feedlot*.

Recycling.⁷ Collecting and reprocessing a resource so it can be used again. An example is collecting aluminum cans, melting them down, and using the aluminum to make new cans or other aluminum products.

Reforestation.² Replanting of forests on lands that have recently been harvested.

Renewable energy.² Energy obtained from sources that are essentially inexhaustible, unlike, for example, the fossil fuels, of which there is a finite supply. Renewable sources of energy include wood, waste, geothermal, wind, photovoltaic, and solar thermal energy. See *hydropower, photovoltaic*.

Residence time.¹ Average time spent in a reservoir by an individual atom or molecule. Also, this term is used to define the age of a molecule when it leaves the reservoir. With respect to greenhouse gases, residence time usually refers to how long a particular molecule remains in the atmosphere. See *lifetime*.

Residential End-Use Sector: Consists of all private residences, whether occupied or vacant, owned or rented, including single family homes, multifamily housing units, and mobile homes. Secondary home, such as summer homes, are also included. Institutional housing, such as school dormitories, hospitals, and military barracks, generally are not included in the residential end-use sector, but are instead included in the commercial end-use sector.

Residual fuel oil.² The heavier oils that remain after the distillate fuel oils and lighter hydrocarbons are distilled away in refinery operations and that conform to ASTM Specifications D396 and D975. Included are No. 5, a residual fuel oil of medium viscosity; Navy Special, for use in steam-powered vessels in government service and in shore power plants; and No. 6, which includes Bunker C fuel oil and is used for commercial and industrial heating, electricity generation, and to power ships. Imports of residual fuel oil include imported crude oil burned as fuel.

Secondary oil recovery.⁷ Injection of water into an oil well after primary oil recovery to force out some of the remaining thicker crude oil. See *enhanced oil recovery, primary oil recovery*.

Sector. Division, most commonly used to denote type of energy consumer (e.g., residential) or according to the Intergovernmental Panel on Climate Change, the type of greenhouse gas emitter (e.g. industrial process). See *Intergovernmental Panel on Climate Change*.

Septic tank.⁷ Underground tank for treatment of wastewater from a home in rural and suburban areas. Bacteria in the tank decompose organic wastes and the sludge settles to the bottom of the tank. The effluent flows out of the tank into the ground through a field of drainpipes.

Sewage treatment (primary).⁷ Mechanical treatment of sewage in which large solids are filtered out by screens and suspended solids settle out as sludge in a sedimentation tank.

Shale oil.⁷ Slow-flowing, dark brown, heavy oil obtained when kerogen in oil shale is vaporized at high temperatures and then condensed. Shale oil can be refined to yield gasoline, heating oil, and other petroleum products. See *kerogen, oil shale*.

Short ton.¹ Common measurement for a ton in the United States. A short ton is equal to 2,000 lbs. or 0.907 metric tons.

Sink.¹ A reservoir that uptakes a pollutant from another part of its cycle. Soil and trees tend to act as natural sinks for carbon.

Sludge.⁷ Goopy solid mixture of bacteria and virus laden organic matter, toxic metals, synthetic organic chemicals, and solid chemicals removed from wastewater at a sewage treatment plant.

Soil.⁷ Complex mixture of inorganic minerals (i.e., mostly clay, silt, and sand), decaying organic matter, water, air, and living organisms.

Soil carbon.⁹ A major component of the terrestrial biosphere pool in the carbon cycle. The amount of carbon in the soil is a function of the historical vegetative cover and productivity, which in turn is dependent in part upon climatic variables.

Solar energy.⁷ Direct radiant energy from the sun. It also includes indirect forms of energy such as wind, falling or flowing water (hydropower), ocean thermal gradients, and biomass, which are produced when direct solar energy interact with the earth. See *solar radiation*.

Solar radiation.¹ Energy from the Sun. Also referred to as short-wave radiation. Of importance to the climate system, solar radiation includes ultra-violet radiation, visible radiation, and infrared radiation.

Source.⁴ Any process or activity that releases a greenhouse gas, an aerosol, or a precursor of a greenhouse gas into the atmosphere.

Special naphtha.² All finished products within the naphtha boiling range that are used as paint thinners, cleaners, or solvents. Those products are refined to a specified flash point.

Still gas.² Any form or mixture of gases produced in refineries by distillation, cracking, reforming, and other processes. Principal constituents are methane, ethane, ethylene, normal butane, butylene, propane, propylene, etc. Used as a refinery fuel and as a petrochemical feedstock.

Stratosphere.⁷ Second layer of the atmosphere, extending from about 19 to 48 kilometers (12 to 30 miles) above the earth's surface. It contains small amounts of gaseous ozone (O₃), which filters out about 99 percent of the incoming harmful ultraviolet (UV) radiation. Most commercial airline flights operate at a cruising altitude in the lower stratosphere. See *ozone layer, ultraviolet radiation*.

Stratospheric ozone. See *ozone layer*.

Strip mining.⁷ Cutting deep trenches to remove minerals such as coal and phosphate found near the earth's surface in flat or rolling terrain. See *surface mining*.

Subbituminous coal.² A dull, black coal of rank intermediate between lignite and bituminous coal.

Sulfur cycle.⁷ Cyclic movement of sulfur in different chemical forms from the environment, to organisms, and then back to the environment.

Sulfur dioxide (SO₂).¹ A compound composed of one sulfur and two oxygen molecules. Sulfur dioxide emitted into the atmosphere through natural and anthropogenic processes is changed in a complex series of chemical reactions in the atmosphere to sulfate aerosols. These aerosols are believed to result in negative radiative forcing (i.e., tending to cool the Earth's surface) and do result in acid deposition (e.g., acid rain). See *aerosols, radiative forcing, acid deposition, acid rain*.

Sulfur hexafluoride (SF₆).¹ A colorless gas soluble in alcohol and ether, slightly soluble in water. A very powerful greenhouse gas used primarily in electrical transmission and distribution systems and as a dielectric in electronics. The global warming potential of SF₆ is 23,900. See *Global Warming Potential*.

Surface mining.⁷ Removal of soil, sub-soil, and other strata and then extracting a mineral deposit found fairly close to the earth's surface. See *strip mining*.

Synthetic fertilizer.⁷ Commercially prepared mixtures of plant nutrients such as nitrates, phosphates, and potassium applied to the soil to restore fertility and increase crop yields. See *organic fertilizer*.

Synthetic natural gas (SNG).³ A manufactured product chemically similar in most respects to natural gas, resulting from the conversion or reforming of petroleum hydrocarbons. It may easily be substituted for, or interchanged with, pipeline quality natural gas.

Tailings.⁷ Rock and other waste materials removed as impurities when minerals are mined and mineral deposits are processed. These materials are usually dumped on the ground or into ponds.

Tar sand.⁷ Swamp-like deposit of a mixture of fine clay, sand, water, and variable amounts of tar-like heavy oil known as bitumen. Bitumen can be extracted from tar sand by heating. It can then be purified and upgraded to synthetic crude oil. See *bitumen*.

Temperature.⁷ Measure of the average speed of motion of the atoms or molecules in a substance or combination of substances at a given moment. See *heat*.

Terrestrial.⁷ Pertaining to land.

Terrestrial radiation.⁹ The total infrared radiation emitted by the Earth and its atmosphere in the temperature range of approximately 200 to 300 Kelvin. Terrestrial radiation provides a major part of the potential energy changes necessary to drive the atmospheric wind system and is responsible for maintaining the surface air temperature within limits of livability.

Trace gas.¹ Any one of the less common gases found in the Earth's atmosphere. Nitrogen, oxygen, and argon make up more than 99 percent of the Earth's atmosphere. Other gases, such as carbon dioxide, water vapor, methane, oxides of nitrogen, ozone, and ammonia, are considered trace gases. Although relatively unimportant in terms of their absolute volume, they have significant effects on the Earth's weather and climate.

Transportation End-Use Sector: Consists of private and public vehicles that move people and commodities. Included are automobiles, trucks, buses, motorcycles, railroads and railways (including streetcars and subways), aircraft, ships, barges, and natural gas pipelines.

Troposphere.^{1&7} The lowest layer of the atmosphere and contains about 95 percent of the mass of air in the Earth's atmosphere. The troposphere extends from the Earth's surface up to about 10 to 15 kilometers. All weather processes take place in the troposphere. Ozone that is formed in the troposphere plays a significant role in both the greenhouse gas effect and urban smog. See *ozone precursor, stratosphere, atmosphere*.

Tropospheric ozone precursor. See *ozone precursor*.

Tropospheric ozone.¹ See *ozone*.

Ultraviolet radiation (UV).¹¹ A portion of the electromagnetic spectrum with wavelengths shorter than visible light. The sun produces UV, which is commonly split into three bands of decreasing wavelength. Shorter wavelength radiation has a greater potential to cause biological damage on living organisms. The longer wavelength ultraviolet band, UVA, is not absorbed by ozone in the atmosphere. UVB is mostly absorbed by ozone, although some reaches the Earth. The shortest wavelength band, UVC, is completely absorbed by ozone and normal oxygen in the atmosphere.

Unfinished oils.³ All oils requiring further refinery processing, except those requiring only mechanical blending. Includes naphtha and lighter oils, kerosene and light gas oils, heavy gas oils, and residuum.

United Nations Framework Convention on Climate Change (UNFCCC).¹ The international treaty unveiled at the United Nations Conference on Environment and Development (UNCED) in June 1992. The UNFCCC commits signatory countries to stabilize anthropogenic (i.e. human-induced) greenhouse gas emissions to “levels that would prevent dangerous anthropogenic interference with the climate system”. The UNFCCC also requires that all signatory parties develop and update national inventories of anthropogenic emissions of all greenhouse gases not otherwise controlled by the Montreal Protocol. Out of 155 countries that have ratified this accord, the United States was the first industrialized nation to do so.

Vehicle miles traveled (VMT).⁸ One vehicle traveling the distance of one mile. Thus, total vehicle miles is the total mileage traveled by all vehicles.

Volatile organic compounds (VOCs).⁶ Organic compounds that evaporate readily into the atmosphere at normal temperatures. VOCs contribute significantly to photochemical smog production and certain health problems. See *non-methane volatile organic compounds*.

Wastewater.² Water that has been used and contains dissolved or suspended waste materials. See *sewage treatment*.

Water vapor.¹ The most abundant greenhouse gas; it is the water present in the atmosphere in gaseous form. Water vapor is an important part of the natural greenhouse effect. While humans are not significantly increasing its concentration, it contributes to the enhanced greenhouse effect because the warming influence of greenhouse gases leads to a positive water vapor feedback. In addition to its role as a natural greenhouse gas, water vapor plays an important role in regulating the temperature of the planet because clouds form when excess water vapor in the atmosphere condenses to form ice and water droplets and precipitation.

Waxes.² Solid or semisolid materials derived from petroleum distillates or residues. Light-colored, more or less translucent crystalline masses, slightly greasy to the touch, consisting of a mixture of solid hydrocarbons in which the paraffin series predominates. Included are all marketable waxes, whether crude scale or fully refined. Used primarily as industrial coating for surface protection.

Weather.¹ Weather is the specific condition of the atmosphere at a particular place and time. It is measured in terms of such things as wind, temperature, humidity, atmospheric pressure, cloudiness, and precipitation. In most places, weather can change from hour-to-hour, day-to-day, and season-to-season. Climate is the average of weather over time and space. A simple way of remembering the difference is that climate is what you expect (e.g. cold winters) and ‘weather’ is what you get (e.g. a blizzard). See *climate*.

Wetland.⁷ Land that stays flooded all or part of the year with fresh or salt water.

Wetlands.² Areas regularly saturated by surface or groundwater and subsequently characterized by a prevalence of vegetation adapted for life in saturated-soil conditions.

Wood energy.² Wood and wood products used as fuel, including roundwood (i.e., cordwood), limbwood, wood chips, bark, sawdust, forest residues, and charcoal.

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